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Volume III

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J6.4-2. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_E(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 1,000,000$ yr that results when only early WP and early DS failures are considered: (a) expected dose $\bar{D}_E(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_E(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_E(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_E(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 500,000$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_E(\tau)$ and quantiles $Q_q[\bar{D}_E(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-52

J6.4-3. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{NE}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 1,000,000$ yr that results when only nominal process failures and early WP and early DS failures are considered: (a) expected dose $\bar{D}_{NE}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{NE}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_{NE}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{NE}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 500,000$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{NE}(\tau)$ and quantiles $Q_q[\bar{D}_{NE}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-53

J6.5-1. Box plots (see Figure J6.2-3 for description) summarizing probabilities $p_A(\mathcal{A}_{ED}|\mathbf{e}_{Ai})$, $p_A(\mathcal{A}_{EW}|\mathbf{e}_{Ai})$ and $p_A(\mathcal{A}_E|\mathbf{e}_{Ai})$ for scenario classes \mathcal{A}_{ED} , \mathcal{A}_{EW} and \mathcal{A}_E obtained with LHS of size $nLHS = 300$ FJ-54

J7.2-1. Dose to RMEI $D_{II}(\tau|[1, t_k, nWP], \mathbf{e}_{M1})$ from igneous intrusive events at times $t_k = 10, 600, 2000, 4000, 6000$ and $10,000$ yr for $i = 1, 2, \dots, nLHS = 300$: (a) 10 yr, (b) 600 yr, (c) 2000 yr, (d) 4000 yr, (e) 6000 yr, and (f) 10,000 yr. FJ-55

J7.2-2. Illustration of interpolation procedure used to obtain estimated doses $\hat{D}_{II}(\tau|[1, \hat{t}_k, nWP], \mathbf{e}_{M1})$ from calculated doses $D_{II}(\tau|[1, t_k, nWP], \mathbf{e}_{M1})$ indicated in Equation J7.2-18 for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 2.0 \times 10^4$ yr]: (a) $D_{II}(\tau|[1, t_k, nWP], \mathbf{e}_{M1})$, $k = 1, 2, \dots, 10$, (b) interpolated values $\hat{D}_{II}(\tau|[1, \hat{t}_k, nWP], \mathbf{e}_{M1})$ for \hat{t}_k between $t_7 = 6000$ and $t_{8+1} = 10,000$,

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and (c) interpolated values $\hat{D}_{II}(\tau|[1, \hat{t}_k, nWP], \mathbf{e}_{M1})$ for \hat{t}_k between 10 yr and 2.0×10^4 yr. FJ-56

J7.2-3. Determination of $\bar{D}_{II}(\tau|\mathbf{e}_1)$ as indicated in conjunction with Equations J7.2-7, J7.2-9 and J7.2-19 from calculated doses $D_{II}(\tau|[1, t_k, nWP], \mathbf{e}_{M1})$ shown in Equation J7.2-18 for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 2.0 \times 10^4$ yr]..... FJ-57

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CCDFs for $D_{II}(10^4 \text{ yr} | \mathbf{a}_{II}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{II}(10^4 \text{ yr} | \mathbf{a}, \mathbf{e}_{Mi}) | \mathbf{a}_{Ai}]$ defined in Equation J7.2-23 for $i = 1, 2, \dots, 50$, and (c) expected (mean) CCDF and quantile curves, $q = 0.05, 0.5, 0.95$, for CCDFs in (a)..... FJ-63

J7.2-10. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{II}(\tau | \mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only igneous intrusive events are considered: (a) Replicated estimates of expected (mean) dose $\bar{D}_{II}(\tau)$ and quantiles $Q_q[\bar{D}_{II}(\tau | \mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{D}_{II}(\tau)$ FJ-64

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J7.2-14. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{II}(\tau | \mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only igneous intrusion is considered: (a) expected dose $\bar{D}_{II}(\tau | \mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{II}(\tau | \mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_{II}(\tau | \mathbf{e})]$ and quantiles $Q_q[\bar{D}_{II}(\tau | \mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 5 \times 10^5$ yr, and (d) expected (mean) dose $\bar{D}_{II}(\tau)$ and quantiles $Q_q[\bar{D}_{II}(\tau | \mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-68

J7.2-15. Summary presentations of epistemic uncertainty in expected dose $\bar{D}_{II}(\tau | \mathbf{e})$ to RMEI that results when only igneous intrusion is considered for $0 \leq \tau \leq 10^6$ yr. FJ-69

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J7.2-16. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{II,r}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr with r corresponding to ^{226}Ra that results when only igneous intrusion is considered: (a) expected dose $\bar{D}_{II,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{II,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $pE[D < \bar{D}_{II,r}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{II,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 5 \times 10^5$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{II,r}(\tau)$ and quantiles $Q_q[\bar{\bar{D}}_{II,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-70

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J7.2-20. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{II}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only igneous intrusion is considered: (a) Replicated estimates of expected (mean) dose $\bar{\bar{D}}_{II}(\tau)$ and quantiles $Q_q[\bar{\bar{D}}_{II}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{\bar{D}}_{II}(\tau)$ FJ-74

J7.3-1. Dose results $D_{IE}(\tau|[1, t_k, 1, \mathbf{u}_l], \mathbf{e}_{M1})$ obtained for times $t_k = 100, 1000, 4000, 10,000$ yr, igneous eruptive properties \mathbf{u}_l , $l = 1, 2, \dots, nU = 40$, and LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$: (a) $t_k = 100$ yr, (b) $t_k = 1000$ yr, (c) $t_k = 4000$ yr, and (d) $t_k = 10,000$ yr. FJ-75

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J7.3-2. Illustration of interpolation procedure used to obtain estimates $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{M1})$ of conditional expected dose $S(\tau|t, \mathbf{e}_{M1})$ to RMEI (mrem/yr) for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 20,000 \text{ yr}]$: (a) $S(\tau|t_k, \mathbf{e}_{M1})$, $k = 1, 2, \dots, 10$, (b) interpolated values $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between $t_7 = 6000 \text{ yr}$ and $t_8 = 10,000 \text{ yr}$, and (c) interpolated values $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between 10 yr and $20,000 \text{ yr}$ FJ-76

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J7.3-5. Summary of results obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{IE}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000 \text{ yr}$ that results when only igneous eruptive events are considered..... FJ-79

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J7.3-7. Results associated with $D_{IE}(\tau|\mathbf{a}_{IE}, \mathbf{e}_1)$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ obtained with sampling-based (Monte Carlo) procedures: (a) CCDF for $D_{IE}(10^4 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_1)$ with exceedance probabilities $p_A[D < D_{IE}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{M1})|\mathbf{e}_{A1}]$ defined in Equation J7.3-20, and (b) expected dose $\bar{D}_{IE}(10^4 \text{ yr}|\mathbf{e}_1)$ associated with $D_{IE}(10^4 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_1)$ as defined in Equation J7.3-18..... FJ-81

J7.3-8. Results associated with $D_{IE}(10^4 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_M)$ obtained with sampling-based (Monte Carlo) procedures for an LHS of size $nLHS = 300$: (a) CCDFs for $D_{IE}(10^4 \text{ yr}|\mathbf{e}_{IE}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{IE}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J7.3-20 for $i = 1, 2, \dots, nLHS = 300$, (b) CCDFs for $D_{IE}(10^4 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{IE}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J7.3-20 for $i = 1, 2, \dots, 50$, and (c) expected (mean) CCDF and quantile curves, $q = 0.05, 0.5, 0.95$, for CCDFs in (a)..... FJ-82

J7.3-9. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected

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dose $\bar{D}_{IE}(\tau)$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only igneous eruptive events are considered: (a) Replicated estimates of expected (mean) dose $\bar{D}_{IE}(\tau)$ and quantiles $Q_q[\bar{D}_{IE}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{\bar{D}}_{IE}(\tau)$ FJ-83

J7.3-10. Dose results $D_{IE}(\tau|[1, t_k, 1, \mathbf{u}_l], \mathbf{e}_{M1})$ obtained for times $t_k = 40,000, 100,000, 200,000, 400,000$ yr, igneous eruptive properties $\mathbf{u}_l, l = 1, 2, \dots, nU = 40$, and LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$: (a) $t_k = 40,000$ yr, (b) $t_k = 100,000$ yr, (c) $t_k = 200,000$ yr, and (d) $t_k = 400,000$ yr. FJ-84

J7.3-11. Illustration of interpolation procedure used to obtain estimates $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{Mi})$ of conditional expected dose $S(\tau|t, \mathbf{e}_{M1})$ to RMEI (mrem/yr) for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 10^6]$ yr: (a) $S(\tau|t_k, \mathbf{e}_{M1}), k = 1, 2, \dots, 10$, (b) interpolated values $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between $t_8 = 200,000$ yr and $t_9 = 400,000$ yr, and (c) interpolated values $\hat{S}(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between 250 yr and 10^6 yr. FJ-85

J7.3-12. Estimate of $\bar{D}_{IE}(\tau|\mathbf{e}_1)$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and $0 \leq \tau \leq 10^6$ yr with integration-based procedure indicated in Equations J7.3-9 and J7.3-16. FJ-86

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J7.3-14. Summary of results obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{IE}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only igneous eruptive events are considered. FJ-88

J7.3-15. Estimates obtained with LHS of size $nLHS = 300$ of expected (mean) dose $\bar{D}_{IE,r}(\tau)$ to RMEI for $0 \leq \tau \leq 10^6$ yr for individual radioactive species that result when only igneous eruptive events are considered. FJ-89

J7.3-16. Results associated with $D_{IE}(\tau|\mathbf{a}_{IE}, \mathbf{e}_{M1})$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ obtained with sampling-based (Monte Carlo) procedures: (a) CCDF for $D_{IE}(500,000 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_{M1})$ with exceedance probabilities $p_A[D < D_{IE}(500,000 \text{ yr}|\mathbf{a}, \mathbf{e}_{M1})|\mathbf{e}_{A1}]$ defined in Equation J7.3-20, and (b) expected dose $\bar{D}_{IE}(500,000 \text{ yr}|\mathbf{e}_1)$ associated with $D_{IE}(500,000 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_{M1})$ as defined in Equation J7.3-18. FJ-90

J7.3-17. Results associated with $D_{IE}(500,000 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_M)$ obtained with sampling-based (Monte Carlo) procedures for an LHS of size $nLHS = 300$: (a) CCDFs

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for $D_{IE}(10^4 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{IE}(500,000 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J7.3-20 for $i = 1, 2, \dots, nLHS = 300$, (b) CCDFs for $D_{IE}(500,000 \text{ yr}|\mathbf{a}_{IE}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{IE}(500,000 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J7.3-20 for $i = 1, 2, \dots, 50$, and (c) expected (mean) CCDF and quantile curves, $q = 0.05, 0.5, 0.95$, for CCDFs in (a). FJ-91

J7.3-18. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{IE}(\tau)$ to RMEI for $0 \leq \tau \leq 10^6 \text{ yr}$ that results when only igneous eruptive events are considered: (a) Replicated estimates of expected (mean) dose $\bar{D}_{IE}(\tau)$ and quantiles $Q_q[\bar{D}_{IE}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{D}_{IE}(\tau)$ FJ-92

J7.5-1. Box plots summarizing probabilities $p_A[\mathcal{A}_{II}(0, \tau)|\mathbf{e}_{Ai}]$, $p_A[\tilde{\mathcal{A}}_{II}(0, \tau)|\mathbf{e}_{Ai}]$, $p_A[\mathcal{A}_{IE}(0, \tau)|\mathbf{e}_{Ai}]$ and $p_A[\tilde{\mathcal{A}}_{IE}(0, \tau)|\mathbf{e}_{Ai}]$ for scenario classes $\mathcal{A}_{II}(0, \tau)$, $\tilde{\mathcal{A}}_{II}(0, \tau)$, $\mathcal{A}_{IE}(0, \tau)$ and $\tilde{\mathcal{A}}_{IE}(0, \tau)$ defined for the time intervals $[0, 20,000 \text{ yr}]$ and $[0, 1,000,000 \text{ yr}]$ obtained with LHS of size $nLHS = 300$ FJ-93

J8.3-1. Summary of results for damage to CDSP WPs obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in doses $D_{SG}(\tau|[1, t_r, A_s], \mathbf{e}_{Mi})$ for $0 \leq \tau \leq 20,000 \text{ yr}$ damaged area $A_s = (10^{-6}) (32.6 \text{ m}^2)$ and the following values for event time t_r : (a) 200 yr, (b) 1000 yr, (c) 3000 yr, (d) 6000 yr, (e) 12,000 yr, and (f) 18,000 yr. FJ-95

J8.3-2. Box plots (see Figure J6.2-3 for description) summarizing results for damage to CDSP WPs obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in dose $D_{SG}(\tau|[1, t_r, A_s], \mathbf{e}_{Mi})$ with $t_r = 200, 1000, 3000$, and 6000 yrs, $A_s = 10^{-8+s} (32.6 \text{ m}^2)$ for $s = 1, 2, 3, 4, 5$, and $i = 1, 2, \dots, nLHS = 300$, where 32.6 m^2 is the assumed surface area for a CDSP WP: (a) $t_r = 200$ yrs, (b) $t_r = 1000$ yrs, (c) $t_r = 3000$ yrs, and (d) $t_r = 6000$ yrs. FJ-96

J8.3-3. Illustration of interpolation procedure to obtain estimated integrals $\hat{l}_1(\tau|t, \mathbf{e}_{M1})$ from calculated doses $D_{SG}(\tau|[1, t_r, A_s], \mathbf{e}_{M1})$ indicated in Equation J8.3-9 for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 2.0 \times 10^4 \text{ yr}]$: (a) $D_{SG}(\tau|[1, t_r, A_s], \mathbf{e}_{M1})$, $r = 1, 2, \dots, 6$ and $s = 1, 2, \dots, 5$, (b) $\hat{l}_1(\tau|t_r, \mathbf{e}_{M1})$, $r = 1, 2, \dots, 6$, (c) interpolated values $\hat{l}_1(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between $t_4 = 6000 \text{ yrs}$ and $t_5 = 12,000 \text{ yrs}$, and (d) interpolated values for $\hat{l}_1(\tau|\hat{t}_k, \mathbf{e}_1)$ for \hat{t}_k between $= 100 \text{ yrs}$ and $20,000 \text{ yrs}$ FJ-97

J8.3-4. Estimate of $\bar{D}_{SG}(\tau|\mathbf{e}_1)$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and $0 \leq \tau \leq 20,000 \text{ yr}$ with integration-based procedure indicated in Equation J8.3-13. FJ-98

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- J8.3-5. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only seismic ground motion events are considered: (a) expected dose $\bar{D}_{SG}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SG}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $pE[D < \bar{D}_{SG}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 10^4$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ and quantiles $Q_q[\bar{\bar{D}}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-99
- J8.3-6. Summary of results obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only seismic ground motion events are considered. FJ-100
- J8.3-7. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr with r corresponding to ^{99}Tc that results when only seismic ground motion events are considered: (a) expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $pE[D < \bar{D}_{SG,r}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SG,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 10^4$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SG,r}(\tau)$ and quantiles $Q_q[\bar{\bar{D}}_{SG,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-101
- J8.3-8. Estimates obtained with LHS of size $nLHS = 300$ of expected (mean) dose $\bar{\bar{D}}_{SG,r}(\tau)$ to RMEI for $0 \leq \tau \leq 20,000$ yr for individual radioactive species that result when only seismic ground motion events are considered. FJ-102
- J8.3-9. Results associated with $DSG(\tau|\mathbf{a}_{SG}, \mathbf{e}_{M1})$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ obtained with sampling-based (Monte Carlo) procedures: (a) CCDF for $DSG(10^4 \text{ yr}|\mathbf{a}_{SG}, \mathbf{e}_{M1})$ with exceedance probabilities $p_A[D < DSG(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{M1})|\mathbf{e}_1]$ defined in Equation J8.3-22, and (b) expected dose $\bar{D}_{SG}(10^4 \text{ yr}|\mathbf{e}_1)$ associated with $DSG(10^4 \text{ yr}|\mathbf{a}_{SG}, \mathbf{e}_{M1})$ defined in Equation J8.3-20. FJ-103
- J8.3-10. Results associated with $DSG(10^4 \text{ yr}|\mathbf{a}_{SG}, \mathbf{e}_M)$ obtained with sampling-based (Monte Carlo) procedures for an LHS of size $nLHS = 300$: (a) CCDFs for $DSG(10^4 \text{ yr}|\mathbf{a}_{SG}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < DSG(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J4.5-19 for $i = 1, 2, \dots, nLHS = 300$, (b) CCDFs for $DSG(10^4 \text{ yr}|\mathbf{a}_{SG}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < DSG(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J4.5-19 for $i = 1, 2, \dots, 50$, and (c) expected (mean) CCDF and quantile curves, $q = 0.05, 0.5, 0.95$, for CCDFs in (a). FJ-104
- J8.3-11. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only seismic

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ground motion events are considered: (a) Replicated estimates of expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ and quantiles $Q_q[\bar{D}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ FJ-105

J8.4-1. Doses $D_{SG}(\tau|\mathbf{a}_{SG,ik}, \mathbf{e}_{Mi})$, $k = 1, 2, \dots, nR = 30$, and estimated expected doses $\bar{D}_{SG}(\tau|\mathbf{e}_i)$ to the RMEI for $0 \leq \tau \leq 10^6$ yr for selected LHS elements $\mathbf{e}_i = [\mathbf{e}_{Ai}, \mathbf{e}_{Mi}]$ when only nominal processes and seismic ground motion events are considered: (a) \mathbf{e}_1 , (b) \mathbf{e}_2 , (c) \mathbf{e}_3 , (d) \mathbf{e}_4 , (e) \mathbf{e}_5 , and (f) \mathbf{e}_6 FJ-106

J8.4-2. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only nominal processes and seismic ground events are considered: (a) expected dose $\bar{D}_{SG}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SG}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_{SG}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 500,000$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ and quantiles $Q_q[\bar{D}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-107

J8.4-3. Summary of results obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only nominal processes and seismic ground events are considered. FJ-108

J8.4-4. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr with r corresponding to ^{99}Tc that results when only nominal processes and seismic ground events are considered: (a) expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SG,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_{SG,r}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SG,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 500,000$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SG,r}(\tau)$ and quantiles $Q_q[\bar{D}_{SG,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-109

J8.4-5. Estimates obtained with LHS of size $nLHS = 300$ of expected (mean) dose $\bar{\bar{D}}_{SG,r}(\tau)$ to RMEI for $0 \leq \tau \leq 10^6$ yr for individual radioactive species that result when only nominal processes and seismic ground events are considered. . FJ-110

J8.4-6. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{SG}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 10^6$ yr that results when only nominal processes and seismic ground motion events are considered: (a) Replicated estimates of expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ and quantiles $Q_q[\bar{D}_{SG}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{\bar{D}}_{SG}(\tau)$ FJ-111

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- J8.6-1. Summary of results for fault displacement damage of CSNF and CDSP WPs obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in doses $D_{SF,r}(\tau|[1, t_j, 100, A_r/3, \sim]|\mathbf{e}_{Mi})$ for $0 \leq \tau \leq 20,000$ yr, $r = 1, 2$, with $1 \sim$ CSNF WPs and $r = 2 \sim$ CDSP WPs, $A_1 = 2.78 \text{ m}^2$ and $A_2 = 3.28 \text{ m}^2$, and $i = 1, 2, \dots, nLHS = 300$: (a, c, e) CSNF WPs for $t_j = 200, 2000$ and 8000 yrs and (b, d, f) CDSP WPs for $t_j = 200, 2000$ and 8000 yrs. FJ-112
- J8.6-2. Box plots (see Figure J6.2-3 for description) summarizing results for fault displacement damage to CSNF and CDSP WPs obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in dose $D_{SF,r}(10^4 \text{ yr}[1, t_j, 100, A_{rk}, \sim]|\mathbf{e}_{Mi})$ for $t_j = 200, 2000, 4000$ and 8000 yrs, $r = 1, 2$, with $1 \sim$ CSNF WPs and $r = 2 \sim$ CDSP WPs, $A_{rk} = A_r/3, 2A_r/3, A_r$, with $A_1 = 2.78 \text{ m}^2$ and $A_2 = 3.28 \text{ m}^2$, and $i = 1, 2, \dots, nLHS = 300$: (a) $t_j = 200$ yr, (b) $t_j = 2000$ yr, (c) $t_j = 4000$ yr, and (d) $t_j = 8000$ yr. FJ-113
- J8.6-3. Illustration of interpolation procedure to estimate integrals $\hat{l}_r(\tau|t, \mathbf{e}_{Mi})$ and expected dose $\bar{D}_{SF}(\tau|\mathbf{e}_i)$ from calculated doses $D_{SF,r}(\tau|[1, t_j, 100, A_{rk}, 1, \sim], \mathbf{e}_{Mi})$ in Equation J8.6-18 for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and the time interval $[0, 2.0 \times 10^4 \text{ yr}]$: (a) $D_{SF,1}(\tau|[1, t_j, 100, A_{rk}, 1, \sim], \mathbf{e}_{M1})$ for $j = 1, 2, \dots, 6$ and $k = 1, 2, 3$ (Notes: Results overlay for $A_{1k}, k = 1, 2, 3$, for each t_j , and results for t_1 and t_2 also overlay), (b) Interpolated values $\hat{l}_1(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between $t_4 = 4000$ yr and $t_5 = 8000$ yr, (c) Interpolated values for $\hat{l}_1(\tau|\hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between 200 and $20,000$ yr, and (d) estimates for $\bar{D}_{SF,1}(\tau|\mathbf{e}_i)$ for CSNF WPs, $\bar{D}_{SF,2}(\tau|\mathbf{e}_i)$ for CDSP WPs and $\bar{D}_{SF}(\tau|\mathbf{e}_i)$ for CSNF and CDSP WPs. FJ-114
- J8.6-4. Estimate of $\bar{D}_{SF}(\tau|\mathbf{e}_1)$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ and $0 \leq \tau \leq 20,000$ yr with integration-based procedure indicated in Equations J8.6-17 and J8.6-25. FJ-115
- J8.6-5. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SF}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that at results when only fault displacement events are considered: (a) expected dose $\bar{D}_{SF}(\tau|\mathbf{e}_i), i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SF}(\tau|\mathbf{e}_i), i = 1, 2, \dots, 50$, (c) exceedance probabilities $pE[D < \bar{D}_{SG}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SF}(\tau|\mathbf{e})], q = 0.05, 0.5$ and 0.95 , for $\tau = 10^4$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SF}(\tau)$ and quantiles $Q_q[\bar{\bar{D}}_{SF}(\tau|\mathbf{e})], q = 0.05, 0.5, 0.95$ FJ-116
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- J8.6-7. Estimate obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in expected dose $\bar{D}_{SF,r}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr with r

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corresponding to ^{99}Tc that results when only fault displacement events are considered: (a) expected dose $\bar{D}_{SF,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, nLHS = 300$, (b) expected dose $\bar{D}_{SF,r}(\tau|\mathbf{e}_i)$, $i = 1, 2, \dots, 50$, (c) exceedance probabilities $p_E[D < \bar{D}_{SG,r}(\tau|\mathbf{e})]$ and quantiles $Q_q[\bar{D}_{SF,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5$ and 0.95 , for $\tau = 10^4$ yr, and (d) expected (mean) dose $\bar{\bar{D}}_{SF,r}(\tau)$ and quantiles $Q_q[\bar{D}_{SF,r}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$ FJ-118

J8.6-8. Estimates obtained with LHS of size $nLHS = 300$ of expected (mean) dose $\bar{\bar{D}}_{SF,r}(\tau)$ to RMEI for $0 \leq \tau \leq 20,000$ yr for individual radioactive species that result when only seismic ground motion events are considered..... FJ-119

J8.6-9. Results associated with $D_{SF}(\tau|\mathbf{a}_{SF}, \mathbf{e}_{M1})$ for LHS element $\mathbf{e}_1 = [\mathbf{e}_{A1}, \mathbf{e}_{M1}]$ obtained with sampling-based (Monte Carlo) procedures: (a) CCDF for $D_{SF}(10^4 \text{ yr}|\mathbf{a}_{SF}, \mathbf{e}_{M1})$ with exceedance probabilities $p_A[D < D_{SF}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{M1})|\mathbf{e}_{A1}]$ defined in Equations J4.5-19 and J8.6-36, and (b) expected dose $\bar{D}_{SF}(10^4 \text{ yr}|\mathbf{e}_1)$ associated with $D_{SF}(10^4|\mathbf{a}_{SF}, \mathbf{e}_{M1})$ as defined in Equation J4.5-20. FJ-120

J8.6-10. Results associated with $D_{SF}(10^4 \text{ yr}|\mathbf{a}_{SF}, \mathbf{e}_M)$ obtained with sampling-based (Monte Carlo) procedures for an LHS of size $nLHS = 300$: (a) CCDFs for $D_{SF}(10^4 \text{ yr}|\mathbf{a}_{SF}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{SF}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J4.5-19 for $i = 1, 2, \dots, nLHS = 300$, (b) CCDFs for $D_{SF}(10^4 \text{ yr}|\mathbf{a}_{SF}, \mathbf{e}_{Mi})$ with exceedance probabilities $p_A[D < D_{SF}(10^4 \text{ yr}|\mathbf{a}, \mathbf{e}_{Mi})|\mathbf{e}_{Ai}]$ defined in Equation J4.5-19 for $i = 1, 2, \dots, 50$, and (c) expected (mean) CCDF and quantile curves, $q = 0.05, 0.5, 0.95$, for CCDFs in (a). FJ-121

J8.6-11. Assessment with replicated sampling of numerical error associated with use of an LHS of size $nLHS = 300$ to determine epistemic uncertainty in expected dose $\bar{D}_{SF}(\tau|\mathbf{e})$ to RMEI for $0 \leq \tau \leq 20,000$ yr that results when only seismic ground motion events are considered: (a) Replicated estimates of expected (mean) dose $\bar{\bar{D}}_{SF}(\tau)$ and quantiles $Q_q[\bar{D}_{SF}(\tau|\mathbf{e})]$, $q = 0.05, 0.5, 0.95$, and (b) confidence intervals for estimates of expected (mean) dose $\bar{\bar{D}}_{SF}(\tau)$ FJ-122

J8.6-12. Summary of results for fault displacement damage of CSNF and CDSP WPs obtained with LHS of size $nLHS = 300$ showing epistemic uncertainty in doses $D_{SF,r}(\tau|[1, t_j, 100, A_r/3, \sim], \mathbf{e}_{Mi})$ for $0 \leq \tau \leq 20,000$ yr, $r = 1, 2$, with $r = 1 \sim$ CSNF WPs and $r = 2 \sim$ CDSP WPs, $A_1 = 2.78 \text{ m}^2$ and $A_2 = 3.28 \text{ m}^2$ and $i = 1, 2, \dots, nLHS = 300$: (a, c, e) CSNF WPs for $t_j = 1000, 80,000$ and $200,000$ yrs and (b, d, f) CDSP WPs for $t_j = 1000, 80,000$ and $200,000$ yrs..... FJ-123

J8.6-13. Box plots (see Figure J6.2-3 for description) summarizing results for fault displacement damage to CSNF and CDSP WPs obtained with LHS of size

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$nLHS = 300$ showing epistemic uncertainty in dose $D_{SF,r}(10^4 \text{ yr} | [1, t_j, 100, A_{rk}, \sim], \mathbf{e}_{Mi})$ for $t_j = 1000, 80,000, 200,000$ and $400,000$ yrs, $r = 1, 2$ with $r = 1 \sim$ CSNF WPs and $r = 2 \sim$ CDSP WPs, $A_{rk} = A_r/3, 2A_r/3, A_r$, with $A_1 = 2.78 \text{ m}^2$, $A_2 = 3.28 \text{ m}^2$, and $i = 1, 2, \dots, nLHS = 300$: (a) $t_j = 1000$ yr, (b) $t_j = 80,000$ yr, (c) $t_j = 200,000$ yr, and (d) $t_j = 400,000$ yr..... FJ-124

J8.6-14. Illustration of interpolation procedure to estimate integrals $\hat{l}_r(\tau | t, \mathbf{e}_{M1})$ and expected dose $\bar{D}_{SF}(\tau | \mathbf{e}_i)$ from calculated doses $D_{SF,r}(\tau | [1, t_j, 100, A_{rk}, 1, \sim], \mathbf{e}_{M1})$ in Equation J8.6-18 for LHS element $\mathbf{e}_1 = [e_{A1}, e_{M1}]$ and the time interval $[0, 1.0 \times 10^6 \text{ yr}]$: (a) $D_{SF,1}(\tau | [1, t_j, 100, A_{rk}, 1, \sim], \mathbf{e}_{M1})$ for $j = 1, 2, \dots, 6$ and $k = 1, 2, 3$ (Notes: Results overlay for $A_{1k}, k = 1, 2, 3$, for each t_j , and results for t_1 and t_2 also overlay), (b) Interpolated values $\hat{l}_1(\tau | \hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between $t_4 = 2.0 \times 10^5 \text{ yr}$ and $t_5 = 4.0 \times 10^5 \text{ yr}$, (c) Interpolated values for $\hat{l}_1(\tau | \hat{t}_k, \mathbf{e}_{M1})$ for \hat{t}_k between 1000 and $1.0 \times 10^6 \text{ yr}$, and (d) estimates for $\bar{D}_{SF,1}(\tau | \mathbf{e}_i)$ for CSNF WPs, $\bar{D}_{SF,2}(\tau | \mathbf{e}_i)$ for CDSP WPs and $\bar{D}_{SF}(\tau | \mathbf{e}_i)$ for CSNF and CDSP WPs. FJ-125

J8.6-15. Estimate of $\bar{D}_{SF}(\tau | \mathbf{e}_1)$ for LHS element $\mathbf{e}_1 = [e_{A1}, e_{M1}]$ and $0 \leq \tau \leq 10^6 \text{ yr}$ with integration-based procedure indicated in Equations J8.6-17 and J8.6-25..... FJ-126

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8. POSTCLOSURE PERFORMANCE DEMONSTRATION

As stated in the introduction (Section 1.0), this Total System Performance Assessment for the License Application (TSPA-LA) report provides the technical basis to support the License Application (LA) for a nuclear waste repository at Yucca Mountain. This particular section of the report presents the postclosure performance assessment results and, in doing so, addresses the fundamental risk triplet (Kaplan and Garrick 1982 [DIRS 100557]), namely: (1) what can happen, (2) how likely is it to happen, and (3) what are the consequences if it does happen. The first two elements of the risk triplet are addressed via consideration of scenario classes, which represent a broad spectrum of potential future system states, and by accounting for the major sources of aleatory uncertainty. The third component is addressed by utilizing Total System Performance Assessment (TSPA) submodels of key processes (e.g., water flow, radionuclide mobilization and release, dissolved and colloidal phase transport through the natural barriers, and radiologic dose) and by accounting for the epistemic uncertainty in the current knowledge of those processes. Risk insights are developed based on:

- Projections of annual doses to the reasonably maximally exposed individual (RMEI) for a set of release scenarios, which are grouped into four classes
- Identifying the scenario classes and modeling cases that produce the most significant releases and dominate the projected doses
- Determining the sources of uncertainty that are most influential in producing the spread of the distribution of projected doses
- Identifying the key radionuclides that contribute the most to the projected annual doses.

The four scenario classes include Nominal, Early Failure, Seismic, and Igneous. Detailed descriptions of the technical basis of these scenario classes and their associated modeling cases are presented in Sections 6.3 through 6.6. These scenario classes collectively encompass all the included features, events, and processes (FEPs) that affect postclosure performance. The procedure used to screen the FEPs is discussed in Section 6.1.1; all the identified FEPs and the screening arguments are documented in Appendix I.

The TSPA presented herein focused on quantifying the postclosure performance of the repository system, which is composed of three major barriers, namely: (1) upper natural barrier, (2) engineered barrier system (EBS), and (3) lower natural barrier. The upper natural barrier consists of the topography and surface soils of the mountain, the unsaturated tuff units above the repository, and rock strata in which the repository is constructed. The EBS includes the waste packages (WPs), drip shields (DSs), waste forms, cladding (associated with the commercial spent nuclear fuel [CSNF], DOE spent nuclear fuel [DSNF], and naval spent nuclear fuel [NSNF]), and the drift invert; no credit is taken, however, for cladding performance. The lower natural barrier below the repository includes the unsaturated rock layers beneath the repository as well as the volcanic rock units and the alluvial material in the saturated zone (SZ) that extends from the repository site to the designated accessible environment boundary. A depiction of these three major barriers and the fundamental processes contributing to postclosure performance is presented on Figure 8-1.

The postclosure performance of this system of multiple barriers was analyzed for scenarios representing a broad range of likely and unlikely conditions. At a fundamental level, the postclosure performance of the repository is quantified through a detailed computer analysis of:

- How the system of multiple barriers would function in response to changes in the hydrologic, geologic, and geochemical conditions expected to be induced by the decay heat, as well as by projected future climate changes
- How the multiple barriers could be affected by likely and unlikely disruptive events potentially induced by seismic and igneous activity at or near the site area
- How radionuclides could be released from the EBS, migrate through the multiple barriers, and potentially enter the biosphere pathways at the designated RMEI location.

As required by the disposal standards, the postclosure performance is quantified for time periods of 10,000 years and 1,000,000 years after closure.

Because of the inherent uncertainties in making such long-term projections, this TSPA is largely based on a probabilistic modeling approach that accounts for uncertainties in parameters and in future conditions. Other sources of uncertainty, such as potential early failure of the EBS components arising from undetected defects, are included in the early failure scenario class. In addition, a TSPA projection was developed for a human intrusion scenario. This separate TSPA addresses the regulatory requirement to demonstrate performance for a hypothetical inadvertent drilling event, which is assumed to release radionuclides directly to the groundwater.

It is important to note that this TSPA utilizes the knowledge base, site characterization data, and EBS design compiled over the past two plus decades. Since the inception of the Yucca Mountain Project, scientists and engineers at the U.S. Department of Energy (DOE) national laboratories and other organizations have been refining the technical basis in preparation for this performance demonstration. The technical basis has been critically reviewed by both internal and external peer review groups to ensure that it is scientifically defensible. In addition, it has been documented in a manner that is traceable and auditable. The technical basis that directly supports the TSPA-LA Model is documented in Volumes I and II, as well as in the numerous analysis and modeling reports that are explicitly cited throughout this document.

In this section, the postclosure performance demonstration results for the proposed Yucca Mountain repository are presented and explained. Probabilistic projections of annual doses and activity concentrations (i.e., radionuclide activity per unit volume) in groundwater, at the designated location of the RMEI, were developed for comparison to the applicable radiation protection standards. The demonstration of conformance with the radiation protection standards is presented using the following regulatory and statistical metrics:

1. Projected annual dose and activity concentration histories at the RMEI location (based on the propagation of epistemic and aleatory uncertainties through the TSPA-LA Model)
2. Statistical metrics for the projected distributions, including the mean, median, and 5th and 95th percentiles of the annual dose and activity concentration

3. Largest or peak mean annual doses computed for comparison with the dose limits specified for 10,000 years after closure and the largest median annual doses for post 10,000 years but within the period of geologic stability (i.e., 1,000,000 years)
4. Peak of the mean activity concentrations computed for comparison with the groundwater protection limits specified for 10,000 years after closure.

With regard to the probabilistic projections of the dose metrics, the following mathematical terminology is used in the subsequent sections: (1) annual dose is the projected annual dose to the RMEI conditional on one specific sampled value of aleatory (irreducible or stochastic) and epistemic (reducible or lack of knowledge) uncertainty; (2) expected annual dose is the expected value of annual dose to the RMEI over aleatory uncertainty, conditional on epistemic uncertainty; (3) mean annual dose is the expected value of annual dose computed over both the aleatory and epistemic uncertainty; and (4) median annual dose is the median of the distribution of expected annual dose.

As demonstrated in Section 7.3.1, the probabilistic projections are statistically stable. Stability is determined by comparing TSPA-LA Model results computed using three independent samples of uncertain parameters, termed replicates. Because the comparison showed that the TSPA-LA Model results are statistically stable, any of the three replicates could be chosen for presentation and comparison with the radiation protection standards. Results from the first replicate are presented in this section and are the subject of analyses reported in the appendices.

Comparisons of the TSPA projections with the applicable U.S. Nuclear Regulatory Commission (NRC) radiation protection standards are presented first and in a summary manner (Section 8.1). Explanation of the performance demonstration results is developed based on the probabilistic projections for the individual scenario classes and modeling cases (Section 8.2). Additional insights to the factors affecting the isolation characteristics of the repository system are developed by examining the performance capabilities of the natural and engineered barriers (Section 8.3). The final part of this section highlights the technical basis for confidence, credibility, and defensibility of the performance demonstration (Section 8.4).

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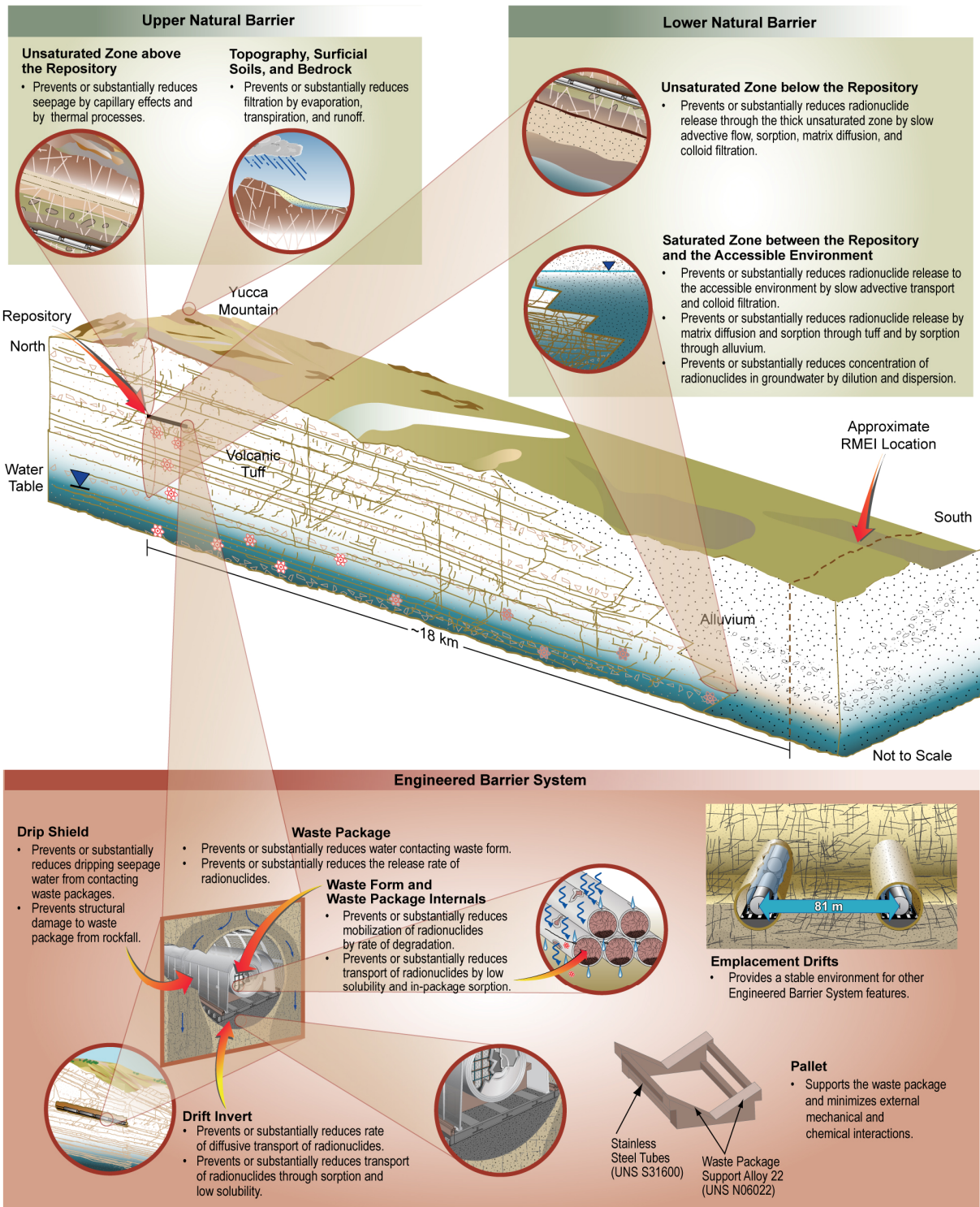


Figure 8-1. Illustration of the Multiple Barriers of Yucca Mountain Repository System

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8.1 CONFORMANCE WITH RADIATION PROTECTION STANDARDS

The U.S. Environmental Protection Agency (EPA) and NRC regulations for a high-level radioactive waste (HLW) repository at Yucca Mountain require that DOE demonstrate a reasonable expectation of compliance with the applicable radiation protection standards. The proposed EPA regulation 40 CFR Part 197 (2005 [DIRS 175755]) establishes three separate and distinct radiation protection standards for the Yucca Mountain Repository. As the licensing agency, the NRC adopted these three radiation protection standards as follows:

- Individual Protection Standard After Permanent Closure (10 CFR 63.311 [DIRS 178394]), which considers the required characteristics of the RMEI as described in 10 CFR 63.312 [DIRS 180319].
- Individual Protection Standard for Human Intrusion (10 CFR 63.321 [DIRS 178394]) according to the Human Intrusion Scenario described in 10 CFR 63.322 [DIRS 180319].
- Separate Standards for Protection of Ground Water (10 CFR 63.331 [DIRS 180319]) using the representative volume specified in 10 CFR 63.332 [DIRS 180319].

The EPA and NRC proposed regulations for Individual Protection and Human Intrusion establish two sets standards for annual doses to the RMEI corresponding to: (1) the time period of 10,000 years after closure, and (2) the time period after 10,000 years but within the period of geologic stability, defined as 1,000,000 years in 10 CFR 63.302 [DIRS 178394]. In contrast, the Separate Standards for Protection of Ground Water set limits for annual dose and activity concentrations (i.e., radionuclide activity per unit volume) for only the 10,000-year period following repository closure. For the purpose of making performance projections for these time periods, the characteristics of the RMEI are defined in 10 CFR 63.312 [DIRS 180319] and the RMEI is taken to reside approximately 18 km (11 mi) downgradient of the repository.

Detailed probabilistic projections developed for this postclosure performance demonstration will be presented and explained in the subsequent sections. At this point, however, it is beneficial to preview the high-level results in a tabulation of peak mean and largest median values for annual doses and activity concentrations for direct comparisons with the numerical limits set in three radiation protection standards. Comparisons of the projected performance metrics and regulatory limits are summarized in Tables 8.1-1 through 8.1-3. To highlight the spread in the computed distribution of expected annual doses, additional statistical parameters are presented in Table 8.1-4 to better clarify the comparison with the limits of the Individual Protection Standard.

As shown in the tabular comparisons, the numerical limits prescribed in all three standards are met, with the peak annual doses and activity concentrations falling well below the limits. These numerical comparisons, however, represent only a part of the postclosure performance demonstration and many detailed calculations and graphical outputs are presented in Sections 8.2 to elucidate the comparisons with the regulatory requirements. For the purposes of explaining the basis for a reasonable expectation of compliance, this section and the subsequent sections address the following fundamental questions regarding the TSPA of the repository and the probabilistic projections of the regulatory metrics:

1. What scenario classes and modeling cases were considered and how were they analyzed to make probabilistic performance projections?
2. What processes or site characteristics modeled in the scenario class were most important to postclosure performance?
3. What multiple-barrier attributes and/or characteristics were most influential in reducing the radionuclide release rates, rate of water or radionuclide movement?
4. What radionuclides in the nuclear waste inventory contribute the most to the annual doses and why?
5. What model parameters are most dominant in determining the uncertainty or spread in the projected performance metrics?

In addition, it is important to identify the plausible and potentially important conservatisms incorporated into the TSPA-LA Model, as well as to explain how the performance projections have been cross-checked to ensure that they are technically sound and defensible.

The next sections describe and explain the projections that constitute the postclosure performance demonstration for the three radiation protection standards. The exposure models and conversion factors used to calculate annual dose and activity concentrations in groundwater are summarized in the Biosphere Component Model section (Section 6.3.11) and described in detail in *Biosphere Model Report* (SNL 2007 [DIRS 177399], Sections 6.11 and 6.12). The annual doses to the RMEI are calculated for two separate exposure scenarios, namely, groundwater and volcanic (SNL 2007 [DIRS 177399], Section 6.1.3).

8.1.1 Individual Protection Standard

To provide a postclosure performance demonstration for the Individual Protection Standard (10 CFR 63.311 [DIRS 178394]), a comprehensive TSPA-LA Model (Sections 1 through 6) of the repository system was developed and applied to:

- Analyze potential releases of radionuclides from the EBS
- Evaluate all potential pathways of radionuclide transport through the multiple barriers and to the accessible environment
- Project (i.e., analyze and calculate as a function of time) the annual doses to the RMEI.

The NRC Proposed Rule 10 CFR 63.311(a) [DIRS 178394] states “(a) DOE must demonstrate, using performance assessment, that there is a reasonable expectation that the reasonably maximally exposed individual receives no more than the following annual dose from releases from the undisturbed Yucca Mountain disposal system: (1) 0.15 mSv (15 mrem) for 10,000 years following disposal; and (2) 3.5 mSv (350 mrem) after 10,000 years, but within the period of geologic stability.”

NRC Proposed Rule 10 CFR 63.311(b) [DIRS 178394] specifies that the “performance assessment must include all potential environmental pathways of radionuclide transport and exposure.” It is important to clarify that, in NRC Proposed Rule 10 CFR 63.311(a) [DIRS 178394], “undisturbed Yucca Mountain disposal system” means that the repository is not affected by human intrusion (10 CFR 63.322 [DIRS 180319]), which is addressed separately in Section 8.1.3. This term is distinct from the regulatory definition of undisturbed performance (10 CFR 63.302 [DIRS 180319]).

To demonstrate compliance with this standard, the peak or largest *mean value* of the projected annual doses to the RMEI is used as the primary performance metric for comparisons with the above numerical limits for 10,000 years and the largest *median value* is used for comparison with the post-10,000-year limit; these metrics are in accordance with 10 CFR 63.303 [DIRS 178394]. The projections of postclosure performance take into account both likely and unlikely FEPs (10 CFR 63.342 [DIRS 178394]) (Appendix I). These FEPs and their affects on waste isolation are quantified in four scenario classes: Nominal, Early Failure, Seismic, and Igneous. Detailed descriptions of the technical basis of these scenario classes and their associated modeling cases are presented in Sections 6.3 through 6.6.

8.1.1.1 Scenario Classes Considered and Calculation of Total Annual Doses

As described in Section 6.1.2, probabilistic projections of the total expected annual dose to the RMEI are computed as a combination of the calculated expected annual doses for each of the four scenario classes; the expectation is taken over all aleatory (i.e., randomly occurring in time) uncertainties. Mathematically, the total expected annual dose to the RMEI, denoted by $\bar{D}_T(\tau|\mathbf{e})$ for time τ in years, is defined as the sum of the expected annual doses for each scenario class (Section 6.1.2.2 and Appendix J4.6):

$$\bar{D}_T(\tau|\mathbf{e}) = \bar{D}_N(\tau|\mathbf{e}) + \bar{D}_E(\tau|\mathbf{e}) + \bar{D}_I(\tau|\mathbf{e}) + \bar{D}_S(\tau|\mathbf{e}) \quad (\text{Eq. 8.1.1-1})$$

In the above equation, the terms $\bar{D}_N(\tau|\mathbf{e})$, $\bar{D}_E(\tau|\mathbf{e})$, $\bar{D}_I(\tau|\mathbf{e})$, and $\bar{D}_S(\tau|\mathbf{e})$ represent the expected annual dose estimates for the four Scenario Classes, namely, Nominal, Early Failure, Igneous, and Seismic, respectively. The term \mathbf{e} represents the set of the epistemically uncertain model parameters, sampled in the probabilistic simulation of postclosure performance.

While the term for the Nominal Scenario Class, $\bar{D}_N(\tau|\mathbf{e})$, consists of only one modeling case, the other expected annual dose terms in Equation 8.1.1-1 are the sum of expected values for the various modeling cases. Specifically, the doses for the Early Failure, Igneous, and Seismic Scenario Classes are computed as following:

$$\bar{D}_E(\tau|\mathbf{e}) = \bar{D}_{WP}(\tau|\mathbf{e}) + \bar{D}_{DS}(\tau|\mathbf{e}) \quad (\text{Eq. 8.1.1-2})$$

$$\bar{D}_I(\tau|\mathbf{e}) = \bar{D}_{II}(\tau|\mathbf{e}) + \bar{D}_{VE}(\tau|\mathbf{e}) \quad (\text{Eq. 8.1.1-3})$$

$$\bar{D}_S(\tau|\mathbf{e}) = \bar{D}_{GM}(\tau|\mathbf{e}) + \bar{D}_{FD}(\tau|\mathbf{e}) \quad (\text{Eq. 8.1.1-4})$$

where the terms $\bar{D}_{WP}(\tau|\mathbf{e})$ and $\bar{D}_{DS}(\tau|\mathbf{e})$ are the expected annual dose estimates for the Waste Package and Drip Shield Early Failure (EF) Modeling Cases, respectively; $\bar{D}_{II}(\tau|\mathbf{e})$ and $\bar{D}_{VE}(\tau|\mathbf{e})$ are the expected annual dose estimates for the Igneous Intrusion and Volcanic Eruption Modeling Cases; and $\bar{D}_{GM}(\tau|\mathbf{e})$ and $\bar{D}_{FD}(\tau|\mathbf{e})$ are the expected annual dose estimates for the Seismic Ground Motion (GM) and Seismic Fault Displacement (FD) Modeling Cases, respectively.

A very important point regarding the TSPA-LA Model implementation of Equation 8.1.1-1 is that, although the expected annual dose for the Nominal Modeling Case, $\bar{D}_N(\tau|\mathbf{e})$, is calculated for the post-10,000-year compliance period, $\bar{D}_N(\tau|\mathbf{e})$ is not added into the total dose calculation for $\bar{D}_T(\tau|\mathbf{e})$. Rather, for the post-10,000-year period, the TSPA-LA Model implicitly produces the sum $\bar{D}_N(\tau|\mathbf{e}) + \bar{D}_S(\tau|\mathbf{e})$ instead of calculating the individual quantities $\bar{D}_N(\tau|\mathbf{e})$ and $\bar{D}_S(\tau|\mathbf{e})$ separately. This combination of modeling cases was performed to ensure proper coupling between the WP and DS damage abstractions in the Seismic GM Modeling Case and the corrosion processes in the Nominal Modeling Case. The damage abstraction takes into account the increased susceptibility to seismic damage as a result of corrosion induced thinning of the WP outer barrier, DS plate, and frame (Section 6.6.1.2.2).

8.1.1.2 Projections of Annual Doses and Major Observations

Projections of postclosure performance for comparison with the Individual Protection Standard were developed using the Monte Carlo simulation methodology, which is described in Appendix J. This methodology incorporates aleatory and epistemic uncertainties into the projections in two separate computational loops: (1) an outer calculational loop that samples probability distributions for model parameters with epistemic uncertainty using the Latin hypercube sampling technique (Helton and Davis 2002 [DIRS 163475]), and (2) an inner loop that evaluates expected annual dose accounting for distributions representing quantities with aleatory uncertainties. The methodology produced an ensemble of expected annual dose outcomes for each scenario class, which were then combined using Equation 8.1.1-1, using the separate terms $\bar{D}_N(\tau|\mathbf{e})$ and $\bar{D}_S(\tau|\mathbf{e})$ for 10,000-year results and including the combined term $\bar{D}_N(\tau|\mathbf{e}) + \bar{D}_S(\tau|\mathbf{e})$ for post-10,000-year results.

The main result of the Monte Carlo simulation process is a set of realizations for the expected annual dose histories for the RMEI, which are generally plotted in the form of a multi-realization plot. The multi-realization plots developed for demonstrating compliance with the Individual Protection Standard are shown on Figure 8.1-1 for 10,000 years after closure and Figure 8.1-2 for the post-10,000-year period (i.e., after 10,000 years but within the period of geologic stability [defined by the proposed NRC rule [DIRS 178394] to end at one million years]). Curves for the mean, median, and 5th and 95th percentiles dose histories are superimposed on each multi-realization plot. The total mean annual dose history, which is plotted as the red curve, was computed by taking the arithmetic average of the 300 expected annual dose values, for individual time planes along the curves. Similarly, the median dose history, plotted as the blue curve, was constructed from points obtained by sorting the 300 expected values from lowest to highest, and

then averaging the two middle values. Curves for the 5th and 95th percentiles are also plotted to illustrate the spread in the expected annual dose histories; 90 percent (or 270 of the 300 epistemic realizations) of the projected dose histories fall between these two percentile curves. For a detailed description of the calculation of total annual dose see Section 6.1.2.2.

For the 10,000-year disposal period, the largest mean and median annual doses to the RMEI are estimated to be about 0.2 mrem and 0.15 mrem, respectively; these projected doses are below the individual protection limit of 15 mrem. Similarly, the largest mean and median annual doses for the post-10,000-year period but within the period of geologic stability are estimated to be about 2 mrem and 1 mrem, respectively; these projected peak values are also below the individual protection limit of 350 mrem. To provide a perspective of the epistemic uncertainty of the performance projections, the annual dose history curves corresponding to the 5th and 95th percentiles are included in the multi-realization plots.

Important risk insights can be gained by disaggregating the total annual doses into the mean annual dose histories for the individual modeling cases; these individual projections are shown on Figure 8.1-3. From those dose curves, the following general observations can be drawn about the projected postclosure performance demonstration:

- Total mean annual doses calculated for both the 10,000-year and post-10,000-year time periods are dominated by releases for the Seismic GM and Igneous Intrusion Modeling Cases.
- Mean annual doses for the Waste Package EF and Seismic FD Modeling Cases are relatively small and are estimated to be on the order of 10^{-2} mrem or less for both the 10,000-year and post-10,000-year time periods.
- Mean annual doses projected for Drip Shield EF and Igneous Volcanic Eruption Modeling Cases are on the order of 10^{-3} mrem or less for both the 10,000-year and post-10,000-year time periods.

With regards to the first observation, the Seismic GM Modeling Case dominates the total mean annual dose for 10,000 years after closure, whereas for the post-10,000-year period, the Igneous Intrusion Modeling Case dominates the total mean annual dose to the RMEI for most of the time period.

An important clarification regarding the projections for the Seismic GM Modeling Case is that the expected annual doses for the 10,000-year time period are only from releases from the damaged co-disposed (CDSP) WPs (Table 6.3.7-1). The contribution to expected annual doses from the commercial spent nuclear fuel (CSNF) WPs were evaluated and found to be insignificant, due to the very low frequency of seismic induced failures of the transportation, aging, and disposal (TAD) packages. The basis for the low frequency is addressed in the following sections and explained in greater detail in Section 7.3.2.6.1.3.7, which examines probability of damage to CSNF WPs by seismic events during the 10,000-year period and the contribution to expected annual dose of these events.

Additional discussion and explanation of the projections for the seven modeling cases is presented in Section 8.2.

8.1.1.3 Disruptive Events Important to Postclosure Performance

From a cause-and-effect view point, clearly the most important disruptive events potentially impacting postclosure performance are the vibratory ground motion and igneous intrusion (i.e., magmatic dike rising through the earth's crust and intersecting the repository). Based on the FEPs screening process, the seismic and igneous disruptions only affect the EBS. In the case of vibratory ground motion, the important aspect is the cumulative damage and failure of the WPs and DSs. It is important to clarify, however, that while WP and DS failures could be caused by a single extreme ground motion event, failure would more typically occur as a result of a sequence of many small to moderate vibratory ground motions over a period of time with stress corrosion cracking (SCC) accumulating from these events. In the case of the igneous intrusion into the repository area, however, the damage and failure of the EBS is a single discrete event that causes complete failure of both WPs and DSs.

For the purposes of risk insights, it is valuable to examine the estimated annual frequencies for these disruptive events and processes. The occurrence of seismic events is described as a Poisson process with the smallest peak ground velocity (PGV) having an annual exceedance frequency of $4.287 \times 10^{-4}/\text{yr}$ and largest PGV with an exceedance frequency of $10^{-8}/\text{yr}$ (Figure 6.6-6; and SNL 2007 [DIRS 176828], Figure 6-7). During the first 10,000 years after closure, only four potentially damaging seismic ground motion events are expected to occur; whereas for the 1,000,000-year period, approximately 430 potentially damaging seismic events are expected to occur. In contrast, the mean annual frequency of a dike intersecting the repository is estimated to be $1.7 \times 10^{-8}/\text{yr}$ (BSC 2004 [DIRS 169989]). This frequency is just slightly greater than the NRC frequency of $10^{-8}/\text{yr}$ for very unlikely events and processes, which are excluded from the performance assessment for the LA by regulation (10 CFR 63.342(b) [DIRS 178394]). While both disruptive events are possible, the potential impacts of seismic ground motion events are much more probable and, therefore, these events are the focus of the subsequent discussion.

Seismically Induced Damage and Failure Mechanisms—Vibratory ground motions can lead to drift degradation and rockfall accumulation on the DS. Subsequent failure of the DS can occur as a result of dynamic loading and deformation and/or by static loading by accumulated rockfall. The ability of the DS to withstand the static and dynamic loading diminishes with time because general corrosion reduces the thickness of the titanium DS plate and framework. Ultimately, the DSs will fail as a result of load-induced buckling or rupture. A histogram of DS failures for the Seismic GM Modeling Case is shown on Figure 8.1-4; the histogram for the Nominal Modeling Case is also shown for the purposes of comparison. As shown, DS failures occur earlier due to the combined effects of nominal and seismic ground motion as compared to nominal process alone. These probabilistic projections of DS and WP failures are described and explained in more detail in Section 8.2.

Seismically induced damage of the WPs is most likely to occur from deformation or denting of the outer wall. These localized areas of deformation or denting develop residual stresses that are susceptible to SCC. Rupture of a WP could potentially occur as a result of kinematic loading caused by package-to-pallet and/or package-to-package impacts. It is important to understand that in the case of the WP, seismic crack damage is more likely to occur than rupture. Projections of CDSP and CSNF WP failures by seismic crack damage as a function of time are

shown on Figure 8.1-5. As can be noted by comparing Figure 8.1-5(a) and Figure 8.1-5(b), the CSNF WP is much less likely to be damaged than is the CDSP WP. A detailed discussion of seismic related damage and failure mechanisms is presented in Section 6.6.1.1.2 and in *Seismic Consequence Abstraction* (SNL 2007 [DIRS 176828]).

All of these potential degradation, damage, and failure mechanisms have been evaluated for this performance demonstration (Section 6.6). The technical basis for the performance analysis of seismic and igneous intrusion consequences are documented in *Seismic Consequence Abstraction* (SNL 2007 [DIRS 176828]) and *Number of Waste Packages Hit by Igneous Intrusion* (SNL 2007 [DIRS 177432]), respectively.

8.1.1.4 Multiple Barrier Processes that Contribute to Postclosure Performance

The multiple barrier processes that are important to postclosure performance have largely been identified in previous TSPAs (Williams 2001[DIRS 157307]) of the Yucca Mountain repository. The modeling studies conducted in support of this TSPA-LA have provided additional insights and understanding, particularly with regard to individual features of natural and engineered barriers. From a high-level perspective, some of the fundamental barrier processes and characteristics that are typically influential in determining how well the system of multiple barriers isolates (i.e., contains and confines) the nuclear waste are:

- Net infiltration into the unsaturated zone (UZ) and seepage into the drifts
- Mechanical strength and corrosion rate properties of the WPs and DSs
- Solubilities for key radioelements such as neptunium, uranium, and plutonium
- Radionuclide sorption onto corrosion products inside the WPs
- Diffusion limited radionuclide releases from failed WPs
- Sorption and matrix diffusion properties of the UZ and SZ underlying the repository.

With regard to infiltration and seepage, these attributes are important because they determine: (1) dripping and non-dripping environments on WP and DS surfaces, and (2) water flow into a failed WP and the ensuing mobilization and release of radionuclides. The strength properties of the WP outer barrier (i.e., Alloy 22 [UNS N06022]) and DS plate (i.e., Titanium Grade 7) and frame (i.e., Titanium Grade 29) determine their capability to withstand dynamic and static loads induced by vibratory ground motion (Section 6.6.1.2). General corrosion and SCC are important processes affecting WP structural integrity because they progressively reduce metal barrier thickness over a period of time. The general corrosion rate of the WP outer barrier is temperature-dependent, which makes the WP surface temperature an important factor. The importance of radioelement solubilities is derived from the fact that they limit the release rates from the waste form. Solubilities of such radioelements as neptunium, uranium, and plutonium are particularly important because they have relatively large initial inventories and produce radionuclide species that have very long half-lives. Corrosion of the WP internals produces metal oxides that provide sorption sites for a variety of fission products and actinide elements (Section 6.3.8). Diffusional release of radionuclides from a failed WP is a function of the breach geometry (e.g., cracks and patches) and effectively limits the rate of release to the natural barrier. The sorption properties of the volcanic tuff and alluvium layers in the UZ and SZ influence the rate of subsurface migration to the RMEI location.

8.1.1.5 Radionuclides Important to Postclosure Performance

In general, the radionuclides in the nuclear waste that dominate the calculation of annual doses typically have a combination of unique characteristics such as: (1) large initial inventories in the nuclear waste, (2) moderately to highly soluble, (3) very long half-lives (e.g., $\geq 10^5$ years), and (4) low to non-sorbing properties. The radionuclides that become important to dose also depend on the time frame considered (i.e., 10,000 years or 1,000,000 years after closure), because of the effect of radionuclide decay on the activity concentrations. In-growth of radionuclides via chain decay, over a very long period of time, can also be an important process that determines the role and importance of actinide elements in the actinium, uranium, neptunium, and thorium series (Figure 6.3.7-4). For the three modeling cases involving groundwater releases, the basic transport processes of advection, dispersion, matrix diffusion, and sorption can play an important role. Moreover, the specific modes (e.g., dissolved and colloidal phases) of transport can also be important; the modes analyzed for each radionuclide are listed in Table 6.3.7-6.

The contributions of the individual radionuclides to the total mean annual dose is shown on Figure 8.1-6 and 8.1-7 for 10,000 years and post-10,000 years, respectively. It is important to acknowledge that some questionable, potentially unrealistic, values of mass dispersivity were found to have been sampled in the probabilistic calculation of the decay chain transport through the SZ; in addition, three radionuclides, namely ^{36}Cl , ^{79}Se , and ^{126}Sn , were inadvertently omitted from the initial TSPA-LA Model runs for the 10,000-year time period. The issues of the dispersivity value and three omitted radionuclides were addressed in a Condition Report (CR 11152). The impact of these issues on the overall analysis is discussed in Appendix P.

Important Radionuclides for the 10,000-Year Performance Projection—From the dose curves shown on Figure 8.1-6, it is clear that the principal contributors to the total mean annual dose, ranked from highest to lowest, are: ^{99}Tc (half-life 2.13×10^5 years), ^{14}C (half-life 5,715 years), ^{239}Pu (half-life 2.41×10^4 years), ^{129}I (half-life 1.57×10^7 years), and ^{240}Pu (half-life 6.56×10^3 years). Collectively, these five radionuclides account for about 98 percent of the peak of the total mean annual dose, which occurs at the end of 10,000 years. The single largest contributor is ^{99}Tc , which accounts for about 56 percent of the peak of the total mean annual dose to the RMEI.

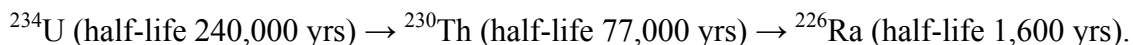
The fission products, ^{99}Tc and ^{129}I , and the activation product, ^{14}C , collectively represent about 83 percent of the total mean annual dose. Their large contribution to total mean annual dose is largely explained by the fact that they are very soluble in water, do not sorb in earth materials, and have long half-lives relative to the 10,000-year time frame. The significance of the technetium, iodine, and carbon radioelements being very soluble is because their release rates are limited only by: (1) the waste form degradation rate, (2) rate and extent of water ingress into the WP, and (3) mass transport mechanism (i.e., diffusion and/or advection) out of the WP. The non-sorbing property is important because these radionuclides are transported from the EBS, through the natural barrier, and to the RMEI at the rate at which the groundwater naturally travels (i.e., no delay by chemical retardation). Their relatively long half-lives, compared to 10,000 years, means that decay would not appreciably reduce their activity level. In contrast to its mobility characteristics, ^{99}Tc has the smallest mean biosphere dose conversion factor (BDCF) (SNL 2007 [DIRS 177399], Figure 6.13-2) of all the major radionuclides; the mean BDCF for ^{14}C is the second lowest. On a mass per package basis, ^{99}Tc has a larger initial inventory than

the ^{129}I and ^{14}C inventory (Table 6.3.7-5), which in part explains the ranking of these three radionuclides.

The two actinides, ^{239}Pu and ^{240}Pu , together contribute about 14 percent to the total mean annual dose with ^{239}Pu by itself contributing about 11 percent. The contribution of these two plutonium species to total mean annual dose is explained by their large initial inventory (Table 6.3.7-5), mode (i.e., dissolved and colloidal) of transport in groundwater, and broad range of transport times in the SZ. With regard to mode of transport, plutonium species are transported through groundwater in dissolved phase and reversible colloids, as well as fast (i.e., not retarded by matrix diffusion or attachment/detachment process) and slow irreversible colloids (Table 6.3.7-6). Detailed probabilistic simulations of plutonium transport through the SZ as reversible colloids, which neglect effects of decay, indicate median transport times ranging from 3,000 years to greater than 1,000,000 years, with a median among all realizations of about 95,000 years (SNL 2007 [DIRS 183750], Table 6-10[a]). In the case of plutonium species in the form of slow irreversible colloids, probabilistic simulations of transport through the SZ that neglect effects of decay indicate transport times ranging from 100 years to about 500,000 years, with a median among all realizations of about 4,500 years (SNL 2007 [DIRS 183750], Table 6-10[a]). In contrast, the simulations for the fast irreversible plutonium colloids range from 10 years to about 1,800 years, with a median among all realizations of about 60 years (SNL 2007 [DIRS 183750], Table 6-10[a]). The TSPA-LA Model projections indicate that the dissolved phase and reversible colloids account for the larger fraction of the contribution to total mean annual dose.

Important Radionuclides for the Post-10,000-Year Performance Projection—Figure 8.1-7 shows the contributors to the total mean annual dose. Ranked from highest to lowest, the four main contributors at 1,000,000 years are: ^{226}Ra (half-life 1,600 years), ^{242}Pu (half-life 3.75×10^5 years), ^{237}Np (half-life 2.14×10^6 years), and ^{129}I (half-life 1.57×10^7 years). These four radionuclides collectively account for nearly 80 percent of the peak mean dose. In marked contrast to ^{242}Pu , ^{237}Np , and ^{129}I , ^{226}Ra has a very small initial inventory in the nuclear waste, relatively short half-life, and sorbs strongly on the geologic media, yet it accounts for about 31 percent of the mean annual dose at 1,000,000 years. This seemingly contrary outcome is explained by: (1) its unique nuclear and chemical properties, and (2) its relatively large BDCF. The nuclear properties of ^{226}Ra explain its persistence in the nuclear waste for potentially millions of years, while its chemical properties explain its rate of migration in the UZ and SZ.

A key nuclear property of ^{226}Ra is that it exhibits in-growth from other radionuclides in the uranium series decay chain (Figure 6.3.7-4). The relevant part of that decay chain consists of the following:



This decay chain is significant because it means that even after ^{226}Ra depletes its initial inventory, it will be continuously replenished so long as there is a source of ^{230}Th and ^{234}U . While both ^{226}Ra and ^{230}Th have relatively small initial inventories in the nuclear waste, the precursor ^{234}U has a significant initial inventory. Also, the large contrast in half-lives between ^{226}Ra and ^{230}Th means that ^{226}Ra will ultimately reach a state of secular equilibrium with ^{230}Th . Similarly, after ^{230}Th depletes its initial inventory, its activity will be in secular equilibrium with

its precursor ^{234}U . The net effect is that ^{226}Ra will persist in the waste form for potentially millions of years.

One of the important chemical properties of ^{226}Ra is that it exhibits high sorption in the unsaturated tuff layers, particularly in the Zeolitic and devitrified tuff (Table 6.3.9-2). This high sorption is also exhibited in the volcanic units and alluvium of the SZ (Table 6.3.10-2). These sorption properties have the effect of greatly slowing the ^{226}Ra rate of migration through the lower natural barrier, to the extent that the activity concentrations of ^{226}Ra would diminish by simple decay before reaching the RMEI location. This transport delay effect is demonstrated in the probabilistic analysis of breakthrough curves (SNL 2007 [DIRS 183750], Figure 6-14[a]) for ^{226}Ra migration through the SZ and to the RMEI location.

The breakthrough curves for ^{226}Ra are reproduced on Figure 8.1-8. It is important to note that the breakthrough curves shown on Figure 8.1-8 do not account for decay during transport; rather, the decay is accounted for when the time-dependent releases of ^{226}Ra from the UZ are computed (SNL 2007 [DIRS 183750], Section 6.5). A statistical analysis of these projections shows that the majority of the ^{226}Ra transport times at the RMEI location are much greater than 10,000 years; more specifically for individual realizations, the median transport times (i.e., the times when relative mass equals 0.5 on Figure 8.1-8a) in the SZ range from 18,000 years to more than 1,000,000 years (SNL 2007 [DIRS 183750], Table 6-10[a]). The 50th percentile of the median ^{226}Ra transport times among all realizations is estimated to be about 731,000 years (SNL 2007 [DIRS 183750], Table 6-10[a]). For these magnitudes of transport times, ^{226}Ra would experience from about 11 to more than 600 half-lives of decay before reaching the RMEI, thus reducing the activity concentration by factors $(1/2)^{11}$ to $(1/2)^{600}$. These decay factors suggest that ^{226}Ra activity concentrations, and dose to the RMEI results from decay during transport of the parents.

To understand how ^{226}Ra actually becomes important to dose, one needs to examine the transport characteristics of its two precursors— ^{230}Th and ^{234}U . The radionuclide ^{230}Th is transported in water as a reversible colloidal phase and a solute (Table 6.3.7-6); whereas ^{234}U is transported only as a solute. Because ^{234}U is weakly sorbed in the geologic media, it is transported through the UZ and SZ at rates many times faster than ^{230}Th or ^{226}Ra . A statistical analysis of the probabilistic breakthrough curves (SNL 2007 [DIRS 183750], Table 6-10[a]) for the SZ that neglect effects of decay shows that the median transport times (through the SZ) for ^{230}Th range from about 1,000 years to over 1,000,000 years, with the 50th percentile of the median transport times among all realizations being greater than 1,000,000 years. In contrast, the median transport times for ^{234}U ranges from 200 to more than 900,000 years; the 50th percentile of the median transport times for all realizations is about 8,900 years. Thus, it follows that the doses from ^{226}Ra are primarily the direct result of the mobilization and groundwater transport of ^{234}U (and to a lesser extent ^{230}Th) with subsequent chain decay to ^{226}Ra . This pattern of radionuclide in-growth occurs along the groundwater flow path.

As noted earlier, ^{226}Ra has a relatively large BDCF value. This is in part due to the fact that its BDCF accounts for several decay products (SNL 2007 [DIRS 177399], Table 6.12-1). In fact, the ^{226}Ra and one of its nuclear progeny, ^{210}Pb , have the highest BDCF values (SNL 2007 [DIRS 177399], Figure 6.13-2) of all the radionuclides considered in the groundwater exposure calculation; moreover, as implemented in the TSPA-LA Model, the ^{226}Ra BDCF includes the

contribution of ^{210}Pb , which on average accounts for about 42 percent of the composite BDCF value.

The results discussed herein contain additional issues identified during analysis. The issues and potential impacts are discussed in Appendix P, in particular in Section P15. Incorporation of any changes related to these issues may modify the interpretation of these results.

8.1.1.6 Model Parameters Influencing Uncertainty in Expected Annual Doses

Uncertainty and sensitivity analyses were conducted to identify the TSPA-LA Model parameters that were most influential in determining the spread in the total expected annual dose projections (Appendix K). Based on those analyses, three model parameters were identified as having the largest influence on the overall uncertainty in the total expected annual doses to the RMEI. These three parameters are (1) occurrence rate of igneous events, IGRATE; (2) residual stress threshold for Alloy 22, SCCTHRP; and (3) temperature dependence parameter for Alloy 22 general corrosion rate, WDGCA22. The importance ranking of these three model parameters varies with time, which is illustrated in Table 8.1-5. Appendix K, Figure K8.1-2, provides information for times prior to 20,000 years, and Figure K8.2-2 provides information for times after 20,000 years. These model parameters are described below; Table K3-4 provides additional details.

IGRATE Parameter—This parameter is the estimated annual frequency of an igneous dike intersecting the repository, which is characterized as an epistemic uncertain quantity. The annual frequency of an igneous event intersecting the repository ranges from approximately $7.4 \times 10^{-10}/\text{yr}$ to $5.5 \times 10^{-8}/\text{yr}$ for the 5th and 95th percentiles, respectively, with a mean annual frequency of $1.7 \times 10^{-8}/\text{yr}$. In a given epistemic realization, the annual frequency of an igneous event is sampled from the cumulative distribution function (CDF) for IGRATE, and is used to determine the probability that an igneous event occurs. As discussed in Section 6.5.1, the Igneous Intrusion Modeling Case assumes that any intrusion penetrating the repository footprint would destroy all DSs and WPs and, therefore, the raw waste form would be exposed to seepage with ensuing radionuclide transport.

SCCTHRP Parameter—This parameter is the residual stress threshold for the Alloy-22 WP outer barrier, which is represented as an epistemically uncertain value. This parameter represents the initiation threshold for the onset of SCC. Thus, when the residual stress in the outer barrier of a WP exceeds this threshold, then SCC is presumed to occur. As explained in Section 6.3.5, the primary causes of residual stresses in the WP outer barrier would be low-frequency, high-peak ground velocity seismic ground motions, which could cause impacts from WP-to-WP, from WP-to-emplacment pallet, and from WP-to-DS; these impacts could potentially cause dynamic loads that dent the outer barrier, which could result in creation of residual stresses. The uncertainty in this model parameter is represented using a uniform distribution.

WDGCA22 Parameter—This parameter relates to the temperature dependence for the general corrosion rate of the Alloy 22 WP outer barrier, which is characterized as an epistemically uncertain quantity. As explained in Section 6.3.5, this parameter determines the magnitude of this temperature dependence and directly influences the short-term and long-term general

corrosion rates of the Alloy 22. Larger values of this parameter correspond to higher general corrosion rates while WP temperatures are above 60 °C, and to lower general corrosion rates when WP temperatures are below 60 °C. This parameter is sampled from a truncated normal distribution

It is important to clarify that the parameter names IGRATE, SCCTHRP, and WDGCA22 are those used in the uncertainty and sensitivity analyses. The corresponding parameter names used in the TSPA-LA Model and the TSPA Input database are Igneous_Event_Prob_a, Stress_Thresh_A22_a, and C1_GenCorr_A22_a, respectively. More detailed discussion of the importance of these and other model parameters is given in Appendix K.

8.1.2 Groundwater Protection

The performance demonstration for the Groundwater Protection Standard was conducted in a manner consistent with the NRC requirements of 10 CFR 63.331 [DIRS 180319]. The radiation protection limits, set in the Groundwater Protection Standard, are restricted to the first 10,000 years after closure; thus, a performance demonstration for this standard beyond this time period was not conducted. The NRC implementation of this standard requires that the DOE TSPA: (1) evaluate potential releases for undisturbed performance, and (2) show a reasonable expectation that the level of radioactivity in a representative volume of groundwater does not exceed the numerical limits in 10 CFR 63.331 ([DIRS 180319], Table 1).

NRC Table 1 establishes limits for selected radionuclides and specific types of radiation emitted. More specifically, Table 1 limits radionuclides in the representative volume in terms of three performance metrics:

1. Combined ^{226}Ra and ^{228}Ra activity concentrations (pCi/L) in groundwater is ≤ 5 pCi/L (natural background included)
2. Gross alpha activity (including ^{226}Ra but excluding radon and uranium isotopes) concentration is ≤ 15 pCi/L (natural background included)
3. Combined beta- and photon-emitting radionuclides doses to the whole body or any organ (based on drinking 2 liters of water per day from the representative volume) is ≤ 4 mrem.

Performance calculations for all three regulatory metrics are for the point of highest concentration in the plume of contamination in the accessible environment. The calculations of whole body and organ doses are based on using the representative volume of groundwater of 3,000 acre-ft/yr, as specified in 10 CFR 63.332 [DIRS 180319].

It is important to clarify that the term undisturbed performance, as defined in the NRC regulations (10 CFR 63.302 [DIRS 180319]), means that “human intrusion or the occurrence of unlikely natural features, events, and processes do not disturb the disposal system.” The NRC defines unlikely as “features, events, and processes, or sequences of events and processes, i.e., those that are estimated to have less than one chance in 10 and at least one chance in 10,000 of occurring within 10,000 years” (10 CFR 63.342(b) [DIRS 178394]). Simply stated, the performance demonstration for the groundwater protection standard requires that the TSPA

assess radionuclide releases caused by “likely” FEPs (i.e., having an annual frequency of 10^{-5} /yr or greater).

Scenario Classes and Modeling Cases Considered—The regulatory definitions for undisturbed performance and unlikely FEPs were used to select specific modeling cases for inclusion in the performance demonstration for this standard. The results of that screening are summarized below:

- Nominal Modeling Case included
- Waste Package EF Modeling Case included
- Drip Shield EF Modeling Case included
- Igneous Intrusion Modeling Cases excluded
- Volcanic Eruption Modeling Cases excluded
- Seismic GM Modeling Case included (for PGVs with annual frequency greater than 10^{-5} /yr only)
- Seismic FD Modeling Case excluded.

The Nominal Modeling Case by definition represents all the expected or likely FEPs. However, because of the inherent characteristics of the natural barrier and performance of the EBS, the effects of corrosion on the WPs (and DSs) are expected to be very small and thus the WPs and DSs are not expected to fail within the 10,000 years for this modeling case.

The Waste Package and Drip Shield EF Modeling Cases were selected for inclusion in the performance demonstration because the probability of an early failure of a WP or DS is greater than 10^{-5} (Section 6.4).

Both the Igneous Intrusion and the Volcanic Eruption Modeling Cases have mean annual frequencies of occurrence that are much less than 10^{-5} per year (Section 6.5). In fact, these two modeling cases are characterized by unlikely FEPs with estimated mean annual frequencies of slightly greater than 10^{-8} per year (Section 6.5). On this basis, the two igneous modeling cases were excluded from consideration in the performance demonstration for groundwater protection.

The Seismic GM Modeling Case represents events and processes with mean exceedance frequencies in the range from 10^{-8} to 4.287×10^{-4} per year (SNL 2007 [DIRS 176828], Section 6.4.3). In contrast, the Seismic FD Modeling Case represents events that do not occur with mean exceedance frequencies greater than 10^{-7} per year (Section 6.6.1.2.3). Therefore, the Seismic GM Modeling Case events with mean exceedance frequencies in the range from 10^{-5} to 4.3×10^{-4} per year are considered to be likely and are included in the TSPA for the groundwater protection standard. The Seismic FD Modeling Case is excluded because of its low annual exceedance frequency near 10^{-7} per year (SNL 2007 [DIRS 176828], Table 6-67).

In summary, the performance demonstration for the Groundwater Protection Standard considered radionuclide releases for FEPs that were represented in the Waste Package and Drip Shield EF

Modeling Cases, and the Seismic GM Modeling Case. The early failure and ground motion modeling cases are implemented in a probabilistic manner to determine the number of WPs failed as a function of time in a given realization (Sections 6.4 and 6.6). As indicated previously, the projections for this standard are based on considering releases only from the damaged CDSP WPs; the potential releases from the CSNF WPs were omitted because the contribution to expected annual dose from seismic damage to CSNF WPs has been shown to be insignificant (Section 7.3.2.6.1.3).

Probabilistic projections for the three groundwater protection metrics were developed for the time period of 10,000 years and compared to the specified numerical limits. Those comparisons are presented and explained below.

8.1.2.1 Projections for Combined ^{226}Ra and ^{228}Ra

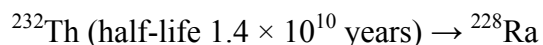
The performance demonstration for this first metric of the Groundwater Protection Standard is based on the combined activity concentration for ^{226}Ra and ^{228}Ra . The probabilistic projections of the time history of the activity concentrations are presented on Figure 8.1-9. The curves shown are background, mean, and 95th percentile activity concentrations of combined radium species (excluding background). As shown on Figure 8.1-9, the 95th percentile value for the combined radium concentration is less than the mean value. This result was a consequence of a few realizations that projected relatively high, but still small, radium concentrations that skewed the distribution of radium concentrations and caused the mean value to be higher than the 95th percentile value.

The curves on Figure 8.1-9 show that the peak mean groundwater concentration of combined ^{226}Ra and ^{228}Ra at the RMEI location is less than 10^{-5} pCi/L, which is well below the 5 pCi/L limit found in 10 CFR 331 ([DIRS 180319], Table 1). The naturally occurring background activity concentration of combined radium at the site is estimated to be ~ 0.5 pCi/L (SNL 2008 [DIRS 183750], Section 6.8.5, Table 6-18). At this background level, the projected peak radium activity concentrations of less than 10^{-5} pCi/L for the postulated repository releases would not make any measurable change to the background level in groundwater at the RMEI location. The fundamental reasons for the very low combined radium concentrations at the RMEI location are as follows:

^{228}Ra Activity Concentration—The chemical and nuclear properties of ^{228}Ra , together with its very small inventory, largely explain its small contribution to the total radium activity concentrations. As previously stated, the radioelement radium sorbs very strongly in the geologic media (Tables 6.3.9-2 and 6.3.10-2). This chemical property translates into ^{228}Ra taking very long transport times to migrate through the lower natural barrier. As previously stated, the projected mass breakthrough curves for radium through the SZ alone are predominantly greater than 10,000 years. However, if one postulates a hypothetical fast pathway with a very unlikely transport time, for example, of 500 years, then ^{228}Ra , with a half-life of 5.8 years, would experience about 86 half-lives of decay before reaching the RMEI location. This would reduce its activity by a factor of $(1/2)^{86} \sim 10^{-26}$. The initial total inventory ^{228}Ra in the CDSP WPs can be estimated by multiplying its specific activity (Table 6.3.7-5) times the number of grams of ^{228}Ra per package (Table 6.3.7-5) times the total number of CDSPs (Table 6.3.7-1) or $(2.72 \times 10^2 \text{ Ci/g} \times \sim 2 \times 10^{-5} \text{ g/pkg} \times 3416 \text{ pkg}) = 18.6 \text{ Ci}$ or less than 20 Ci. This means that,

after 86 half-lives of decay, the original quantity of ^{228}Ra would be reduced to 2×10^{-13} pCi. Mixing this activity in the representative volume of 3,000 acre-ft would suggest that the ^{228}Ra activity concentration in the groundwater would likely be undetectable.

Considering that ^{228}Ra is a member of the thorium series, it could potentially reach the RMEI location via transport of its precursor ^{232}Th and subsequent in-growth. The relevant part of the decay chain is:



Because of its extraordinarily long half-life, the ^{232}Th is for all practical purposes a stable element for time periods of 10,000 years. For this reason, it is conservatively assumed that ^{228}Ra and ^{232}Th are in secular equilibrium and the activity of ^{228}Ra is the same as the ^{232}Th .

In the absence of ^{228}Ra inventory regeneration by in-growth and considering characteristically long radium transport times (through the SZ alone), it is reasonable to expect the activity concentrations for ^{228}Ra to show effectively undetectable levels for 10,000 years after disposal.

^{226}Ra Activity Concentration—The explanation for the projected low ^{226}Ra activity concentrations is similar to that for ^{228}Ra . The primary difference is that ^{226}Ra is a member of the uranium series with its two primary precursors being ^{234}U (half-life 240,000 years) and ^{230}Th (half-life 7.54×10^4 years). Over a time period of 10,000 years, the ^{234}U activity would experience a decay change of about 3 percent, which is a relatively small amount of in-growth to ^{230}Th . Similarly, about 9 percent of ^{230}Th activity would decay to ^{226}Ra in 10,000 years. In contrast, ^{226}Ra activity would diminish by almost 99 percent over the 10,000 years.

As noted previously, the median ^{226}Ra transport times through SZ for individual Monte Carlo realizations range from 18,000 years to more than 1,000,000 years, with the median over all realizations exceeding 700,000 years. The median ^{230}Th transport times through SZ for individual realizations range from about 1,000 years to more than 1,000,000 years, with a median for all realizations of greater than 1,000,000 years. Because the ^{226}Ra will be inventory limited, and its projected long transport times in the SZ, it is reasonable to expect that ^{226}Ra activity concentration would be low. Moreover, ^{226}Ra is limited by the small in-growth from ^{230}Th , occurring along the groundwater flow path.

Thus, the probabilistic projections for the combined radium groundwater standard demonstrate a reasonable expectation that the level of radioactivity in a representative volume of groundwater (i.e., 3,000 acre-ft) would not exceed the numerical limit of 5 pCi/L for groundwater protection.

8.1.2.2 Projections for Gross Alpha Activity

The performance demonstration for this metric of the Groundwater Protection Standard is based on a calculation of the gross alpha activity, including ^{226}Ra but excluding ^{222}Rn and Uranium species (10 CFR 63.331). The TSPA Biosphere Component Model, documented in DTN: MO0702PAGWPROS001_R0 [DIRS 179328], identifies 15 primary radionuclides that have one or more alpha emitters in their decay chain to the next tracked radionuclide. The specific alpha emitting radionuclides to consider in estimating the gross alpha are:

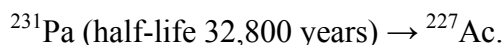
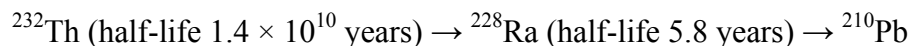
- ^{210}Pb (half-life 22.3 years)
- ^{226}Ra (half-life 1,600 years)
- ^{227}Ac (half-life 21.8 years)
- ^{228}Th (half-life 1.913 years)
- ^{229}Th (half-life 7,300 years)
- ^{230}Th (half-life 75,400 years)
- ^{232}Th (half-life 1.4×10^{10} years)
- ^{231}Pa (half-life 32,800 years)
- ^{237}Np (half-life 2.14×10^6 years)
- ^{238}Pu (half-life 87.7 years)
- ^{239}Pu (half-life 24,100 years)
- ^{240}Pu (half-life 6,560 years)
- ^{242}Pu (half-life 3.75×10^5 years)
- ^{241}Am (half-life 433 years)
- ^{243}Am (half-life 7,370 years).

The probabilistic projections for gross alpha activity concentrations over the 10,000-year time period are shown on Figure 8.1-10. The activity concentration for gross alpha was calculated based on the annual mass flux of the alpha emitting radionuclides across the boundary of the accessible environment and collected in the representative groundwater volume of 3,000 acre-ft/yr. The plot on this figure shows curves for the mean and 95th percentile for gross alpha activity concentration. The peak mean activity concentration is projected to be less than 1×10^{-4} pCi/L (excluding background). This activity concentration is below the ground-water protection standard of 15 pCi/L (found in 10 CFR 331 [DIRS 180319], Table 1) for gross alpha and well below the natural background level of ~ 0.5 pCi/L in the site groundwater. As a result, the postulated releases of the alpha emitting radionuclides are effectively limited by the EBS and sufficiently confined to the lower natural barrier.

The radionuclides making the largest contributions to mean gross alpha activity are ^{237}Np , ^{239}Pu , ^{240}Pu , and ^{242}Pu . These four radionuclides have relatively large initial inventories (in the CSNF waste form) and long to very long half-lives. Moreover, the three plutonium species are transported through the groundwater pathway in both dissolved and colloidal phases. The low contributions of the other alpha emitters can be plausibly explained by examining their initial inventories (Table 6.3.7-5), decay chain properties (Figure 6.3.7-4), half-lives (Figure 6.3.7-4), and sorption properties (Tables 6.3.9-2 and 6.3.10-2) (i.e., distribution coefficient or K_d). A qualitative examination leads to the following observations and inferences explained below.

Low Inventory, Short Half-Lives, and Strong Sorption—The alpha emitting radionuclide ^{228}Th is easily eliminated as a potentially significant contributor to gross alpha because its very short half-life of 1.913 years and low initial inventory (provided by the decay of ^{245}Cm) will most likely deplete before loss of containment. Similarly, the radionuclides ^{238}Pu and ^{241}Am have relatively short half-lives; equally important is the fact that they are strongly sorbed in the UZ rock layers and in the volcanic units and alluvium of the SZ. Thus, it is reasonable to expect that these two alpha emitters are not likely to reach the accessible environment in most TSPA-LA realizations. Thus, based on these qualitative arguments, ^{228}Th , ^{238}Pu , and ^{241}Am will likely not make a significant contribution to the gross alpha activity concentrations.

Secular Equilibrium and Sorption—The alpha emitting radionuclides ^{210}Pb , ^{226}Ra , and ^{227}Ac are not likely to be significant contributors to gross alpha because their activity is controlled via secular equilibrium with their precursor radionuclides, namely:



As was explained earlier for the combined radium groundwater standard, these radionuclide species are unlikely to produce significant activity concentrations at the accessible environment boundary because of the sorption properties of both radium and its precursor thorium. In the case of the radioelement protactinium, its initial inventory in the WPs is low and its sorption properties in the unsaturated rock layers is relatively high (e.g., mean K_{ds} of $\sim 5,000$ mL/g), which suggests that the precursors (^{232}Th , ^{230}Th , and ^{231}Pa) or the daughters (^{210}Pb , ^{226}Ra , and ^{227}Ac) are not likely to contribute appreciably to the gross alpha activity levels.

Very High Sorption—The radioelement americium has the property of being very strongly sorbed in the unsaturated rock layers (e.g., mean K_{ds} of 400 to about 5,000 mL/g), as well as in the volcanic units (e.g., mean K_{ds} of 400 to about 5,000 mL/g) and alluvium (e.g., mean K_{ds} of $\sim 5,000$ mL/g) of the SZ. On the other hand, this radioelement has a significant initial inventory in the CSNF and can also be transported in colloidal phase (i.e., reversible and irreversible).

Thus, the probabilistic projections for this regulatory performance metric show a reasonable expectation that the level of radioactivity in a representative volume of groundwater would not exceed the numerical limit of 15 pCi/L for groundwater protection.

8.1.2.3 Projections for Combined Beta- and Photon-Emitting Radionuclides

The performance demonstration for this metric of the Groundwater Protection Standard is based on combined beta- and photon-emitting radionuclides; both the primary beta emitters and any daughter products that decay by beta-emission are considered. The annual doses from exposure to beta-photon emitters are quantified in terms of both whole body and organ dose. The TSPA Biosphere Component Model documented in the *Biosphere Model Report* (SNL 2007 [DIRS 177399], Table 6.15-4), identifies a total of 19 primary radionuclides that are used to compute this groundwater protection metric.

Some of the more prominent beta emitters are: ^{14}C , ^{36}Cl , ^{79}Se , ^{90}Sr , ^{99}Tc , ^{129}I , ^{135}Cs , and ^{137}Cs . Of this set, only ^{90}Sr and ^{137}Cs have short half-lives (~ 30 years) relative to the 10,000 year time period. Some of the beta-photon emitters are daughter products of alpha and beta emitters (SNL 2007 [DIRS 177399], Table 6.15-4) such as ^{137m}Ba , ^{228}Ac , ^{212}Pb , and ^{208}Tl . Because this group of radioisotopes has half-lives ranging from minutes to several hours, they are not included in radionuclide transport calculations; however, the associated conversion factors are included in calculating the beta-photon dose. The projections of annual doses for these radionuclides are evaluated as a function of the release rates from the repository, in-growth, and groundwater transport to the accessible environment.

Probabilistic projections for the mean whole body and thyroid annual doses for the 10,000-year time period are shown on Figure 8.1-11; a curve for the 95th percentile is also shown. These annual doses were calculated by summing all the annual doses from the beta- and photon-emitting radionuclides included in the TSPA-LA Model. The peak mean annual drinking water dose to the thyroid is estimated to be about 0.2 mrem. The whole body dose shown on the same figure takes into account the effect on all organs and includes the organ-dose weighting factors. The peak mean annual drinking water dose to the whole body in this case is estimated to be about 4×10^{-2} mrem. The radionuclide that dominates the mean thyroid dose is ^{129}I , whereas ^{99}Tc dominates the mean whole body dose.

Thus, the probabilistic projections for combined beta-photon dose demonstrate a reasonable expectation that the annual doses to the whole body or any organ would not exceed the numerical limit of 4 mrem.

8.1.3 Human Intrusion Protection

As required by the proposed NRC regulation, a performance demonstration was developed for the Human Intrusion Standard (10 CFR 63.321 [DIRS 178394]). The probabilistic projections provide an estimate of the annual dose to the RMEI resulting from a stylized human intrusion drilling scenario and compare the result to the dose limits that the geologic repository at the Yucca Mountain site must meet. The EPA regulation 40 CFR Part 197 (2005 [DIRS 175755]) clarifies that the performance assessment for the Human Intrusion Scenario is to be presented separately and not integrated into the TSPA for the Individual Protection Standard.

The Human Intrusion Standard parallels the Individual Protection Standard in that the same numerical limits for annual dose must be met. More specifically, the Human Intrusion Standard (10 CFR 63.321 (a) and (b) in the proposed NRC regulation [DIRS 178394]) specifies that the DOE must:

1. Determine the earliest time after disposal that a WP would degrade sufficiently that a drilling intrusion could occur without recognition by the drillers
2. Demonstrate a reasonable expectation that the RMEI, as a result of the human intrusion, will not receive an annual dose of:
 - a. 0.15 mSv (15 mrem) for 10,000 years following disposal
 - b. 3.5 mSv (350 mrem) after 10,000 years, but within the period of geologic stability
3. Include all potential environmental pathways of radionuclide transport and exposure subject to the requirements of 10 CFR 63.322 [DIRS 180319].

The characteristics and assumptions of the stylized Human Intrusion Scenario are specified in the NRC regulation 10 CFR 63.322 [DIRS 180319]. That section of the regulation prescribes the following scenario characteristics for the drilling event:

- There is a single human intrusion as a result of exploratory drilling for ground water, per 10 CFR 63.322(a).
- The intruders drill a borehole directly through a degraded WP into the uppermost aquifer underlying the Yucca Mountain Repository, per 10 CFR 63.322 (b).
- The drillers use the common techniques and practices that are currently employed in exploratory drilling for ground water in the region surrounding Yucca Mountain, per 10 CFR 63.322 (c).
- Careful sealing of the borehole does not occur, instead natural degradation processes gradually modify the borehole, per 10 CFR 63.322 (d).
- No particulate waste material falls into the borehole, per 10 CFR 63.322 (e).
- The exposure scenario includes only those radionuclides transported to the SZ by water (e.g., water enters the WP, releases radionuclides, and transports radionuclides by way of the borehole to the SZ), per 10 CFR 63.322 (f).
- No releases are included which are caused by unlikely natural processes and events, per 10 CFR 63.322 (g).

With regard to the last specification, the NRC regulation defines unlikely natural processes and events as those having a probability of less than one chance in 10 and at least one chance in 10,000 of occurring within 10,000 years of disposal, as defined in 10 CFR 63.342(b) [DIRS 178394]. This probability statement is equivalent to a cut-off annual frequency criterion of $\leq 10^{-5}/\text{yr}$ for exclusion of unlikely processes and events.

For the purposes of the TSPA, it is assumed that inadvertent drilling into the repository results in a penetration of the DS and WP, as well as creating a direct pathway to the groundwater. The conceptualization of the scenario includes radionuclides transported vertically through the UZ, horizontally along the SZ, and then withdrawn with the groundwater at the location of the RMEI. The exposure characteristics for the RMEI are as defined in 10 CFR 63.312(a) through (e) [DIRS 180319].

8.1.3.1 Determination of Earliest Time for Drilling Intrusion

A detailed technical study was conducted to establish a technical basis to address the first requirement of the Human Intrusion Standard (i.e., 10 CFR 63.321(a) [DIRS 178394]). That study, which is presented in Section 6.7.2, examined three general aspects of the hypothetical drilling intrusion, namely:

1. Capability of intact DSs and WPs to resist penetration by drilling technology typically used in groundwater exploration
2. Drilling operating characteristics (i.e., change-in-conditions) that would or would not indicate drill-bit impingement on a metallic anthropogenic structure

3. Degradation of DSs and WPs by nominal processes and likely disruptive events that could provide a physical pathway for a drill-bit to penetrate the WP without recognition by the drillers.

With regard to the scenario conditions, likely processes and events refers to those events with a frequency of occurrence greater than or equal to 10^{-5} /yr, as specified by the NRC regulation.

The technical basis, which is summarized herein, was used to develop a bounding estimate of 200,000 years for the earliest time of WP penetration (without recognition by drillers). That bounding estimate was used in the probabilistic projections of annual dose to the RMEI, which are presented in Section 8.1.3.2

Capability of Intact DSs and WPs to Resist Drilling Penetration

The 15-mm (0.59 in.) thick Titanium Grade 7 DS plate and 25-mm (0.98 in.) thick Alloy 22 WP outer barrier constitutes a significant double barrier to potential groundwater drilling intrusion. In the case of the CSNF package, which is a canister within a canister, there would be an additional 50.8-mm (2 in.) thick stainless steel barrier to resisting potential drilling intrusion. The capability of the DS and WP barriers to resist a drilling penetration was examined from two perspectives: (1) drill-bits typically used in groundwater exploration in rock formations, and (2) scenario of drill string free-fall through an open drift and potentially penetrating a DS and WP.

Drilling Technology—Drilling technology that would be used in drilling water wells through welded tuff at Yucca Mountain would typically utilize roller bits (IADC 1992 [DIRS 155232]) to drill through the rock formations. Roller bits are very effective in drilling through welded tuff formations and basically cause brittle failure of the rock matrix and break and crush the rock to facilitate removal. This type of drill-bit, however, would be ineffective in attempting to bore through the metal barriers because of the high compressive strength and ductility of titanium and stainless steel. Penetration of these intact metal barriers would require a drill-bit designed to induce metal failure; such types of drill-bits are not typically used in exploratory drilling for water in the Southwestern United States. Moreover, attempts to drill through intact DSs and/or WPs would result in significant changes to the drilling operation characteristics, which would be clear indicators of the drill-string encountering a metallic anthropogenic structure.

Drill-String Free-Fall and Impingement on WP—A drill-string entering an open emplacement drift could experience a free-fall and, in turn, directly impacting the intact DS and WP. The potential for WP penetration was evaluated for the case of a 300 m, 14-metric ton drill-string assembly (i.e., drill pipe, drill collar, and drill-bit) assuming no DS present with drop heights of 1.8 m (i.e., approximate distance to CSNF) and 1.6 m (i.e., approximate distance to CDSP). A bound for the maximum impact velocity of such a drill-string free-fall was estimated to be about 5.9 m/s for both the CSNF and CDSP packages.

The potential for rupture of the WP outer barrier for the above conditions was evaluated by utilizing calculational results from detailed structural analyses for WP-to-pallet impacts (SNL 2007 [DIRS 178851], Section 6.3.2), which provide strain-based rupture conditions as a function of impact velocities. This comparative analysis indicates that the drill-string impact would not penetrate an intact WP. A more likely outcome of a drill-string free-fall would be impingement on and deformation of the DS. In the unlikely event that the DS was actually penetrated by the drill-string, it would absorb the impact energy, which would reduce the impact velocity, and further reduce the potential for penetration of the WP.

Drilling Operating Characteristics

Drilling operations generally involve monitoring of the bit operating conditions as the drill-string penetrates the rock formations. Typical characteristics that are monitored by drillers include circulating fluid properties and rates, drilling stability, bit weight, and rotary speed. Significant changes in any of these types of operating characteristics generally prompt the driller to examine and resolve the change in operating characteristics. In the case of a drill-string entering an open emplacement drift and impinging the DS and/or WP, the following changes in the operating characteristics would likely occur:

- Loss of circulation fluid
- Sudden changes in rotary speed
- Rapid increase in drill-string vibration.

Moreover, driller attempts to correct volumetric problems with the circulating fluid (i.e., by spot cementing and setting cases) would be largely thwarted by the large drift volume and physical presence of the DS and WP. These problems would likely prompt the driller to remove the drill-string from the borehole and examine the drill-bit. Such diagnostic actions by the driller would lead to recognition that a metallic anthropogenic structure had been encountered.

Under conditions of a drift collapse, however, changes in some operational characteristics could be muted and, therefore, less recognizable. For example, the loss of circulating fluid in a collapsed drift (i.e., filled with rubble) would still occur but could be mitigated by the driller. Changes in drill-bit rotation speed would be unaffected until the drill-bit encountered the DS, at which time the drill-string would exhibit vibration. Drift collapse would primarily occur as a result of cumulative effects of seismic events (i.e., vibratory ground motion) in the lithophysal regions of the repository. Such drift collapse events are currently projected to occur in the post-10,000-year time period.

Degradation of Drip Shields and Waste Packages by Nominal and Disruptive Scenario Conditions

The study considered four scenarios for WP and DS degradation: (1) nominal degradation of the DSs and WPs, (2) DS early failures (i.e., having an undetected defect that could cause an early failure), (3) DS and WP failures as a result of igneous intrusion disruptive events, and (4) DS and

WP failures as a result of seismic events (i.e., seismic ground motion and seismic fault displacement). Of these scenario conditions, three were screened-out because their frequency of occurrence was less than the NRC cut-off of 1 chance in 10 in 10,000 years or $10^{-5}/\text{yr}$, namely:

- Early Failure: The probability that a drilling intrusion encounters a DS early failure occurring at the same location is estimated to be 7×10^{-6}
- Igneous Intrusion: Exceedance frequency of igneous events in the repository footprint estimated to be less than $10^{-5}/\text{yr}$ but greater than $10^{-8}/\text{yr}$
- Seismic Fault displacement: Exceedance frequency of seismic fault displacement in the repository footprint is estimated to be less than $2.5 \times 10^{-7}/\text{yr}$.
- Seismic Ground Motion: (1) peak ground velocities, with exceedance frequencies greater than or equal to $10^{-5}/\text{yr}$, are relatively low and unlikely to induce dynamic loads sufficient to cause DS plate failure, and (2) exceedance frequency of seismic-induced large block movements, which could rupture DSs and damage WPs, are less than $1.17 \times 10^{-6}/\text{yr}$ (DTN MO0712PBANLNWP.000_R0 [DIRS 184664], file *Nonlith LC Calculation Rev03.xmcd*).

Thus, nominal degradation processes are the most probable way that a penetration (i.e., opening) could develop in the DSs and WPs that could potentially serve as a physical pathway to groundwater drilling—facilitating a penetration without recognition by the drillers.

In the case of the DSs, thinning of the titanium DS plate would occur over geologic time periods. Over a period of 10,000 years, for example, the extent of thinning of the 15-mm-thick DS plate by general corrosion is projected to be approximately 0.66 mm or a thinning of about 4 percent of the total thickness. This projection is based on a high corrosion rate corresponding to a 0.9999 quantile (Section 6.3.5.1.3). Assuming this same corrosion rate, the DS plate failure time is projected to occur at about 230,000 years; this estimate is corroborated by more detailed probabilistic projections for nominal conditions that calculate the first DS plate failure occurring at about 270,000 years (Figure 8.1-4). Those detailed probabilistic projections, also for nominal conditions, indicate that the mean failure time of the first penetration of WP outer barrier by general corrosion patches occurs at about 440,000 years (DTN: MO0709TSPA WPDS.000 [DIRS 183170]).

Based on the above considerations and evaluations, the estimate of the earliest time a driller could penetrate a WP, without recognition by the driller, was taken to be 200,000 years. This time is a conservative bound to DS fail time and neglects the lifetime of the WPs.

8.1.3.2 Projections of Annual Doses for Human Intrusion

To address the second requirement of the Human Intrusion Standard (10 CFR 63.321(b) [DIRS 178394]), a probabilistic TSPA methodology, analogous to that used to demonstrate

performance with the Individual Protection and Groundwater Protection Standards, was used to make projections of the annual dose. The calculations of annual dose to the RMEI were performed for all environmental pathways, as specified in 10 CFR 63.321(c) [DIRS 178394]. Based on the analysis described above, the earliest time after disposal for the drilling intrusion was taken to be 200,000 years.

As described in Section 6.1.2.5, the type of WP (i.e., CSNF or CDSP) assumed to be penetrated is sampled in the analysis so that probabilistic projections of annual dose reflect the radionuclide releases from the various waste forms (Section 6.3.5). Based on the proportion of WP types, the probability of sampling a CSNF WP is ~ 0.7 , whereas the probability of selecting a CDSP WP is ~ 0.3 . The location of the penetration in the repository footprint was also sampled so as to reflect the range of percolation fluxes that would induce waste dissolution and releases.

The probabilistic projections of expected annual dose for the Human Intrusion Scenario are presented on Figure 8.1-12. The plots show the curves for the mean, median, and 5th and 95th percentiles of the distribution of expected annual doses for the period of geologic stability. The peak of the mean annual dose to the RMEI occurs within a few thousand years after inclusion. The peak mean and median annual individual doses are projected to be less than 0.008 mrem and 0.006 mrem, respectively, well below the regulatory limit of 350 mrem. The annual doses at the 5th and 95th percentiles are calculated to be approximately 0.001 mrem and 0.017 mrem, respectively.

The contributions of individual radionuclides to the total mean annual dose for the Human Intrusion Scenario for 1,000,000 years after repository closure, is shown on Figure 8.1-13. ^{99}Tc (half-life 2.13×10^5 yrs), ^{129}I (half-life 1.57×10^7 yrs), ^{36}Cl (half-life 3.01×10^5 yrs), and ^{237}Np (half-life 2.14×10^6 yrs) dominate the mean annual dose from 200,000 years to about 300,000 years; thereafter, dominance then transitions to ^{79}Se (half-life 2.95×10^5 yrs), ^{135}Cs (half-life 2.30×10^6 yrs), ^{242}Pu (half-life 3.75×10^5 yrs), and ^{237}Np (half-life 2.14×10^6 yrs). The ^{242}Pu , which is transported through the SZ in both dissolved and colloidal form, dominates the peak mean annual dose at 1,000,000 years.

Thus, these projections demonstrate that there is a reasonable expectation that the mean annual doses to the RMEI would be well below the limits for the Human Intrusion Standard. Moreover, the projections indicate that the system of multiple barriers would be sufficiently robust and resilient to limit annual doses for the prescribed Human Intrusion Scenario.

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Table 8.1-1. Performance Demonstration Results for Individual Protection Standard

Time After Closure (yrs)	Projected Peak Mean Annual Dose (mrem)	Time of Peak Mean Annual Dose (yr)	Projected Peak Median Annual Dose (mrem)	Time of Peak Median Annual Dose (yr)	Limit for Annual Dose (mrem)
10,000	0.24	10,000	0.12	10,000	15 (mean)
1,000,000	2.30	1,000,000	0.99	~760,000	350 (median)

Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]; Regulatory Limits from 10 CFR 63.311(a) ([DIRS 178394])

Table 8.1-2. Performance Demonstration Results for Groundwater Protection Standard

Type of Limit	Projected Peak Mean Activity Concentration or Annual Dose	Natural Background Level	Limit for Activity Concentration or Annual Dose
Combined Ra-226 & Ra-228	$<10^{-5}$ pCi/L	0.5 pCi/L	5 pCi/L
Gross Alpha Activity	$<10^{-4}$ pCi/L	0.5 pCi/L	15 pCi/L
Dose from Combined Beta & Photon Emitting Radionuclides	Whole Body ~ 0.04 mrem Thyroid ~ 0.17 mrem	Background level excluded in regulatory requirement	4 mrem

Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]; Regulatory Limits from 10 CFR 63.331 ([DIRS 180319])

Table 8.1-3. Performance Demonstration Results for Human Intrusion Standard with Drilling Event at 200,000 years After Closure

Time After Closure (yrs)	Projected Peak Mean Annual Dose (mrem)	Limit for Annual Dose (mrem)
10,000	0	15 (mean)
1,000,000	$< 10^{-2}$	350 (median)

Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]; Regulatory Limits from 10 CFR 63.321(b) ([DIRS 178394])

Table 8.1-4. Uncertainty in Projections of Peak Mean Annual Dose (mrem) for Individual Protection Standard

Time After Closure (yrs)	Mean	Median	5th Percentile	95th Percentile
10,000	0.24	0.12	6.43×10^{-3}	0.71
1,000,000	2.30	0.88	0.18	9.55

Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Table 8.1-5. Uncertainty Importance Ranking as a function of time for Key TSPA-LA Model Parameters

Time After Closure (yrs)	Two Most Important Parameters at the Selected Time	
3,000	SCCTHRP	IGRATE
5,000	SCCTHRP	IGRATE
10,000	SCCTHRP	IGRATE
125,000	IGRATE	SCCTHRP
250,000	WDGCA22	IGRATE
500,000	IGRATE	WDGCA22
1,000,000	IGRATE	WDGCA22

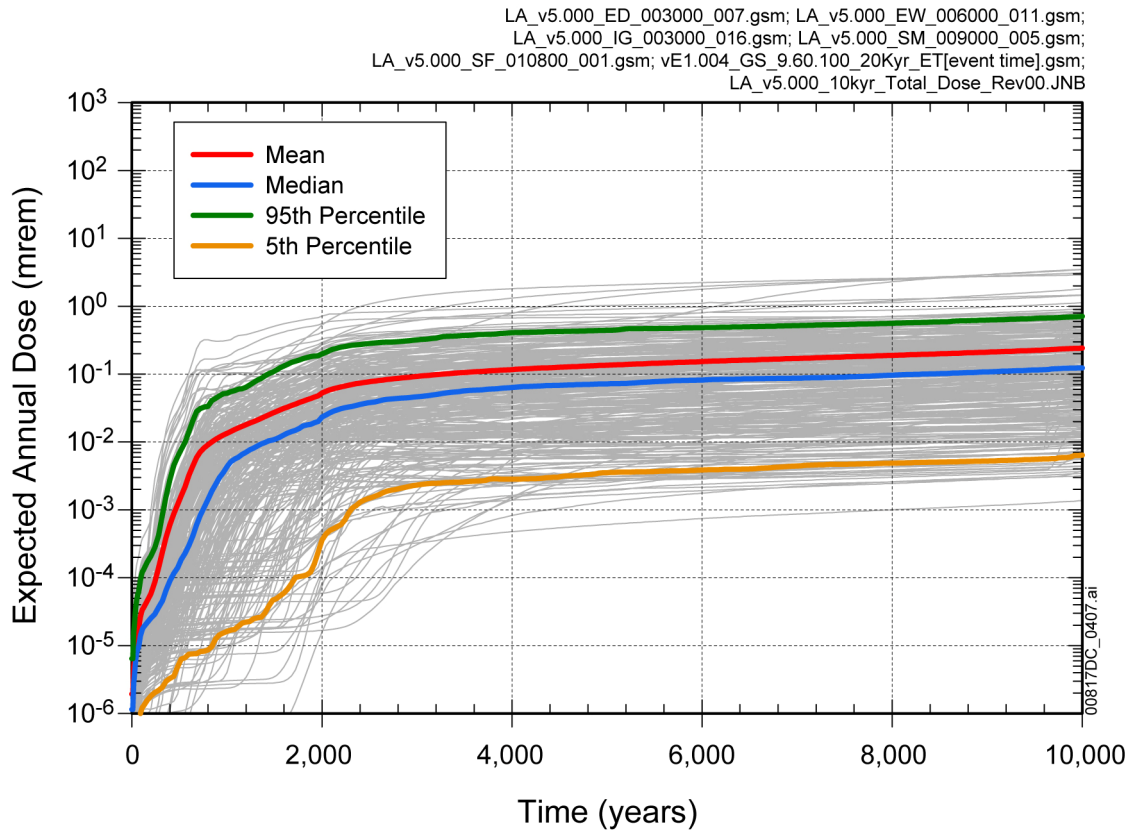
Source: Output DTN: MO0709TSPASENS.000 [DIRS 182982]

NOTE: IGRATE = occurrence rate of igneous events

SCCTHRP = SCC stress threshold

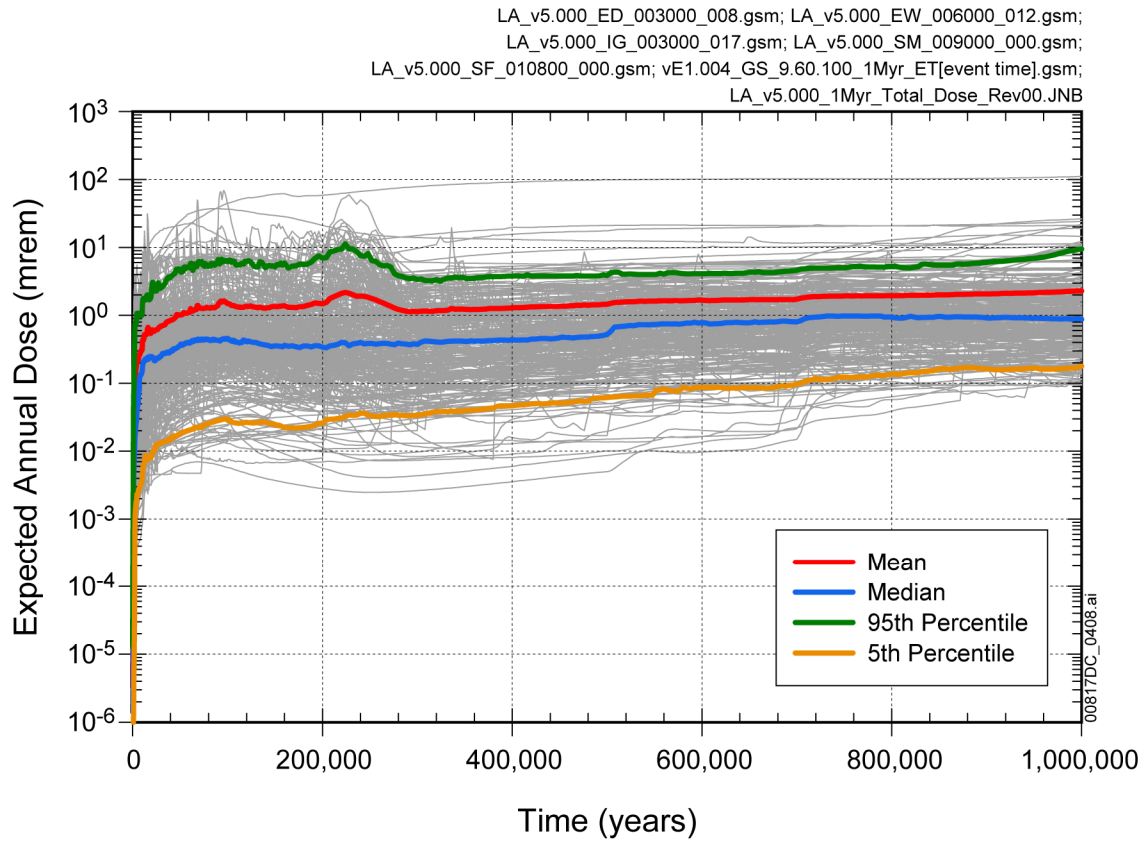
WDGCA22 = temperature dependence parameter for alloy 22
general corrosion rate

Importance related to expected dose to RMEI for all scenario
classes.



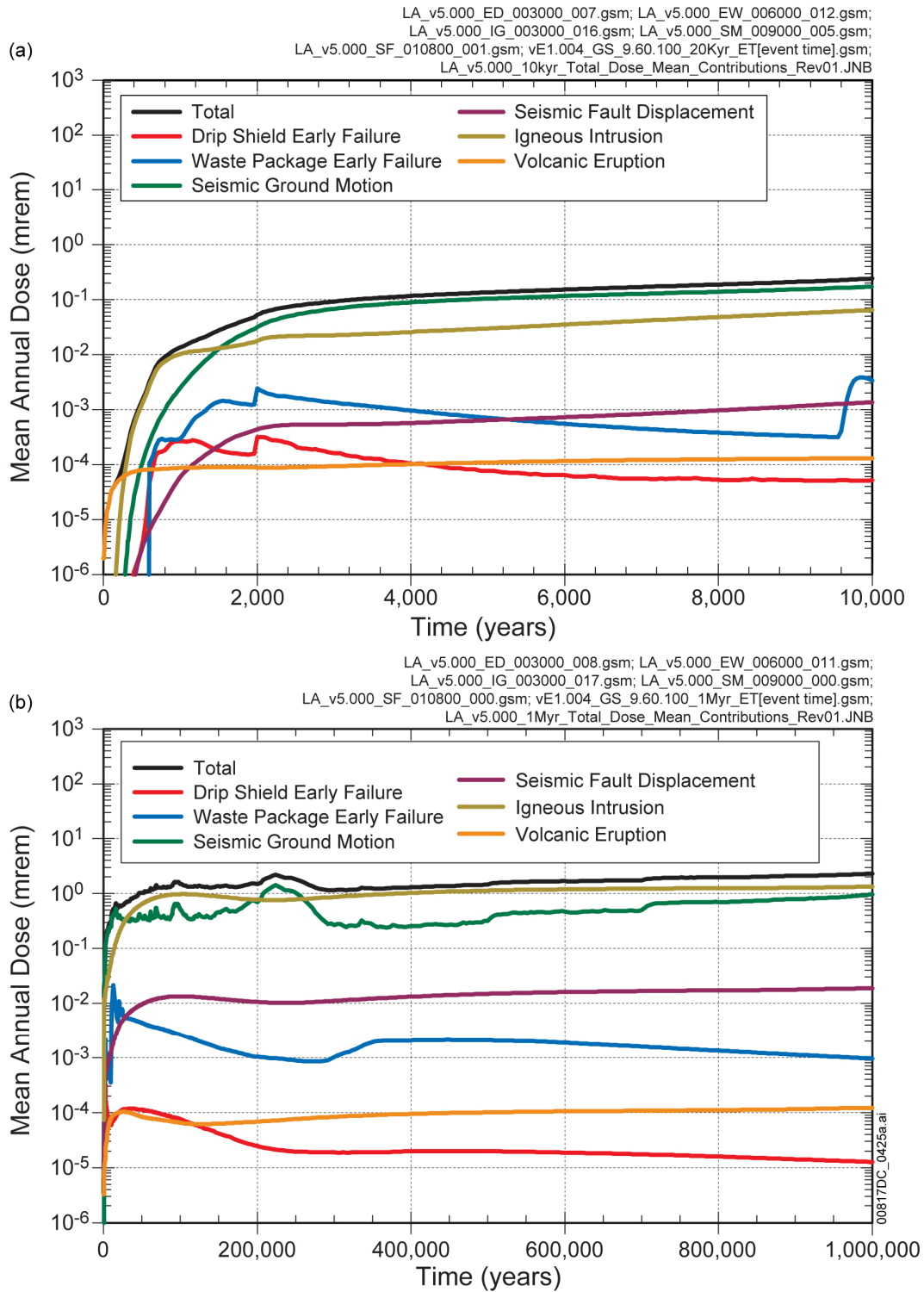
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-1. Probabilistic Projections of Total Expected Annual Dose for 10,000 Years after Closure



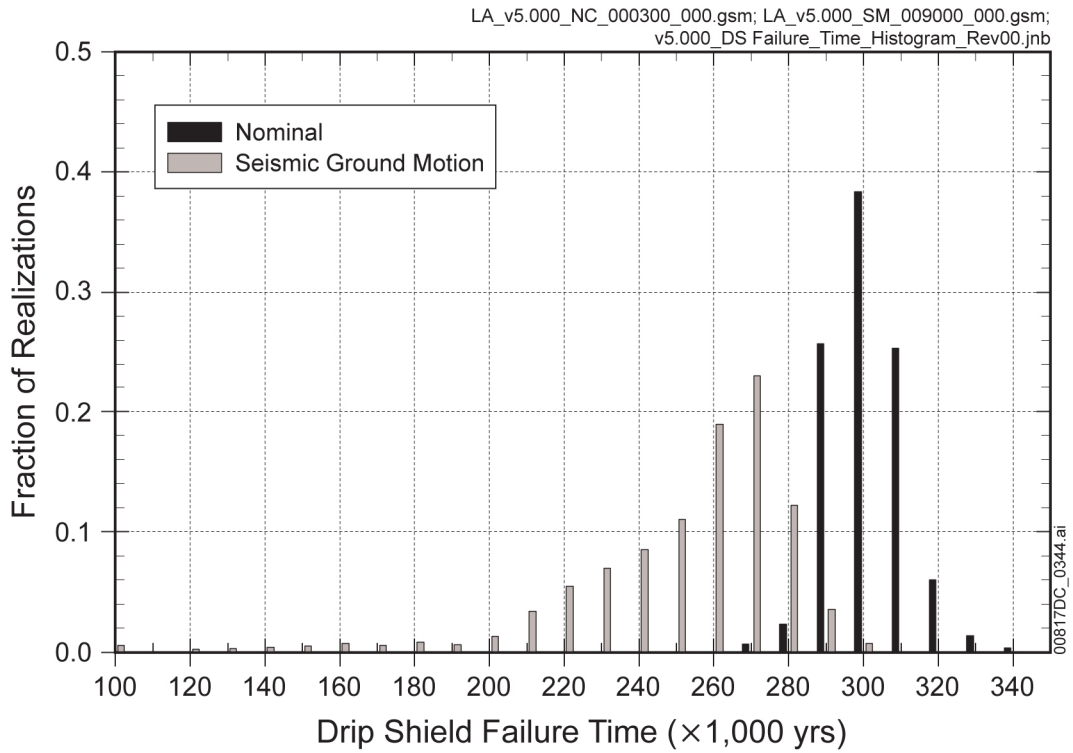
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-2. Probabilistic Projections of Total Expected Annual Dose for 1,000,000 Years after Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

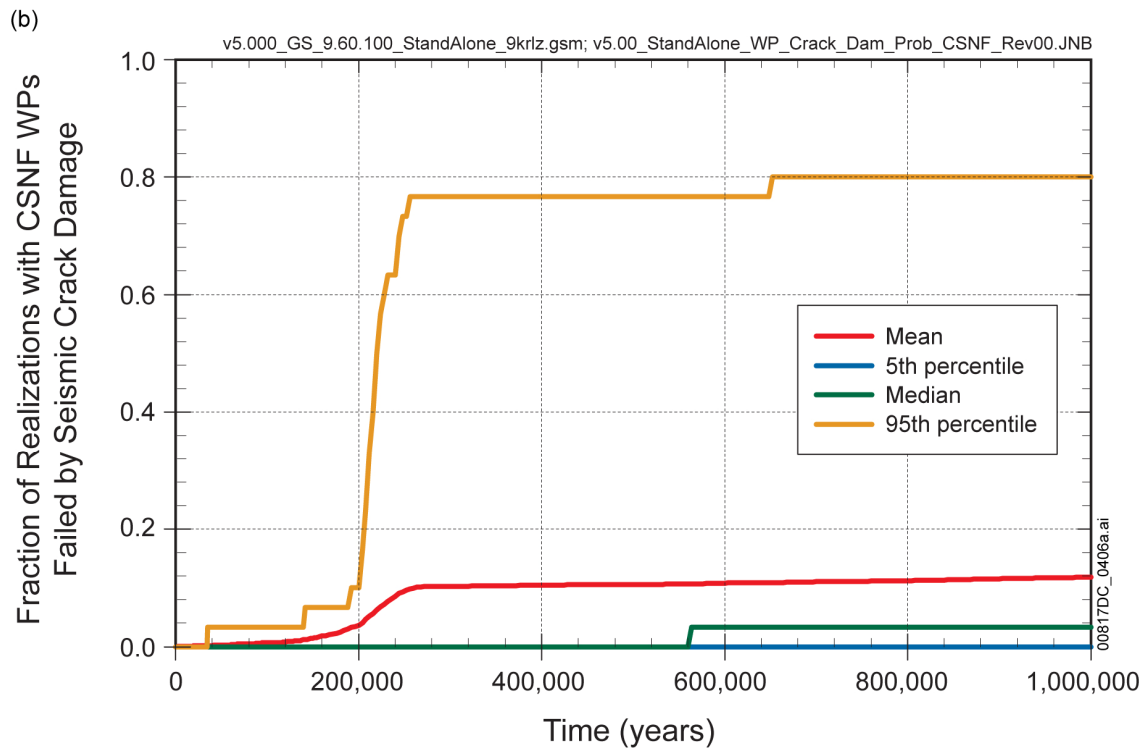
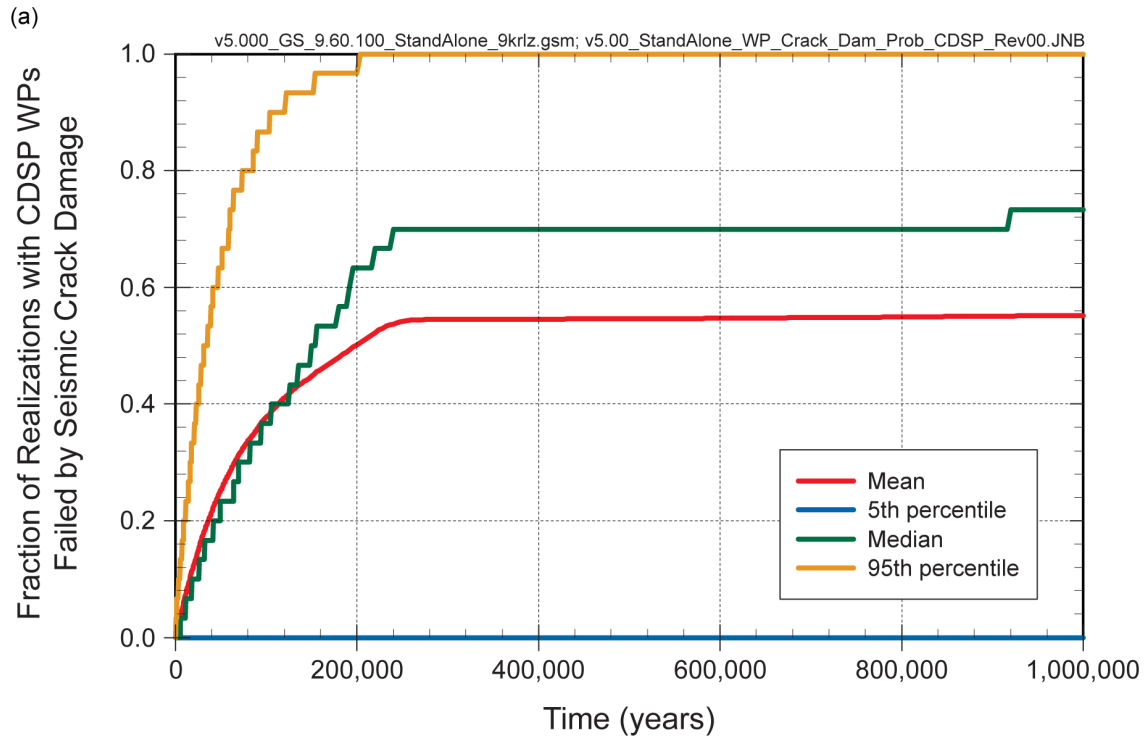
Figure 8.1-3. Relative Contributions of Scenario Modeling Cases to Total Mean Annual Dose for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

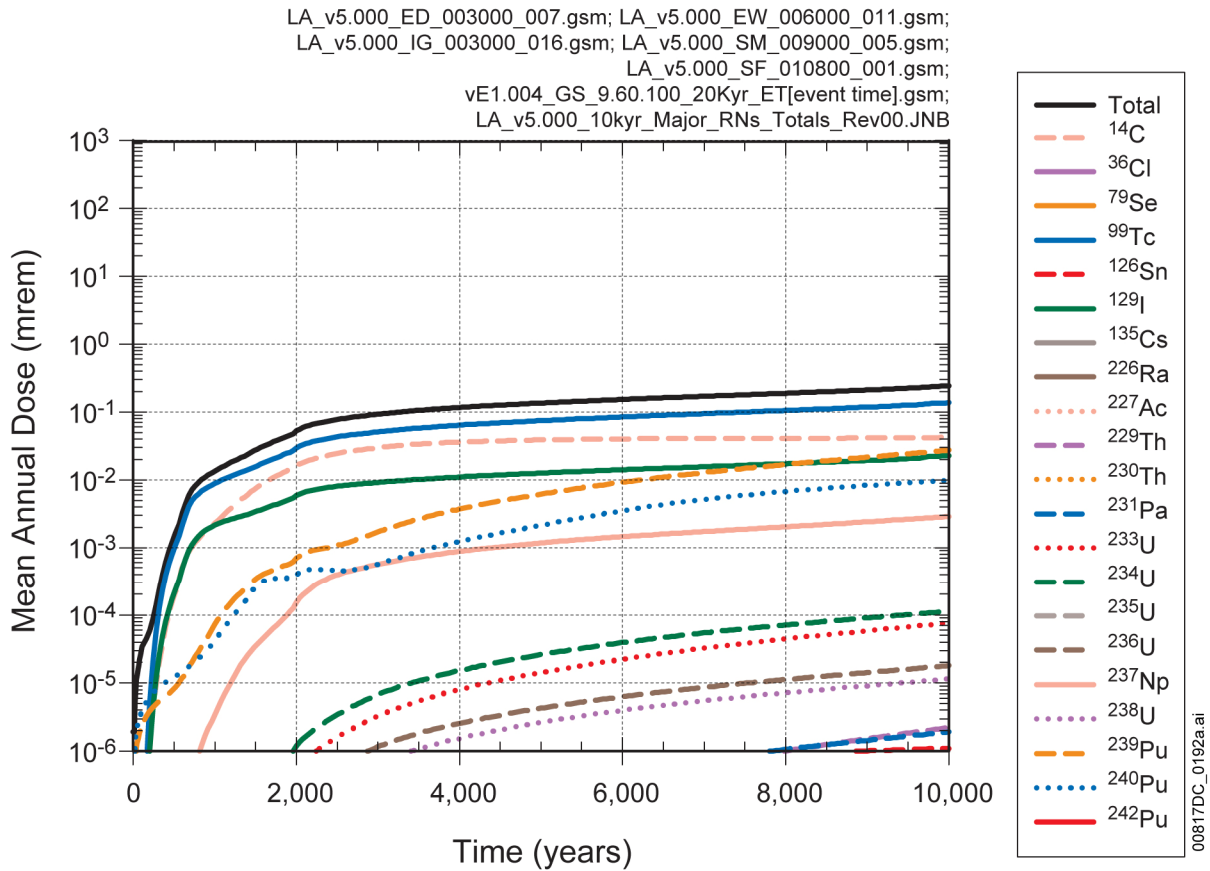
NOTE: Nominal failures are due to general corrosion. Seismic ground motion failures are caused by the combined effects of general corrosion, vibratory ground motion, and rockfall.

Figure 8.1-4. Histogram of Drip Shield Failure for the Nominal and Seismic Ground Motion Modeling Cases



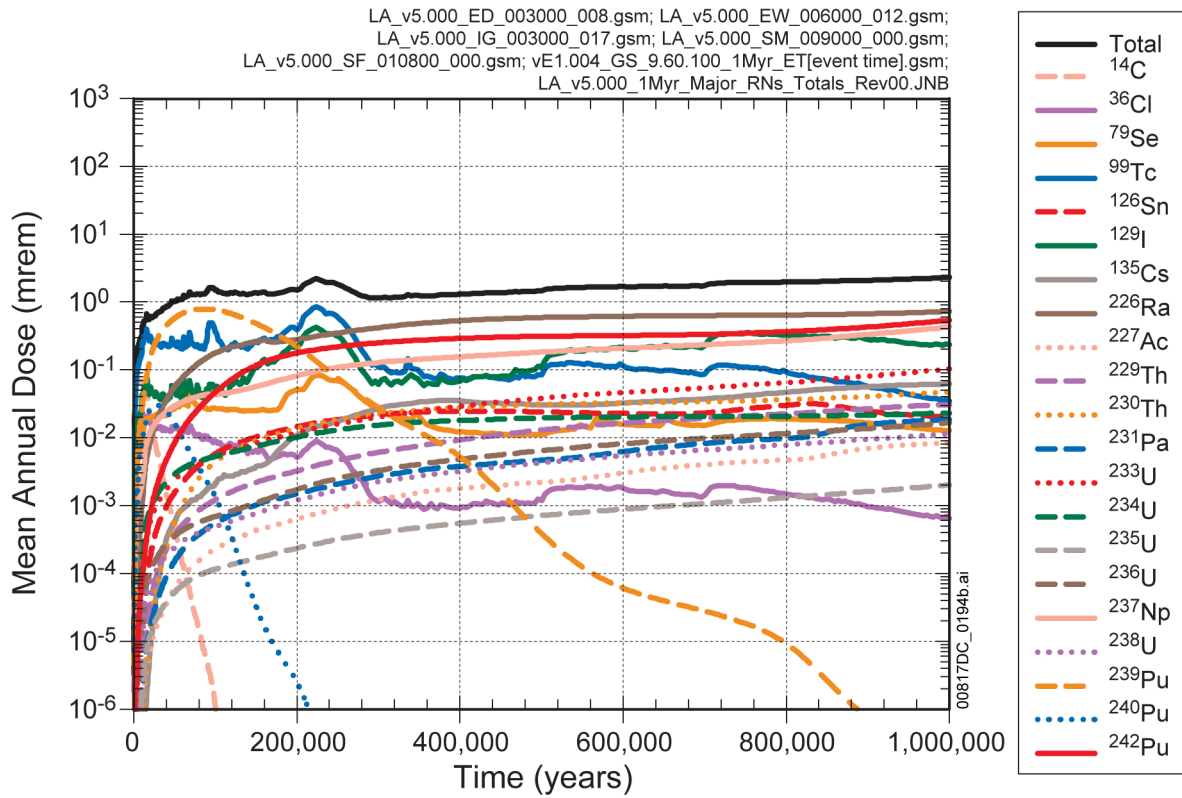
Source: Output DTNs: MO0709TSPAREGS.000 [DIRS 182976]; and MO0709TSPAWPDS.000 [DIRS 183170].

Figure 8.1-5. Fraction of (a) CDSP WPs and (b) CSNF WPs Failed for by Seismic Crack Damage as a Function of Time for Percolation Subregion 3



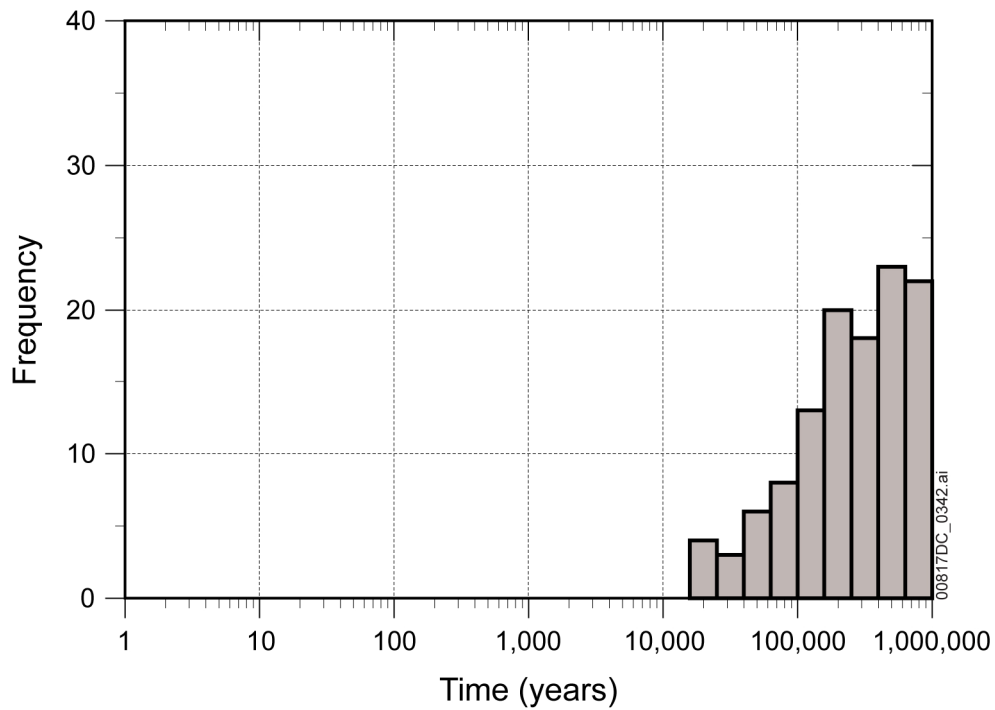
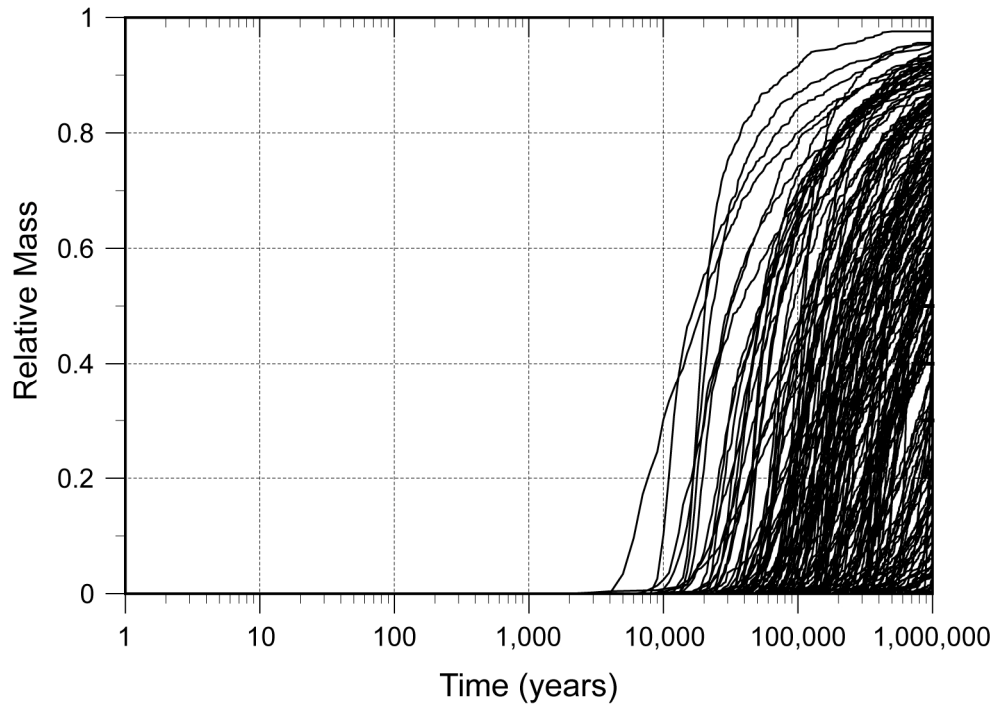
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-6. Contribution of Individual Radionuclides to Total Mean Annual Dose for 10,000 Years after Repository Closure



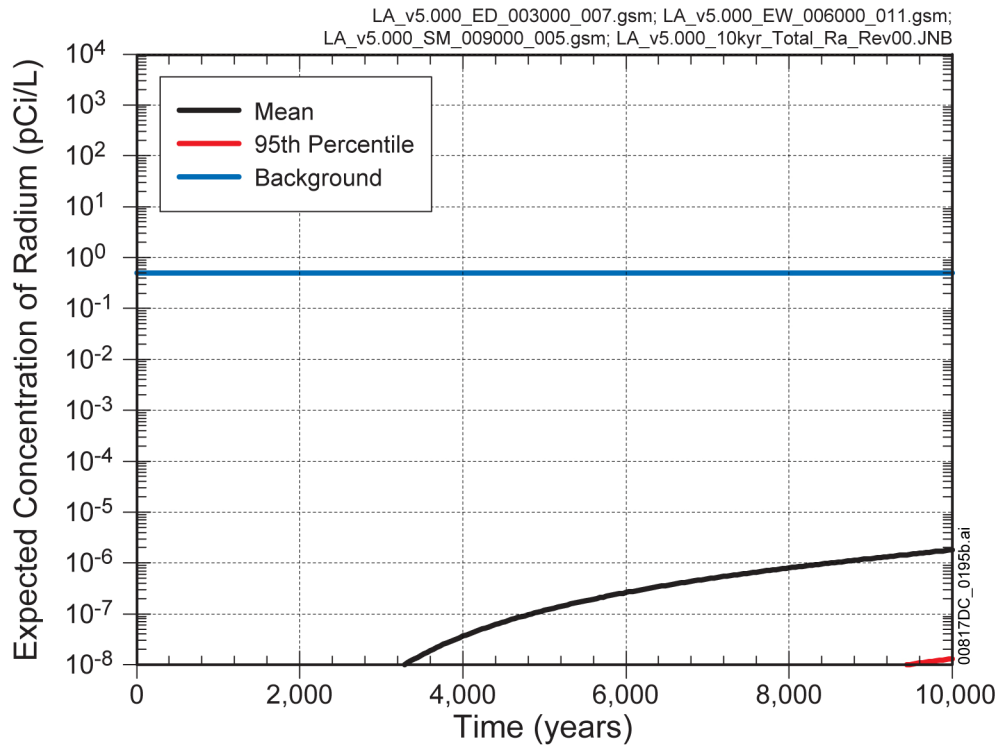
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-7. Contribution of Individual Radionuclides to Total Mean Annual Dose for 1,000,000 Years after Repository Closure



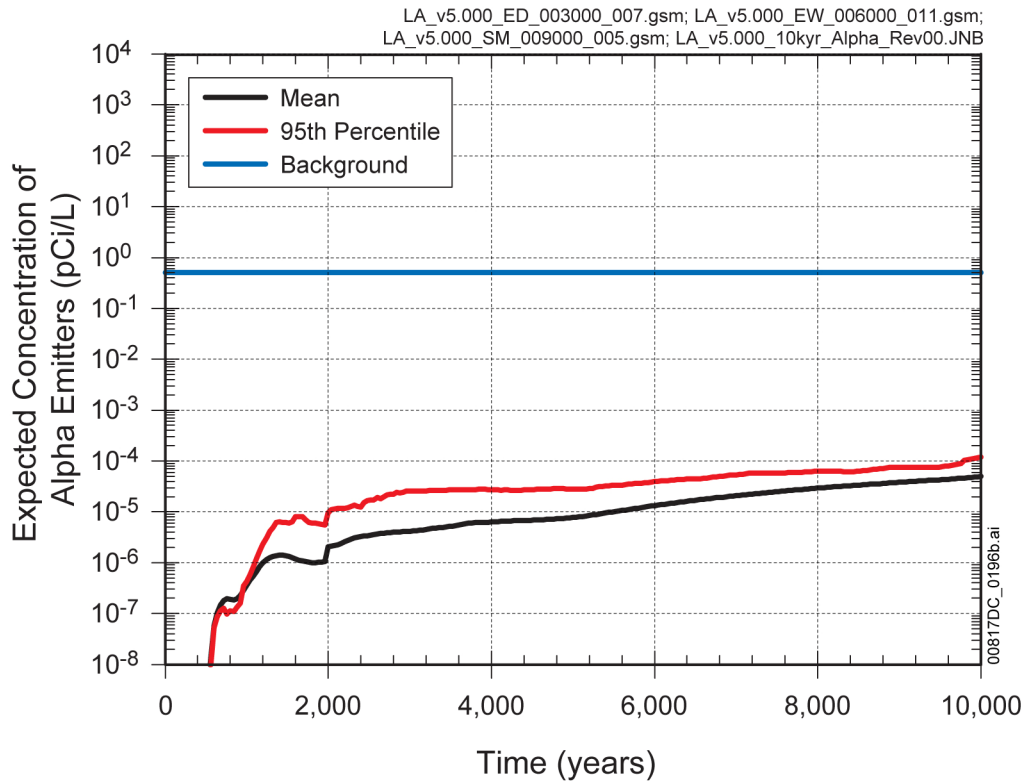
Sources: SNL 2007 [DIRS 183750], Figure 6-14[a].

Figure 8.1-8. Radium Mass Breakthrough Curves (upper) and Median Transport Times (lower) at the RMEI Location



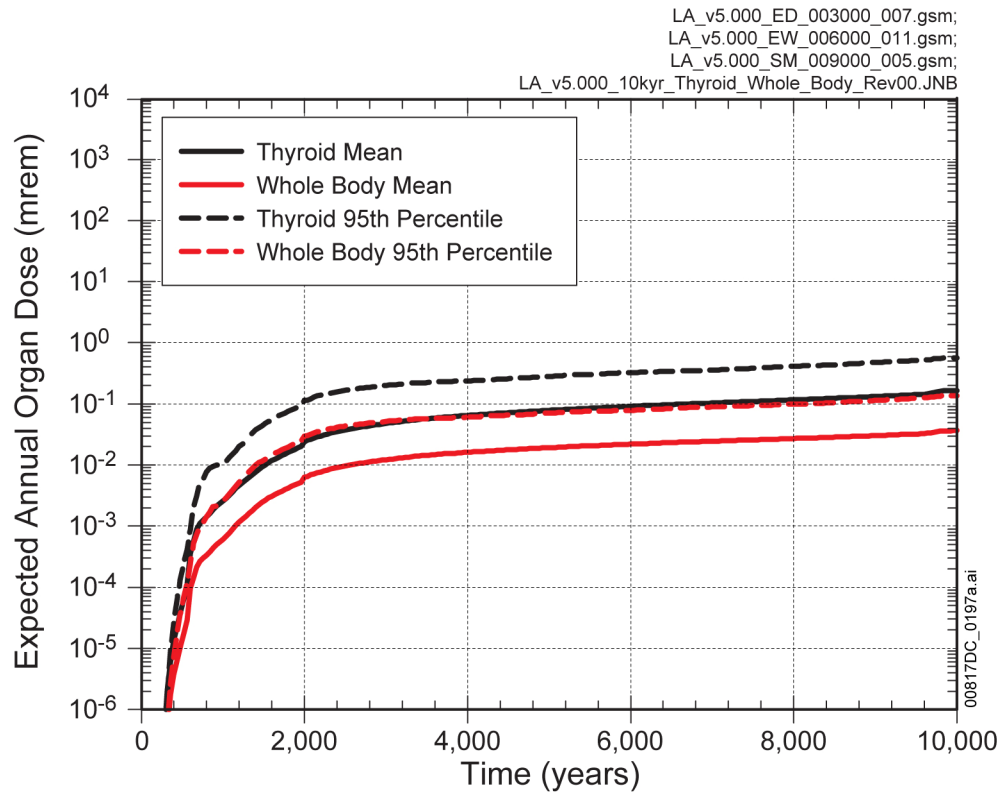
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-9. Probabilistic Projections of Activity Concentrations Total Radium (²²⁶Ra and ²²⁸Ra) in Groundwater, Excluding Natural Background, for 10,000 Years after Closure



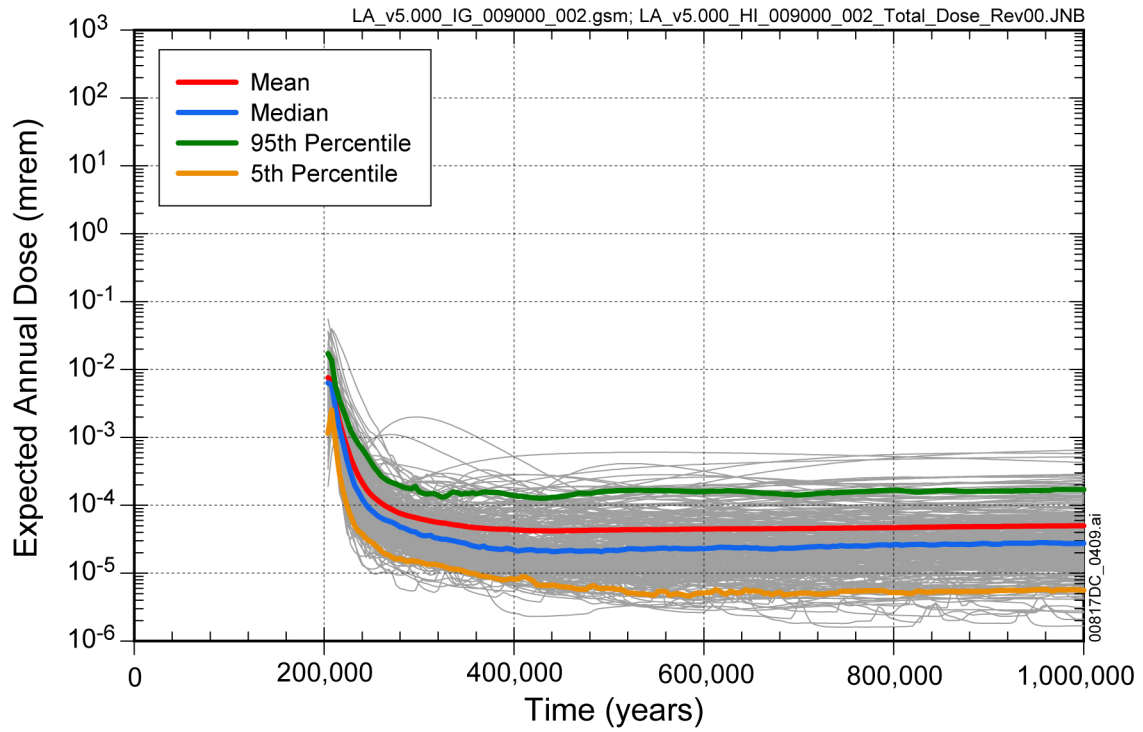
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-10. Probabilistic Projections of Activity Concentration of Gross Alpha and ²²⁶Ra (Excluding Radon and Uranium) in Groundwater for 10,000 Years after Closure



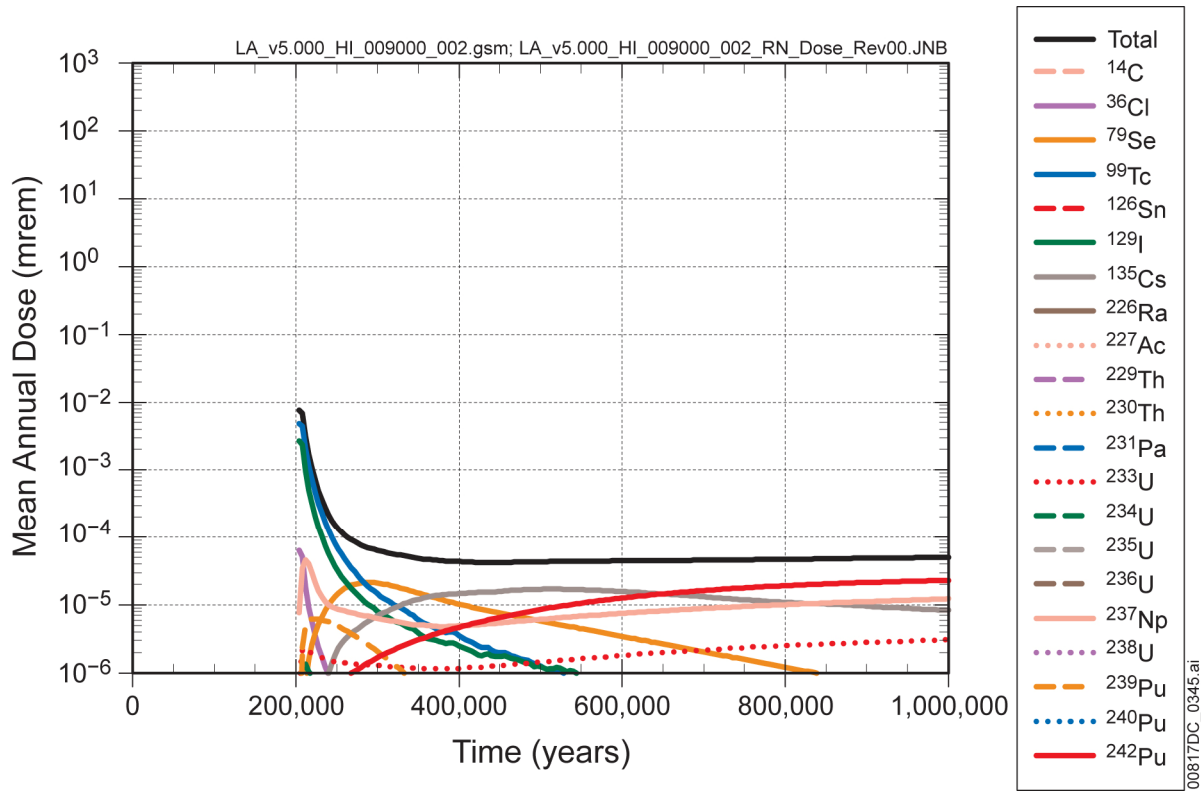
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-11. Probabilistic Projections of Annual Drinking Water Doses for Combined Beta and Photon Emitting Radionuclides for 10,000 Years after Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-12. Probabilistic Projections for Expected Annual Doses for the Human Intrusion Scenario for 1,000,000 Years after Closure with Drilling Event at 200,000 Years



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976].

Figure 8.1-13. Contribution of Individual Radionuclides to Mean Annual Dose for the Human Intrusion Scenario for 1,000,000 Years after Repository Closure

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8.2 PROJECTIONS FOR INDIVIDUAL MODELING CASES

To clarify and help explain the performance demonstration for the EPA Individual Protection Standard, this section presents the probabilistic projections for the individual modeling cases. As indicated previously, the projections of total expected annual dose to the RMEI are based on a summation (Equation 8.1.1-1) of the expected annual doses over scenario modeling cases. The individual scenario class modeling cases include: (1) Nominal, (2) Waste Package EF, (3) Drip Shield EF, (4) Igneous Intrusion, (5) Volcanic Eruption, (6) Seismic GM, and (7) Seismic FD. As stated earlier, the Nominal Modeling Case is not explicitly included in the sum of total expected annual doses because it is embedded in the Seismic GM Modeling Case.

The performance projections presented herein for these modeling cases are organized in a manner so as to address the five fundamental performance questions identified earlier in Section 8.1. The supporting technical basis for the inferences and observations made regarding characteristics important to postclosure performance are also presented or cited. As will be explained in subsections that follow, the examination of the expected annual doses for the individual modeling cases leads to the following fundamental observations about relative importance of these modeling cases:

- Total mean annual doses to the RMEI for time periods after a few hundred years are largely dominated by Seismic GM and Igneous Intrusion Modeling Cases
- Mean annual doses for Waste Package EF and Seismic FD Modeling Cases are similar and estimated to contribute about 10^{-2} mrem or less to the total mean annual dose
- Mean annual doses for the Drip Shield EF and Volcanic Eruption Modeling Cases are estimated to contribute less than 10^{-3} mrem to the total mean annual dose.

For the 10,000-year period, the Seismic GM Modeling Case contributed about 71 percent and the Igneous Intrusion Modeling Case contributed about 27 percent of the peak total mean annual dose (i.e., 0.2 mrem). In contrast, for the post-10,000-year period, the Seismic GM and Igneous Intrusion Modeling Cases contributed almost equally to the peak total mean annual dose of 1 to 2 mrem.

For the Seismic GM Modeling Case, the important phenomenon is the accumulated WP damage induced by the vibratory ground motion. As indicated on Figure 8.1-5, the CDSP WPs, which contain the defense high-level waste (DHLW) glass and DOE spent nuclear fuel (DSNF), have a higher probability of failure from vibratory ground motion than the more robust CSNF WPs. The probabilistic projections for this modeling case are presented later in Section 8.2.4.

With regard to the Igneous Intrusion Modeling Case, it is important to understand that this particular modeling case represents the consequences of unlikely disruptive events with a mean annual frequency estimated to be 1.7×10^{-8} per year (Section 6.5.1.1). Moreover, the TSPA submodel for these phenomena assumes that all the 3,416 CDSP WPs and 8,213 CSNF WPs (Table 6.3.7-1) are completely failed by the hypothetical magmatic intrusion.

8.2.1 Nominal Modeling Case

The Nominal Scenario Class consists of a single modeling case that is a representation of a set of possible repository futures in which: (1) disruptive events and conditions and (2) DS and WP early failures are excluded. Moreover, the Nominal Modeling Case serves as a 'reference system state' from which all other modeling cases are developed. The system attributes for the Nominal Modeling Case are defined in terms of likely and unlikely FEPs that describe natural evolution of the natural and engineered barriers over geologic time frames, in the absence of disruptive events. That evolution includes changes in climate, infiltration and seepage into the drifts, exposure of the EBS components to water, progressive chemical degradation, and failure of the DSs and WPs, with eventual mobilization, release, and migration of radionuclides; this evolution of the reference state is included in the other modeling cases. The technical description for the representation of this nominal evolution is summarized in Section 6.3.

From a high-level point of view, this modeling case basically describes the evolution of the nuclear waste and repository system in the absence of disruptive events and processes. The system response for the Nominal Modeling Case is a function of a spectrum of processes:

1. Climate changes and attendant changes in net infiltration into the Upper Natural Barrier
2. Seepage of percolating water into the repository drifts
3. Changes to the water chemistry induced by repository heating
4. Progressive degradation of the engineered barriers by corrosion processes (i.e., general corrosion, localized corrosion, and stress corrosion cracking [SCC]) and failure
5. Accumulation of moisture inside the WPs, degradation and dissolution of the waste forms, and mobilization of the radionuclides
6. Migration of dissolved and colloidal-phase radionuclides from the EBS
7. Water flow and radionuclide transport through the Lower Natural Barrier and to the accessible environment
8. Water withdrawal and exposure to the RMEI.

The detailed technical basis (i.e., conceptual models, model abstractions, supporting data, and TSPA model parameter values and their uncertainty characterizations) for the Nominal Modeling Case is documented in numerous analysis and/or modeling reports.

The probabilistic projections of expected annual dose for this modeling case are shown on Figure 8.2-1. As can be seen from this plot, there are no doses to the RMEI in the 10,000-year regulatory time period. This is consistent with calculations for the EBS performance that indicate no realizations of DS failures in the 10,000-year time period (Section 8.1, Figure 8.1-4 and DTN: MO0709TSPA WPDS.000 [DIRS 183170]). With the DSs intact, the WPs would experience very little degradation by general corrosion in a humid-air environment.

The first realization with a failed (i.e., creation of a breach) CDSP or CSNF WP by nominal corrosion processes is projected to occur between 100,000 years to 150,000 years. The WP failure mechanism is typically SCC of the closure-lid welds. The extent of cracking gradually increases with time, leading to the ingress of moisture inside the WP and attendant diffusive radionuclide releases. Realizations with WP failures by corrosion penetrations through the Alloy 22 outer barrier are observed at about 450,000 years. By 1,000,000 years, approximately 60 percent of the WPs are projected to be failed by SCC and about 10 percent by general corrosion breaches. The projections indicate that the DSs are intact until about 270,000 years (Figure 8.1-4), with failure occurring by thinning of the 15-mm thick titanium DS (Section 6.3.5.1). All the DSs are projected to be completely failed by about 340,000 years.

The probabilistic projections for the post-10,000-year period indicate that there is no radiologic exposure until after about 140,000 years after closure. The peak mean and median annual doses for this reference modeling case are 0.5 mrem and 0.3 mrem, respectively. The peak mean annual dose at the RMEI location occurs at about 730,000 years, with two radionuclides dominating the peak annual dose. These radionuclides are the highly soluble, long-lived, and mobile radionuclide species ^{129}I and ^{99}Tc . The second order contributors to this peak dose value are ^{135}Cs and ^{79}Se (Figure 8.2-2). This peak is a modeling artifact driven by the time stepping and WP failure methodology but is similar to the overall peak at 1,000,000 years.

It is very important to differentiate the Nominal Modeling Case results from those for other modeling cases. The Nominal Modeling Case projection of mean annual dose should not be taken as a representation of compliance with radiation protection limits. The mean annual dose for the Nominal Modeling Case is not included in the calculation of the total mean annual doses for the 10,000-year period because there are no projected WP failures by nominal degradation processes in this time period. Rather, the effect of nominal DS and WP corrosion processes for the post-10,000-year period are accounted for in the Seismic GM Modeling Case. The Nominal Modeling Case only provides a reference system state for comparison with the other six modeling cases.

8.2.2 Early Failure Scenario Class Modeling Cases

The projections for early failure modeling cases demonstrate postclosure performance for conditions of DS and WP early failures. As described in Section 6.4, DS and WP early failures are generally attributed to the presence of undetected flaws (e.g., weld flaws and improper weld filler material) and manufacturing defects. Such undetected flaws and defects could possibly be introduced during manufacturing or construction of the barriers, or possibly during handling and emplacement (e.g., damaged welds). In the case of a DS, this type of defect would diminish the DS's ability to withstand the dynamic and static loadings; however, such defects are treated in the TSPA-LA Model as having failed immediately. Similarly, a WP with a defective closure-lid weld, for example, would lead to a shorter period of containment than for nominal performance, but also is treated as an immediate failure at the time of closure.

The DS and WP early failures were simulated in the TSPA-LA Model as a random process described by a Poisson distribution; thus, the number of failed DSs and WPs is a random variable and is sampled from applicable probability distributions (Sections 6.4.1 and 6.4.2). The

following information and calculations (Sections 6.4.1 and 6.4.2, Tables 6.4-1 and 6.4-2) are useful in developing risk insights to DS and WP early failures:

Drip Shield Early Failure

- Probability of 1 or more DS early failures: 0.0166
- Expected number of DS early failures: 0.0181
- Expected number of DS early failures conditional on one or more occurring: 1.09
- Mean probability of DS early failure: 1.56×10^{-6}

Waste Package Early Failure

- Probability of 1 or more WP early failures: 0.442
- Expected number of WP early failures: 1.09
- Expected number of WP early failures conditional on one or more occurring: 2.46
- Mean probability of WP early failure: 9.36×10^{-5} .

The results listed above are documented in output DTN: MO0707WPDRIPSD.000 [DIRS 183005].

For these two early failure modeling cases, the repository system response is developed by applying the process models described for the Nominal Modeling Case to the DSs and WPs that are affected by early failure. Generally speaking, the two early failure modeling cases describe the repository system response as a function of:

1. Changes of climate states, net infiltration, seepage into the drift, and water chemistry as represented in the Nominal Modeling Case (except no degradation of WPs by corrosion)
2. Total number of DS early failures in each realization
3. Total number of CDSP WP and CSNF WP early failures in each realization
4. Ingress of water to the failed WPs, mobilization, and the eventual release of radionuclides from the EBS
5. Water flow and transport of dissolved and colloidal phase radionuclides through the Lower Natural Barrier and to the accessible environment
6. Water withdrawal and radiologic exposure at the RMEI location.

It is important to emphasize that WPs associated with early failed DSs are also assumed to be failed in the TSPA-LA Model. The associated radionuclide releases from these early failed DSs and WPs are assumed to occur at the time of repository closure.

The models and TSPA methodology used to simulate the occurrence of early failures and their impacts on the performance of the affected WPs and DSs are presented in Section 6.4. Section 6.4 also summarizes the supporting technical basis for the abstraction of early failures.

Model parameter values and their uncertainty characterizations are documented in *Analysis of Mechanisms for Early Waste Package/Drip Shield Failure* (SNL 2007 [DIRS 178765]).

8.2.2.1 Drip Shield Early Failure Modeling Case

As indicated by probabilistic analysis of early failures, the total number of DS early failures is estimated to be very small. For example, the probability of three or more DS early failures is about 2×10^{-4} (Table 6.4-1). As implemented in the TSPA-LA Model, the DS early failures are accounted for by simply removing the DS as a barrier to seepage for a given realization. In its implementation, the number of 'n' failed DSs means that the associated 'n' WPs are also assumed to have failed; thus, 'n' failed DSs define the inventory at risk for this modeling case.

The expected annual dose histories for the Drip Shield EF Modeling Case are shown in Figure 8.2-3; this plot shows multi-realization projections for both the (a) 10,000-year period after closure, and (b) post-10,000 years to 1,000,000 years. The mean, median, and 5th and 95th percentile curves are superimposed on the plot to illustrate the central tendency and uncertainty. The uncertainty or spread reflects the epistemic uncertainty in the TSPA-LA Model parameters and representation of the evolution of the future conditions. The projections for the first 10,000 years show an early peak around 1,000 years (not the peak annual dose for the time period, however), which is primarily due to contribution from early failed CDSP WPs (see Section 7.7.1.1). Because the CDSP WPs produce less decay heat than the CSNF WPs, the relative humidity in the CDSP WP emplacement locations goes above 95 percent and diffusive transport of radionuclide starts before 10,000 years. The peak mean and median annual doses of about 3×10^{-4} mrem and 4×10^{-5} mrem, respectively, occur at approximately 2,000 years; the mean and median annual doses decline thereafter and drop to about 4×10^{-5} mrem and 7×10^{-6} mrem, respectively, at 10,000 years. The abrupt increase in the dose histories at 2,000 years is due to the change in the climate (monsoonal to glacial transition). In the post-10,000-year period, the figure shows a second peak occurring at about 40,000 years with a mean of $\sim 10^{-4}$ mrem and median of $\sim 10^{-5}$ mrem.

The primary radionuclides that contribute to the mean annual dose for the Drip Shield EF Modeling Case are shown on Figure 8.2-4. In the first 2,000 years after repository closure, three soluble and mobile radionuclides dominate dose: ^{99}Tc , ^{129}I , and ^{14}C . From 2,000 years to 10,000 years, the ^{14}C drops in importance and is replaced by ^{239}Pu . At the outset of the post-10,000-year period, the mean annual dose is dominated by ^{239}Pu and ^{99}Tc . The radionuclide ^{239}Pu creates the secondary peak mean annual dose at about 40,000 years, but diminishes in importance by about 200,000 years due to radionuclide decay. From that point forward, ^{242}Pu , ^{237}Np , and ^{226}Ra dominate the mean annual doses.

8.2.2.2 Waste Package Early Failure Modeling Case

From the probabilistic analysis of early failures, the number of WP early failures is estimated to be relatively small. For example, the probability of three or more WP early failures is about 0.12 (Table 6.4-1). As implemented in the TSPA-LA Model, the WP early failure is conceptualized as complete failure with respect to radionuclide containment (Section 6.4.2.2), at the time of repository closure. The overlying DS is not affected by the early failure of the WP. The number of early failed WPs defines the radionuclide source term for this modeling case. The

radionuclide releases from these WP early failures occur by diffusive transport from the waste form with sorbing radionuclides delayed by sorption on corrosion products (Section 6.3.8).

The expected annual dose histories for this modeling case are shown in the multi-realization projections in Figure 8.2-5 for both the (a) 10,000-year period after closure, and (b) post-10,000 years. The expected annual doses account for aleatory uncertainties associated with characteristics of the early failed WPs, such as the number of early failed WPs, type of early failed WPs (CDSP or CSNF), and their locations in the repository. The mean, median, and 5th and 95th percentile curves are shown on Figure 8.2-5 to highlight the uncertainty in the expected annual doses.

For the first 10,000 years after repository closure, there is an initial peak between 1,000 and 2,000 years which is primarily due to contribution from early failed CDSP WPs. Because the CDSP WPs produce less decay heat than the CSNF WPs, the relative humidity in the CDSP WP emplacement locations goes above 95 percent and diffusive transport of radionuclides starts. The abrupt jump in the dose histories at 2,000 years is due to the change in the climate state (monsoonal to glacial transition). Between 9,000 and 10,000 years, the dose histories increase with the peak mean and median annual doses estimated to be about 4×10^{-3} mrem and 2×10^{-4} mrem. This peak is due to the early failed CSNF WPs, which have cooled and have relative humidity environments that promote diffusive transport. For post-10,000 years, the peak mean and median annual doses reach levels of approximately 2×10^{-2} mrem and 4×10^{-3} mrem, respectively, at about 12,000 years and then gradually decline. At about 260,000 years the mean annual dose curve again ascends forming one last broad peak at about 450,000 years; the onset of this increase coincides with the onset of DS failures by general corrosion at about 270,000 years (Figure 8.1-4) and, in turn, seepage directly onto the early failed WPs. The mean and median doses at one million years are about 1×10^{-3} mrem and 2×10^{-4} mrem, respectively.

The major radionuclides that contribute to the mean annual dose for the Waste Package EF Modeling Case are shown on Figure 8.2-6. In the first 10,000 years postclosure, soluble and mobile radionuclides, in particular, ^{99}Tc , ^{14}C , and ^{129}I , dominate the estimate of mean annual dose. In the post-10,000-year period, after ^{99}Tc and ^{129}I decline, the peak mean annual dose is dominated by ^{239}Pu up to about 200,000 years; thereafter, ^{242}Pu , ^{226}Ra , and ^{237}Np are the primary contributors to the peak mean annual dose.

8.2.3 Igneous Scenario Class Modeling Cases

The projections for the Igneous Scenario Class demonstrate postclosure performance for unlikely igneous events and processes that could disrupt the repository system. As noted earlier, the estimated annual frequency of igneous activity at the repository site is 1.7×10^{-8} /yr (Section 6.5.1.1). This frequency is just slightly greater than the NRC frequency of 10^{-8} /yr for very unlikely events and processes, which are excluded from the performance demonstration for the LA by regulation (10 CFR 63.342 [DIRS 178394]). As described in Section 6.5, the Igneous Scenario Class consists of two modeling cases: (1) the Igneous Intrusion Modeling Case that represents the interaction of a hypothetical intrusive magma with the repository and attendant release of radionuclides to the groundwater pathway, and (2) the Volcanic Eruption Modeling Case that represents a hypothetical volcanic eruption at the land surface and the release of radionuclides to the atmospheric pathway.

8.2.3.1 Igneous Intrusion Modeling Case

In this modeling case, a simulated magmatic dike intersects the footprint of the repository causing failure of the EBS. Radionuclide releases that are attributed to the intrusion, and are transported away from the repository, are analyzed in a manner analogous to that for the Nominal Modeling Case, but differ in the conceptualization of the EBS failure. There are two main components to the model: (1) the behavior of the WPs and other EBS elements damaged by an igneous intrusion, and (2) groundwater flow and radionuclide transport away from the WPs. For the purposes of ensuring conservatism in the annual doses to the RMEI, the modeling case assumes that all the WPs in the repository are completely destroyed, exposing the waste forms to percolating groundwater with subsequent degradation, radionuclide mobilization, and transport.

Radionuclide transport occurs through the invert and into the UZ, depending on solubility limits and the rate of water flux through the intruded drifts. It is assumed that the drifts do not act as a capillary barrier, and the seepage water flux into a magma-intruded drift is equal to the percolation flux in the overlying host rock. No barrier performance credit is taken for water diversion by the remnants of the DS or WP, and cladding is assumed to be fully degraded. Because the thermal, chemical, hydrological, and mechanical conditions in the drift following igneous intrusion are uncertain, the EBS is assumed to be completely failed.

The expected annual dose histories for the Igneous Intrusion Modeling Case are shown in Figure 8.2-7 for both the (a) 10,000-year period and (b) post-10,000-year period. The mean, median, and 5th and 95th percentile curves are displayed on a multi-realization plot. The expected annual dose to the RMEI takes into account aleatory uncertainty associated with characteristics of the igneous intrusion, such as the number of future events and the time at which they may occur. These figures indicate that the peak mean annual dose to the RMEI for the 10,000-year time period is less than 0.1 mrem and for the post-10,000-year time period is about 1.3 mrem. The median annual dose at 1,000,000 years is about 0.4 mrem.

The radionuclides that contribute most to the mean annual dose are shown on Figure 8.2-8. Figure 8.2-8a shows that radionuclides ^{99}Tc and ^{129}I dominate the estimate of the mean for the first 4,000 years and ^{239}Pu , ^{99}Tc , and ^{240}Pu dominate the estimate of the mean for the remainder of the 10,000-year postclosure period. Figure 8.2-8b shows that ^{239}Pu , which is transported both in dissolved and colloidal form, dominates the peak mean annual dose for the first 200,000 years and radionuclides ^{242}Pu , ^{237}Np , and ^{226}Ra dominate the estimate of the mean for the remainder of the post-10,000-year time period.

8.2.3.2 Volcanic Eruption Modeling Case

In this modeling case, the disruption is conceptualized as a volcanic eruptive conduit intersecting the repository footprint resulting in the dispersal of waste-contaminated tephra in the atmosphere, with attendant deposition of contaminated ash on the land surface. The performance projections evaluate the post-eruption consequences due to waste redistributed from upstream in the Fortymile Wash watershed and deposited at the RMEI location. WPs in the direct path of the conduit are assumed to be destroyed and entrained into the hypothetical eruption. As described in Section 6.3.11, the radiologic exposure scenario is that the RMEI exposure arises from

contaminated volcanic ash deposited on surface soil and the subsequent radionuclide transport from surface soil to other environmental media (e.g., air and plants).

As noted in Section 6.5.2.1.1, the probabilistic calculations of igneous disruptions indicate that there is about a 70 percent probability that no WPs would be hit by a volcanic conduit intersecting the repository. The small conduit diameters relative to drift spacing implies that 70 percent of the conduits potentially intersecting the repository footprint would intersect between drifts and therefore not impact any WPs. In the 30 percent of the realizations in which one or more packages are intersected, the most likely number hit is four and the maximum number hit is seven.

The expected annual dose histories for this modeling case are shown on Figure 8.2-9 for both the (a) 10,000-year period and (b) post-10,000-year period. The expected annual dose takes into account aleatory uncertainty associated with characteristics of the eruption such as number of WPs intersected by the eruption, the fraction of waste-containing magma ejected in the atmosphere, eruption power, wind direction, and wind speed. The mean, median, and 5th and 95th percentile curves on Figure 8.2-9 show uncertainty in the value of the expected annual dose, taking into account epistemic uncertainty associated with incomplete knowledge of the behavior of the physical system during and after the disruptive event. These figures show that the mean annual dose for 10,000 years postclosure is about 1×10^{-4} mrem and is largely uniform for the post-10,000-year time period with the peak mean less than 2×10^{-4} mrem. The median annual dose is less than 6×10^{-5} mrem at 1,000,000 year.

The radionuclides contributions to the mean annual dose are shown on Figure 8.2-10. Because transport of radionuclides to the location of the RMEI is more rapid in the Volcanic Eruption Modeling Case than in the Igneous Intrusion Modeling Case, radionuclides with short half-lives are able to contribute to the estimate of the mean annual dose estimate. Examples of two short-lived radionuclides are ^{137}Cs and ^{238}Pu , which make significant contributions to the dose at early times, but their contributions drop off rapidly because of radioactive decay. At 300 years, ^{241}Am dominates the total, but its contribution rapidly diminishes after about 1,000 years, also due to radioactive decay. These short-lived radionuclides are able to reach the location of the RMEI before they decay because atmospheric transport to this location is relatively rapid. After 1,000 years, ^{239}Pu and ^{240}Pu are the dominant contributors until approximately 100,000 years; thereafter ^{226}Ra , ^{229}Th , and ^{237}Np become the dominant contributors out to 1,000,000 years.

8.2.4 Seismic Scenario Class Modeling Cases

The probabilistic projections for the Seismic Scenario Class demonstrate postclosure performance for likely and unlikely seismic events. As described in Section 6.6, the Seismic Scenario Class consists of two modeling cases: (1) Seismic GM Modeling Case, and (2) Seismic FD Modeling Case. These modeling cases take into account the aleatory uncertainty of the timing of these events, effects of the seismic events on DSs and WPs, and attendant radionuclide releases. This scenario class also takes into account changes in seepage, WP degradation, and flow in the EBS, as well as the conditions associated with the nominal evolution of the repository system. The likelihood and intensity/magnitude of these seismic events are defined by hazard curves, which were developed as part of the probabilistic seismic hazard analysis (CRWMS M&O 1998 [DIRS 103731]).

8.2.4.1 Seismic Ground Motion Modeling Case

The Seismic GM Modeling Case focuses on postclosure performance as a function of disruptions caused by vibratory ground motion. As described in Section 6.6.1, the likelihood and intensity of ground motion is defined by the mean seismic hazard curve (i.e., relation of peak ground velocities to exceedance frequencies (Figure 6.6-6)). Depending on the timing, sequence, and intensity of the ground motion events, the DSs and WPs can accumulate damage and/or fail. In the case of the DSs, seismic ground motion could cause damage or failure by dynamic loading and/or static loading by rockfall resulting in bucking or rupture (Section 6.6.1.1.2). Similarly, the WPs could be damaged or fail by: (1) local strain that exceeds ultimate tensile strength, (2) deformations creating residual stresses that induce SCC, and (3) lithostatic loading from rubble that causes a puncture of the outer barrier by the WP internals (Section 6.6.1.1.2). Both the DSs and WPs are also degraded by the general corrosion (Section 6.6.2.3). The number of failed CDSP and CSNF WPs determine the source term for the radionuclide transport calculations and projections of expected annual dose to the RMEI.

The expected annual dose histories for the Seismic GM Modeling Case are shown on Figure 8.2-11; multi-realization projections are presented for both the (a) 10,000-year period, and (b) post-10,000-year period. The expected annual dose histories take into account aleatory uncertainty associated with characteristics of future events such as number of events, times of events, and the peak ground velocity of the event. The mean, median, and 5th and 95th percentile curves on Figure 8.2-11 show uncertainty in the value of the expected annual dose, taking into account epistemic uncertainty associated with the incomplete knowledge of the behavior of the physical system during and after the disruptive event. These figures show that the mean annual dose for 10,000 years postclosure is less than 0.2 mrem, while for the post-10,000-year period it is less than 2 mrem. The median annual dose at one million years is less than 0.5 mrem.

The radionuclides that contribute most to the estimate of mean annual dose for this modeling case are presented on Figure 8.2-12. The mean dose curves on Figure 8.2-12a illustrate that three radionuclides, ^{99}Tc , ^{14}C , and ^{129}I , contribute the most to the peak mean annual dose for the 10,000-year time period; during this time period CDSP WPs are the primary containers damaged because the CSNF WPs are much more resistant to seismic damage. The predominant mechanism causing damage to the CDSP WPs and CSNF WPs consisted of small cracks that result in releases from the WPs by diffusion. As can be seen on Figure 8.2-12b, the dominant radionuclides contributing to the peak at approximately 230,000 years are ^{99}Tc , ^{129}I , and ^{79}Se . From 200,000 years to about 800,000 years, the dominant radionuclides are ^{99}Tc , ^{129}I , ^{226}Ra , ^{135}Cs , and ^{242}Pu . After 800,000 years, ^{242}Pu , ^{129}I , and ^{237}Np are the three major contributors to the mean annual dose for this modeling case.

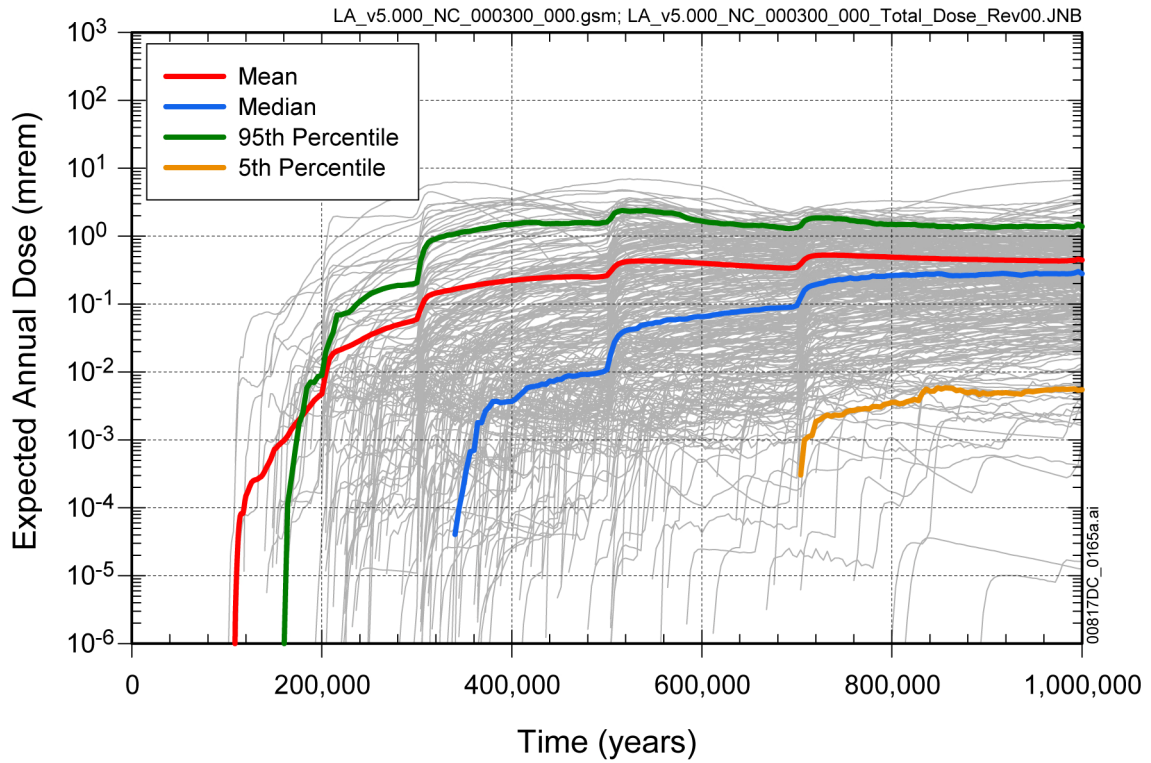
8.2.4.2 Seismic Fault Displacement Modeling Case

The Seismic FD Modeling Case demonstrates postclosure performance as a function of disruptions caused by fault displacement. As described in Section 6.6.1, the characterization of the fault is based on data from known faults in the vicinity of the site area. The disruption in this modeling case is conceptualized as a sudden discontinuity in the profile of the repository. The location and magnitude of the fault displacement determines the number of DSs and WPs that

are disrupted. A mean displacement hazard curve provides the basis for calculating the magnitude of the simulated fault displacement. The WPs are presumed to be sheared if the displacement exceeds a threshold value. As noted in Section 6.6.1.2.3, damage of the DSs and WPs by fault displacement is not expected to occur if the mean annual frequency is greater than $2.5 \times 10^{-7}/\text{yr}$.

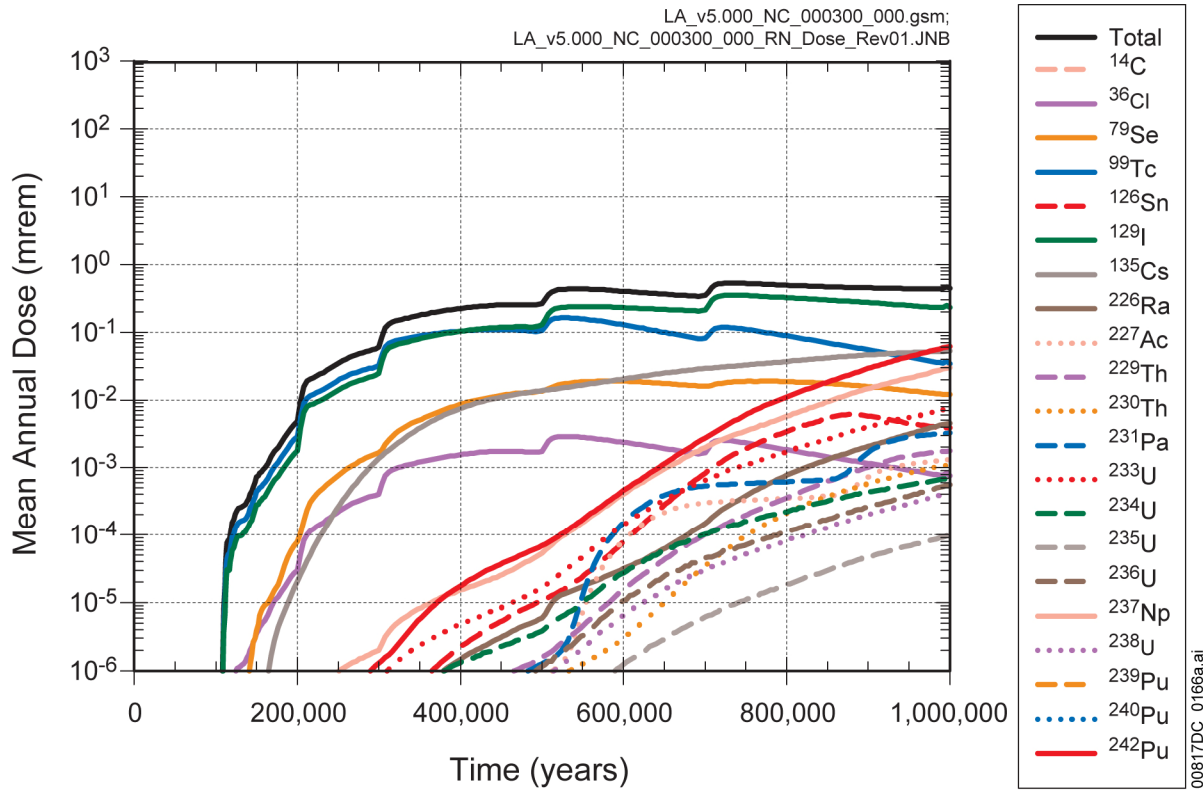
The expected annual dose histories for the Seismic FD Modeling Case are shown on Figure 8.2-13; the figure shows expected annual dose to the RMEI for both the (a) 10,000-year period, and (b) post-10,000-year period. The expected annual dose takes into account aleatory uncertainty associated with characteristics associated with the number of DSs and WPs disrupted. The mean, median, and 5th and 95th percentile curves on Figure 8.2-13 show uncertainty in the value of the expected annual dose, taking into account epistemic uncertainty associated with incomplete knowledge of the behavior of the physical system during and after the disruptive event. These figures show that the mean annual dose for 10,000 years postclosure is less than 2×10^{-3} mrem and for 1,000,000 years postclosure less than 2×10^{-2} mrem. The peak median annual dose for the 1,000,000-year time period is approximately 10^{-2} mrem.

The contribution of individual radionuclides to mean annual dose are shown in the results presented on Figure 8.2-14 for both the 10,000-year and post-10,000-year period after closure. The plot for the 10,000-year period (Figure 8.2-14(a)) shows that ^{99}Tc and ^{129}I dominate the dose for the first 4,000 years after closure, and ^{99}Tc , ^{239}Pu , ^{129}I , and ^{240}Pu are the dominant radionuclides contributing to dose at 10,000 years. As can be noted from Figure 8.2-14b, ^{239}Pu dominates the mean annual doses up to about 200,000 years; thereafter, the radionuclides ^{242}Pu , ^{226}Ra , and ^{237}Np remain dominant contributors for the remainder of the 1,000,000 years.



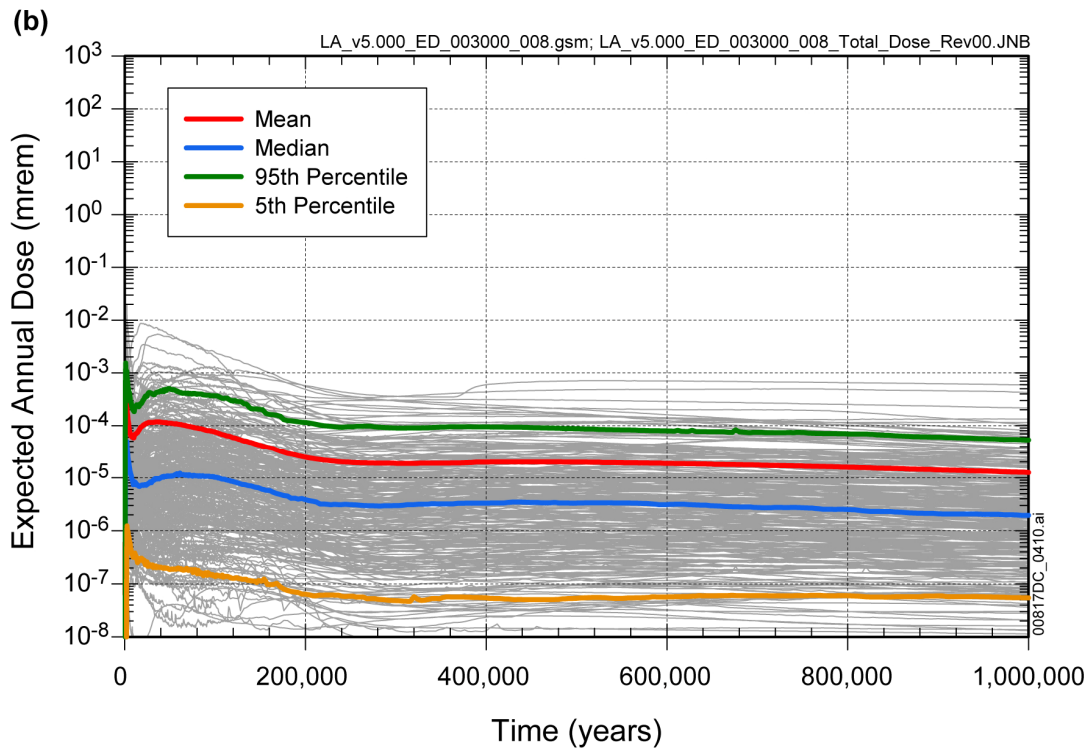
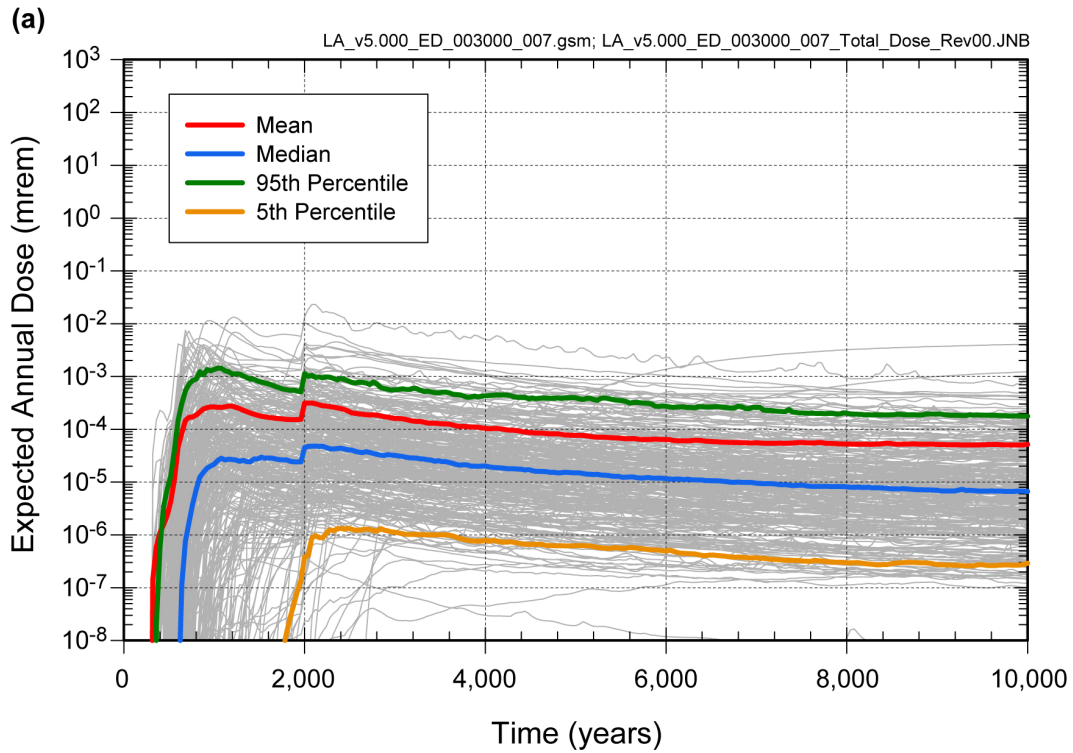
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-1. Probabilistic Projections of Expected Annual Dose for the Nominal Modeling Case for 1,000,000 Years after Repository Closure



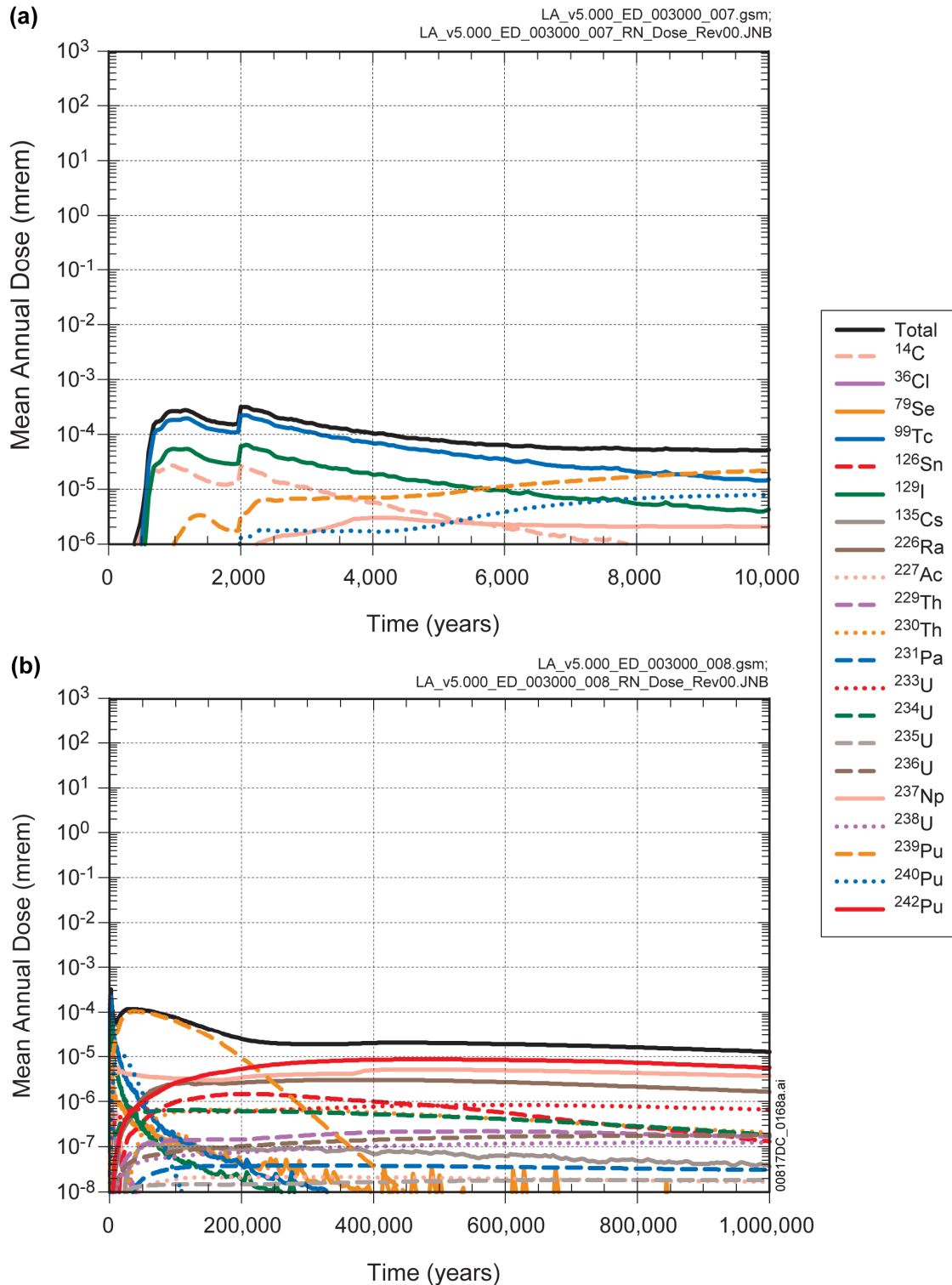
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-2. Contribution of Individual Radionuclides to Mean Annual Dose for the Nominal Modeling Case for 1,000,000 Years after Repository Closure



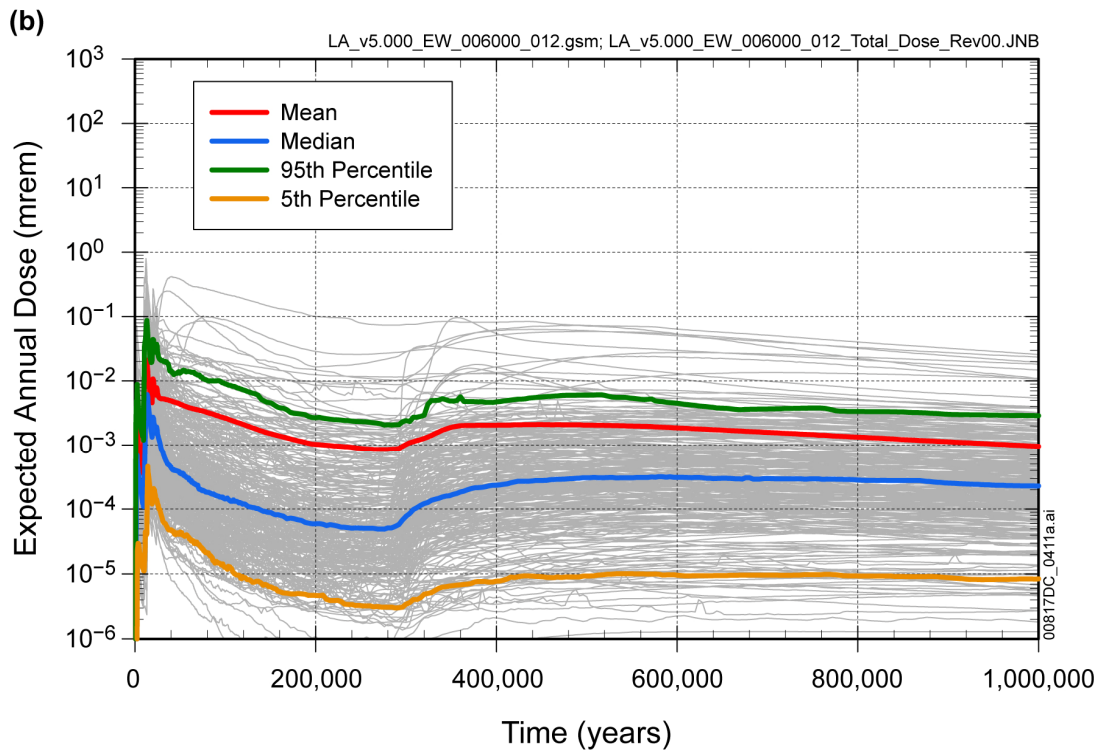
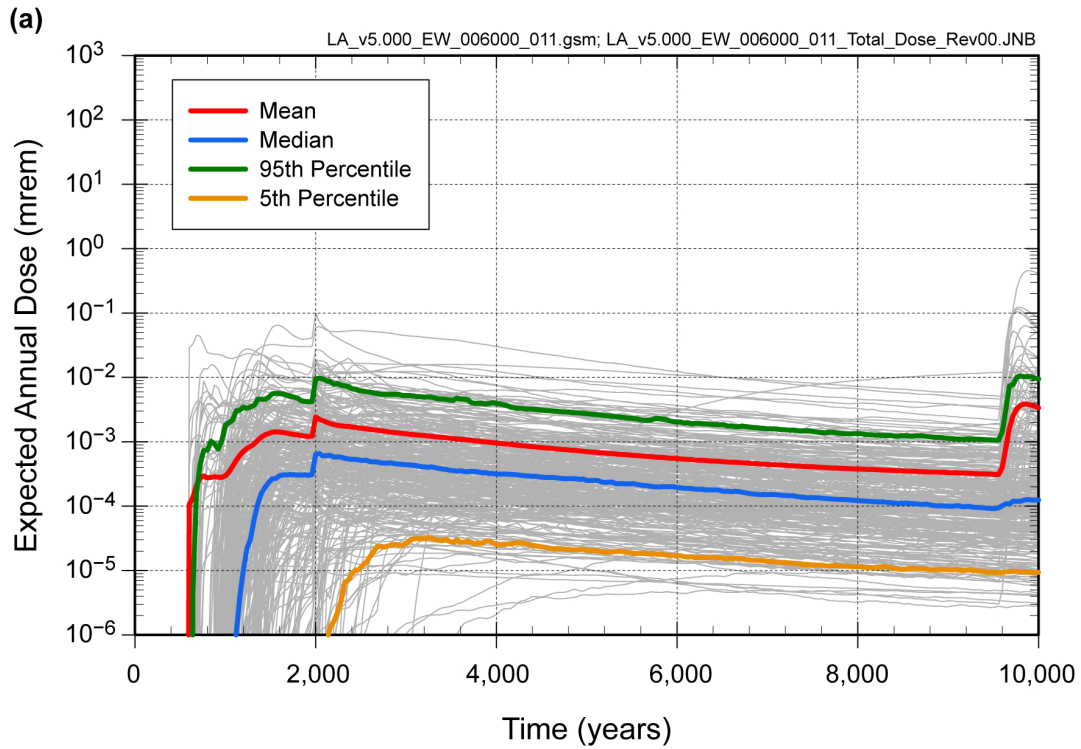
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-3. Probabilistic Projections of Expected Annual Dose for the Drip Shield Early Failure Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



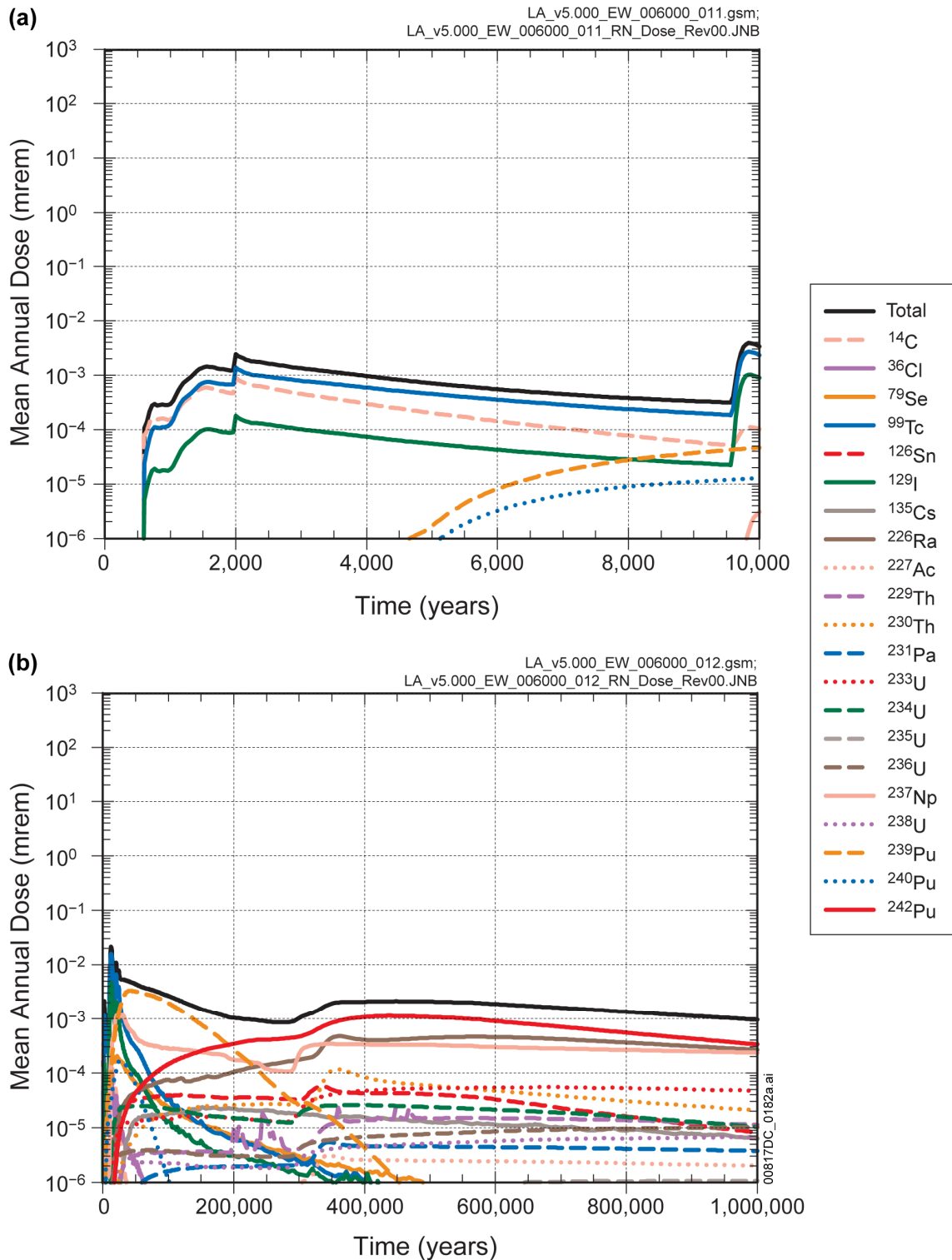
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-4. Contribution of Individual Radionuclides to Mean Annual Dose for Drip Shield Early Failure Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years After Repository Closure



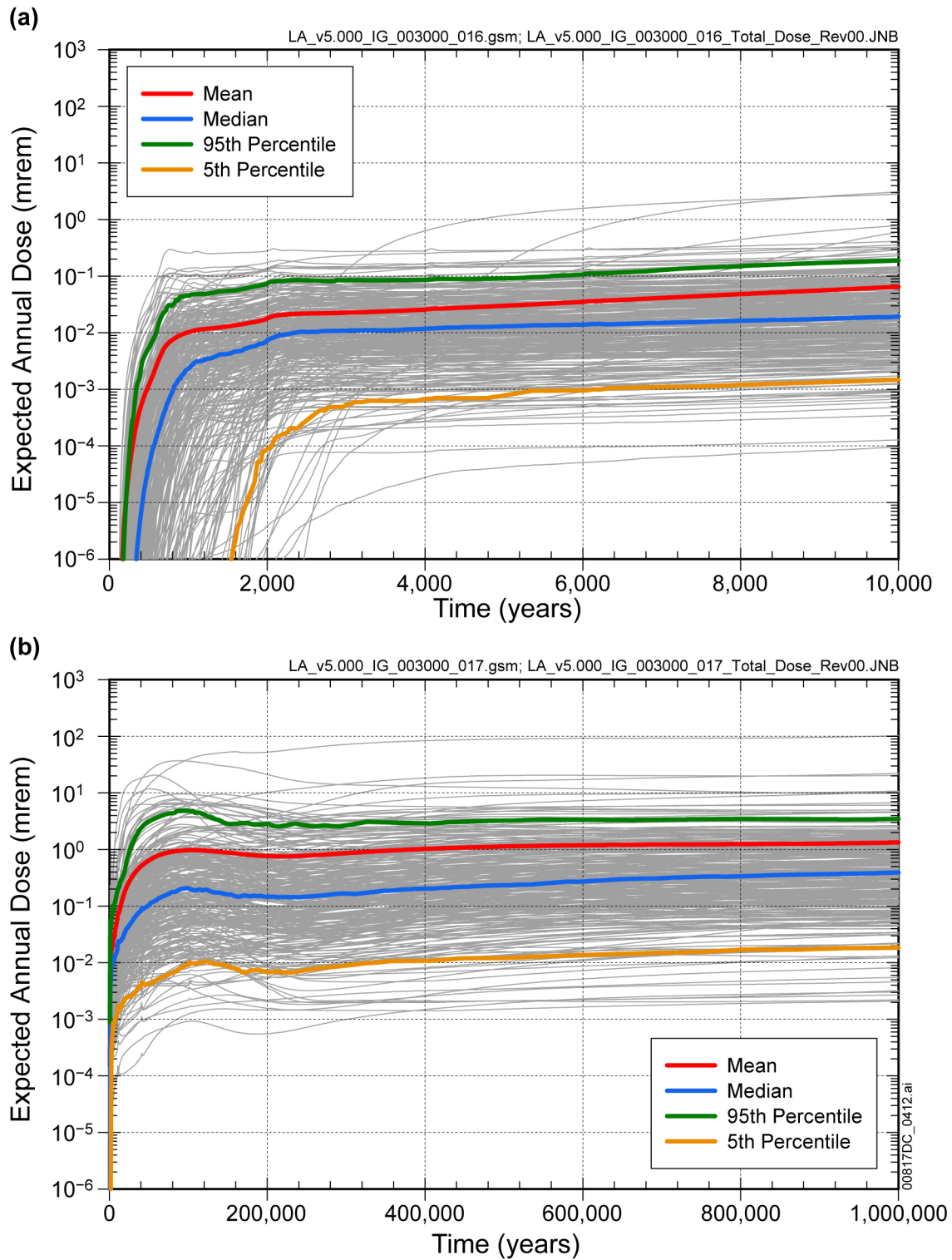
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-5. Probabilistic Projections of Expected Annual Dose for Waste Package Early Failure Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



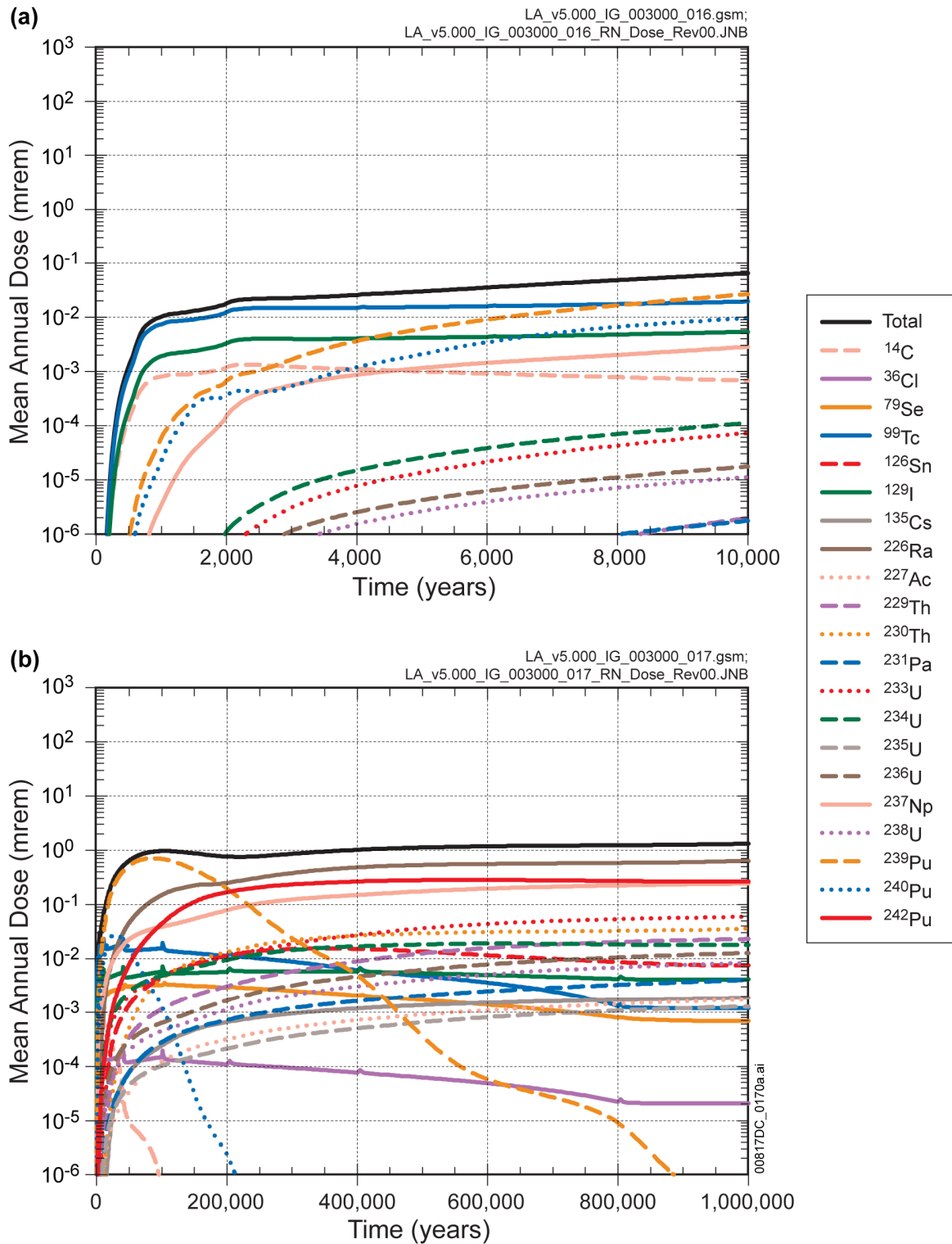
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-6. Contribution of Individual Radionuclides to Mean Annual Dose for Waste Package Early Failure Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



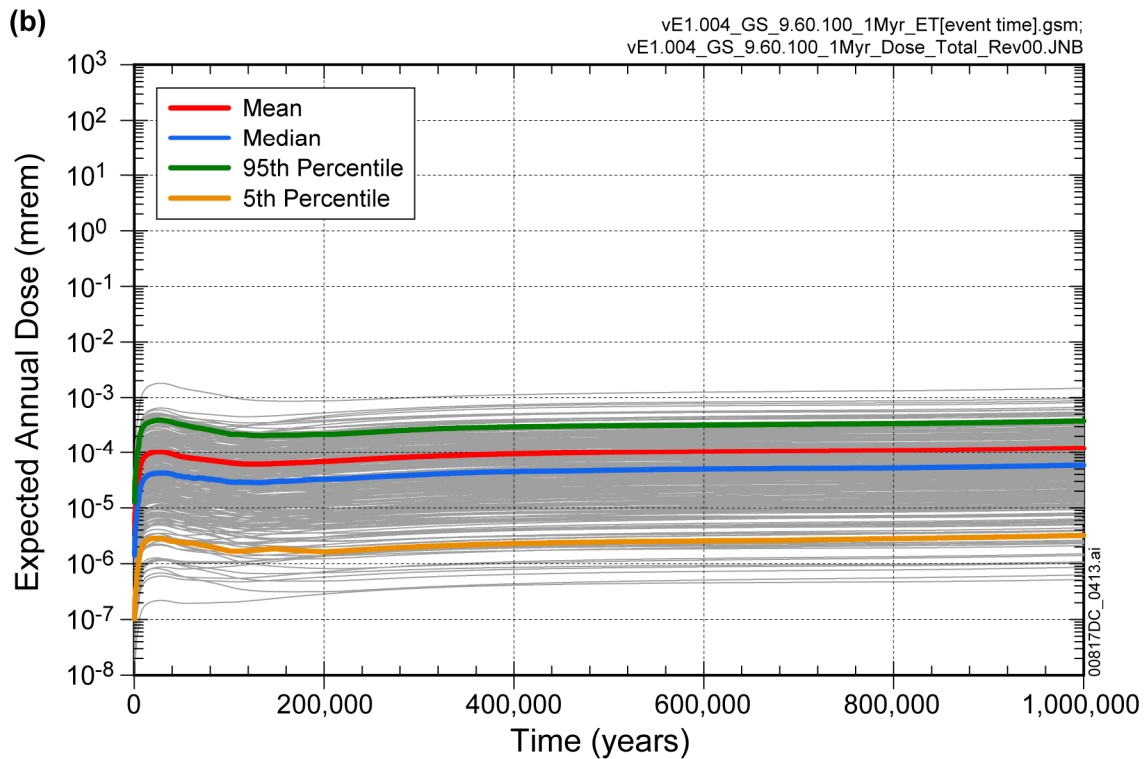
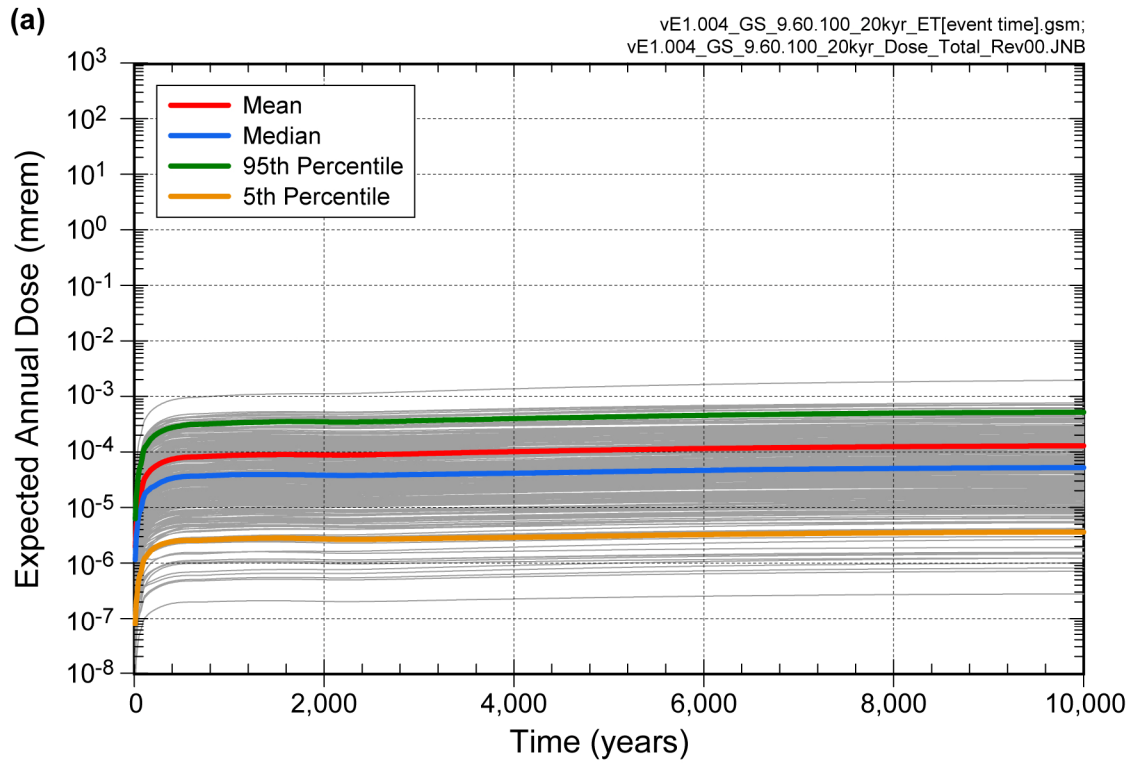
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-7. Probabilistic Projections of Expected Annual Dose for the Igneous Intrusion Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



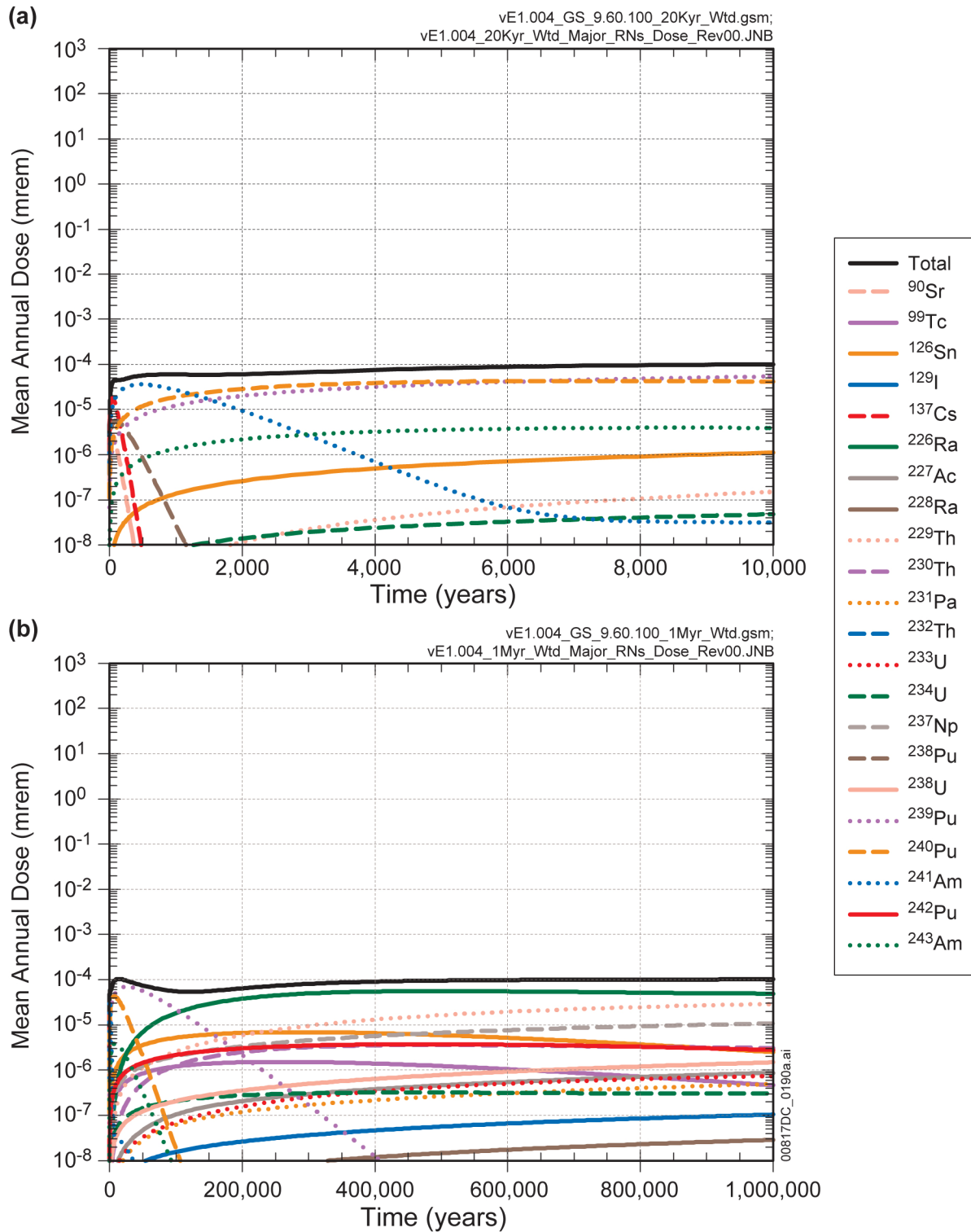
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-8. Contribution of Individual Radionuclides to Mean Annual Dose for the Igneous Intrusion Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years After Repository Closure



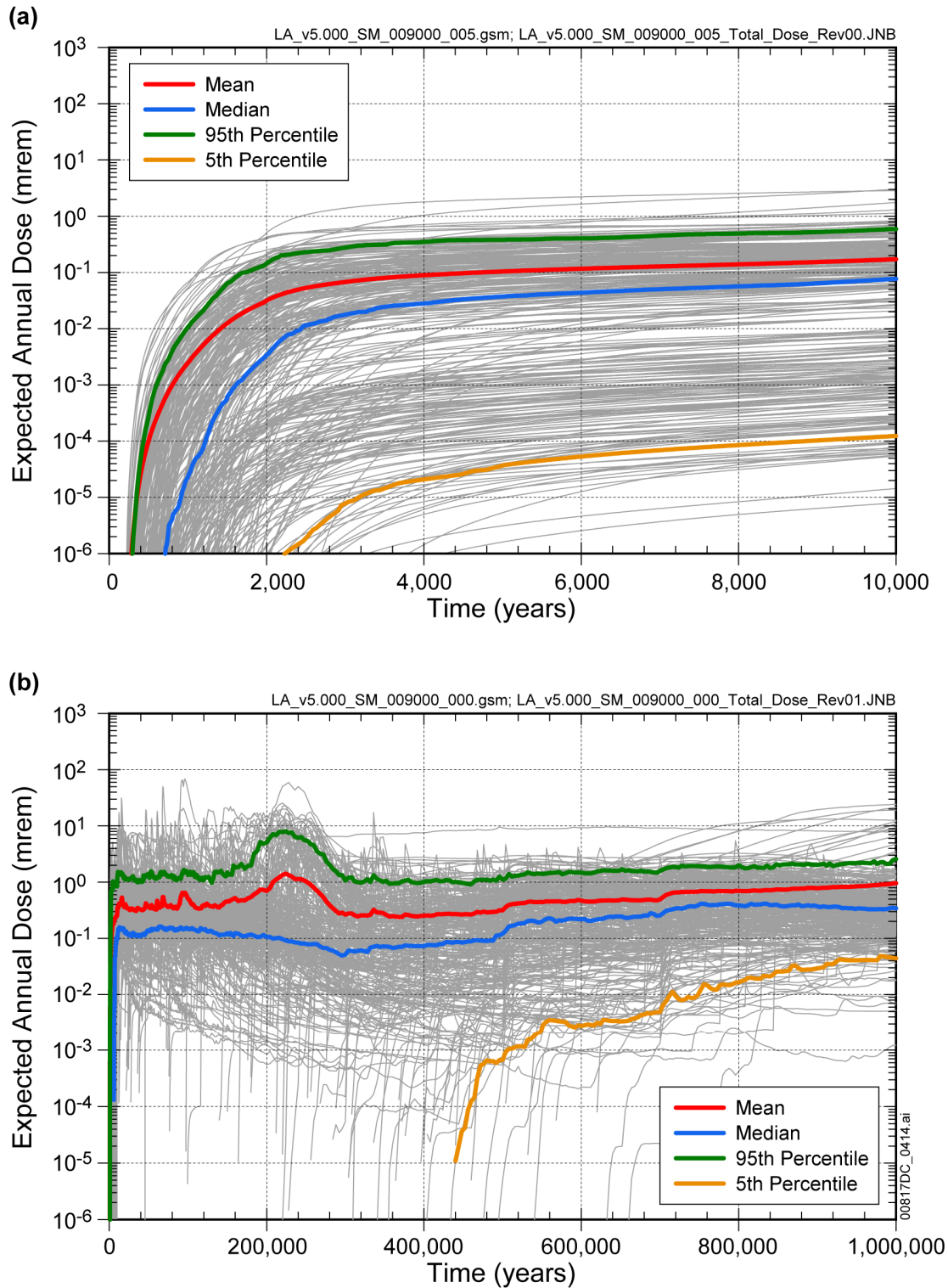
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-9. Probabilistic Projections of Expected Annual Dose for the Volcanic Eruption Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



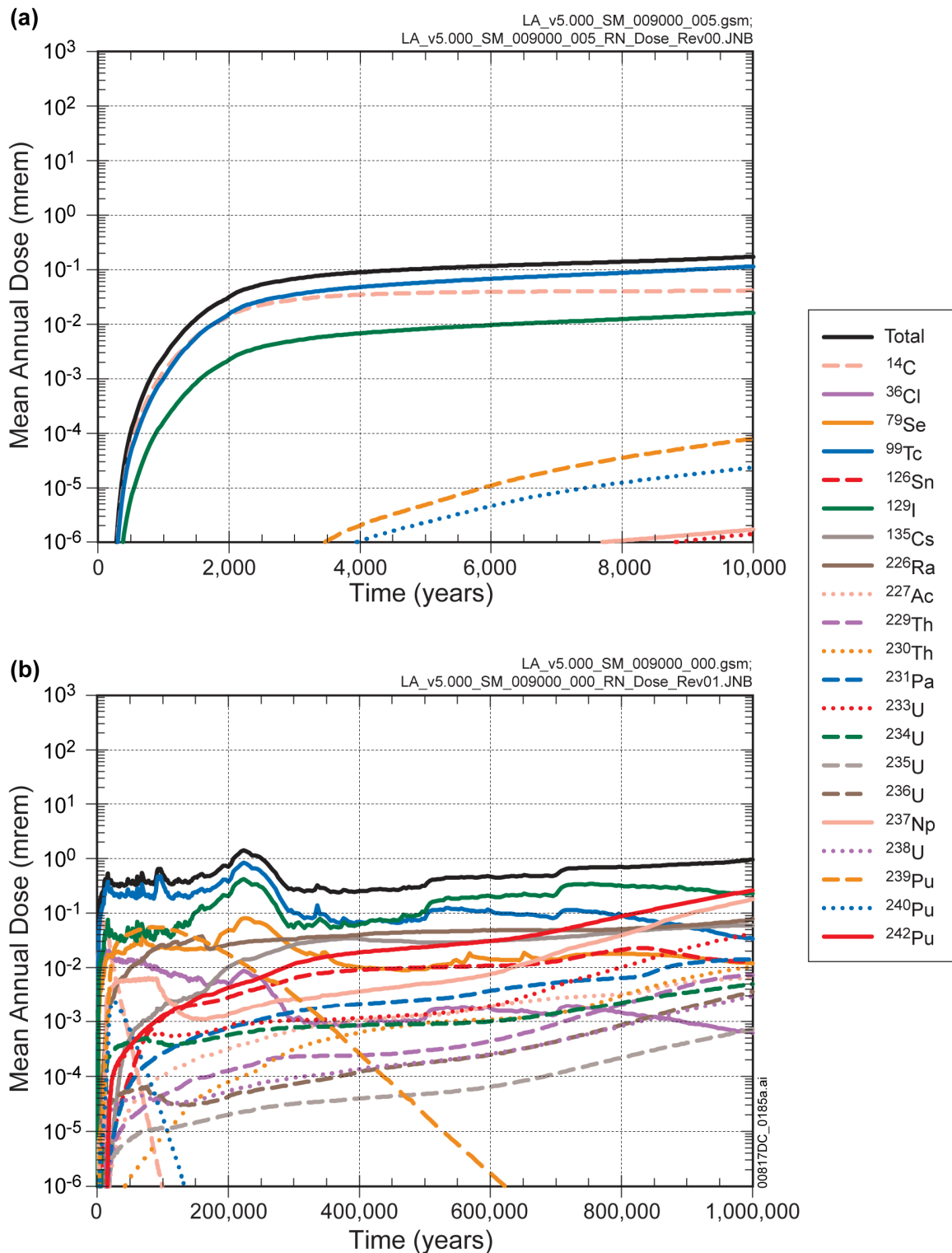
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-10. Contribution of Individual Radionuclides to Mean Annual Dose for the Volcanic Eruption Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



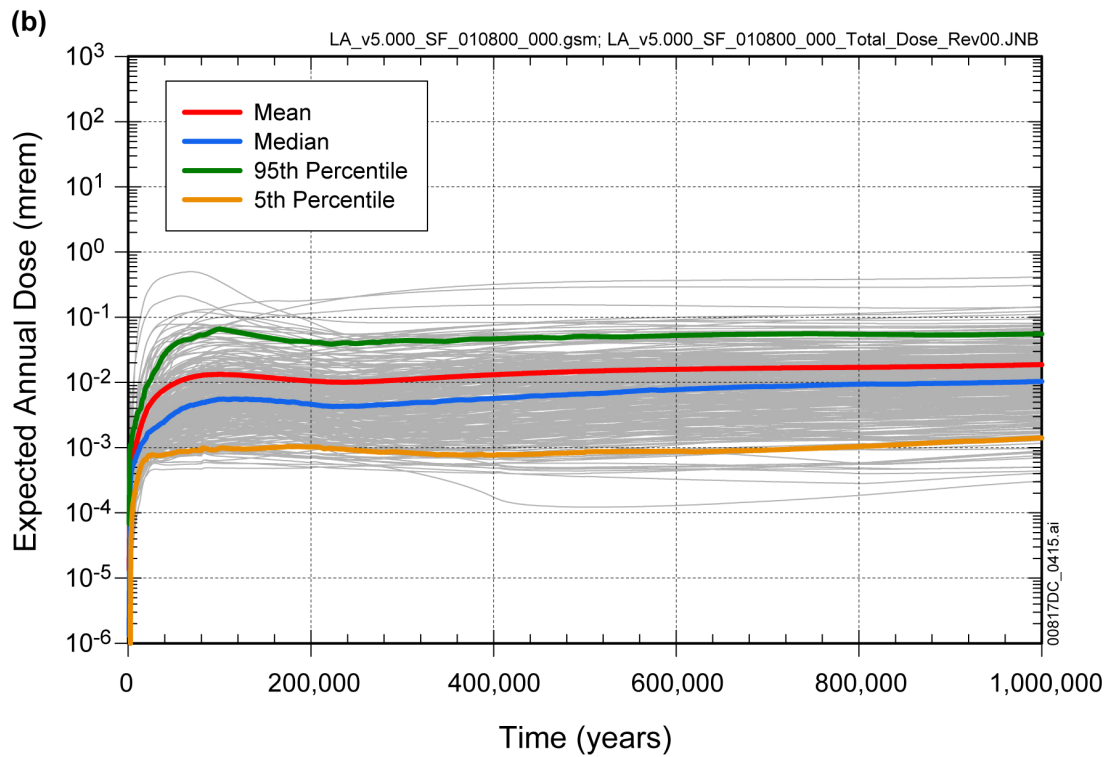
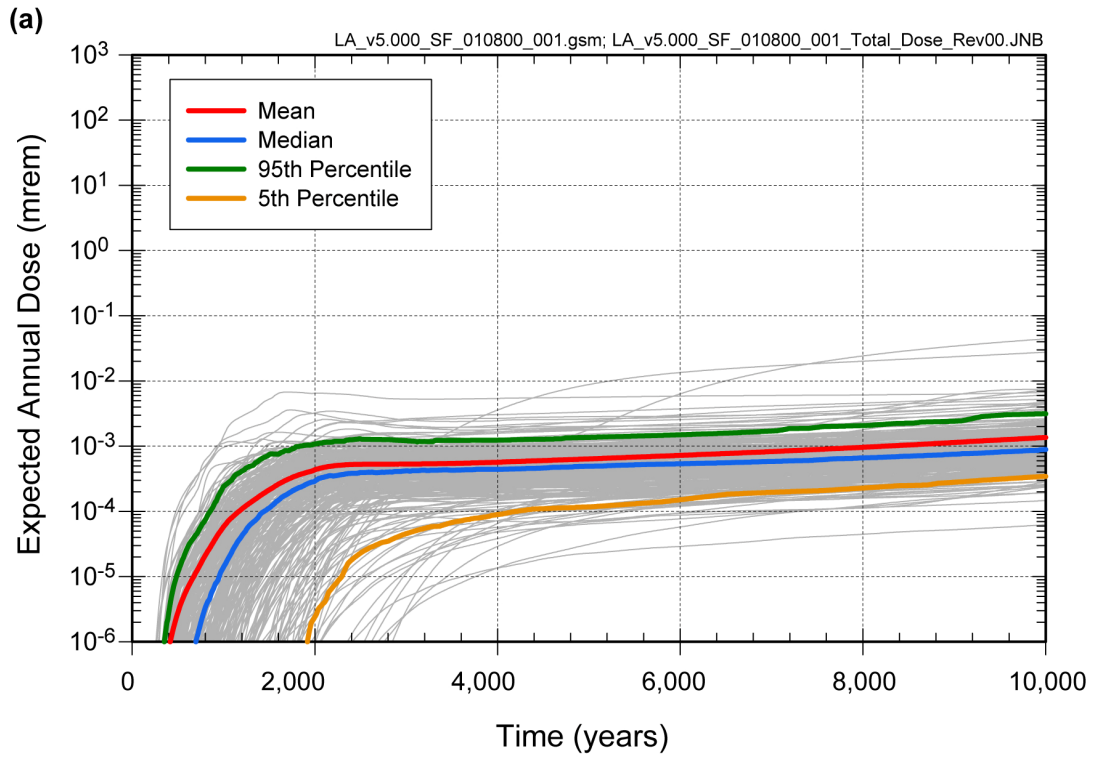
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-11. Probabilistic Projections of Expected Annual Dose for the Seismic Ground Motion Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



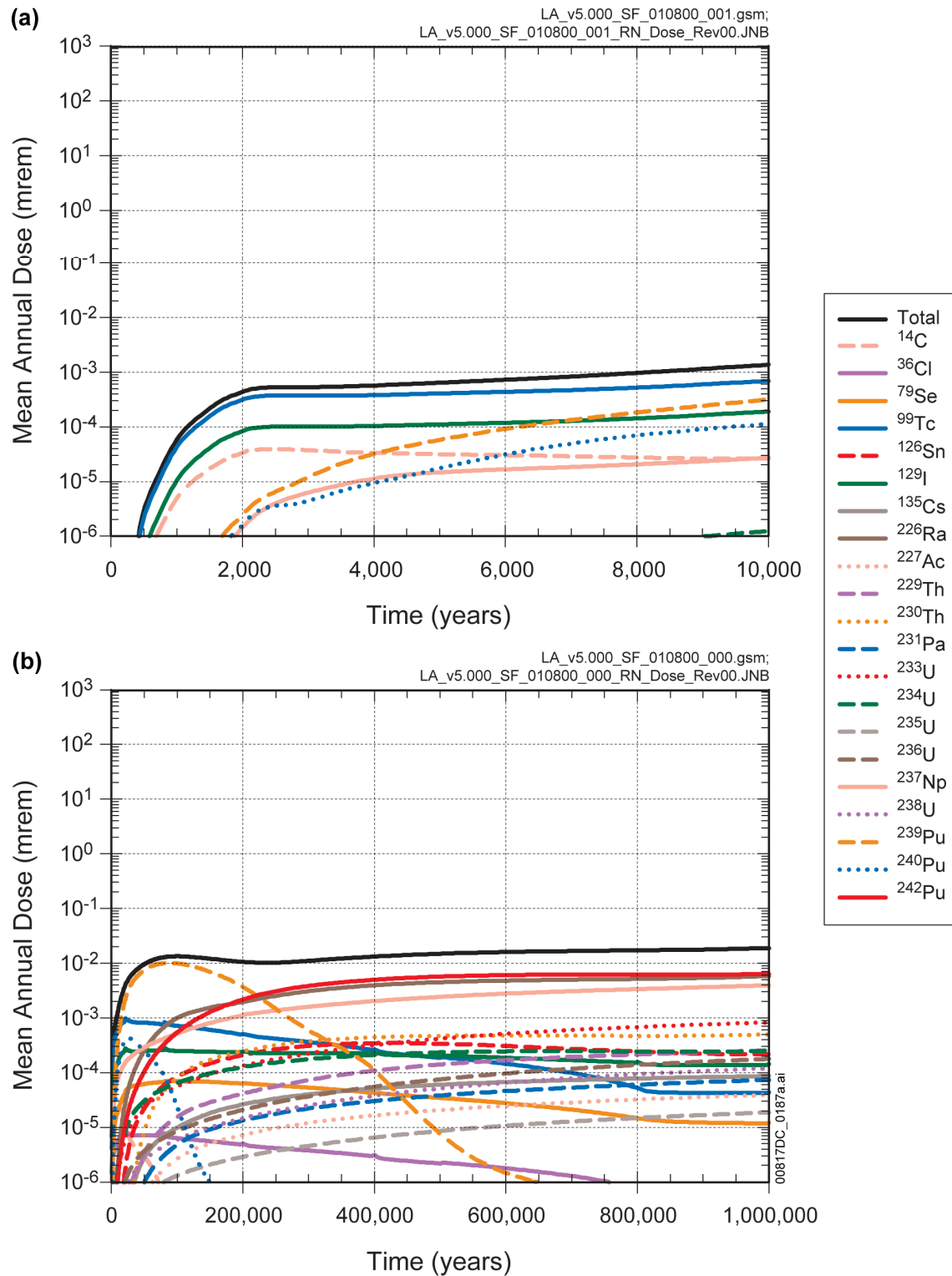
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-12. Contribution of Individual Radionuclides to Mean Annual Dose for the Seismic Ground Motion Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-13. Probabilistic Projections of Expected Annual Dose for the Seismic Fault Displacement Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.2-14. Contribution of Individual Radionuclides to Mean Annual Dose for the Seismic Fault Displacement Modeling Case for (a) 10,000 Years and (b) 1,000,000 Years After Repository Closure

8.3 DESCRIPTION OF MULTIPLE BARRIER CAPABILITY

In the previous sections of this report, the DOE performance demonstrations for the three NRC radiation protections standards: (1) Individual-Protection (10 CFR 63.311) [DIRS 178394], (2) Human-Intrusion (10 CFR 63.321) [DIRS 178394], and (3) Groundwater Protection (10 CFR 63.331) [DIRS 180319], were presented and explained. The performance projections for these three standards demonstrated the isolation capability of the multiple barriers acting as an integrated system. In this section, both qualitative and quantitative descriptions are presented to explain the performance characteristics and capabilities of the individual barriers. In addition to providing valuable insights to barrier capability, this information also addresses the NRC requirements of 10 CFR 63.115 [DIRS 180319] for multiple barriers, which specifies that the DOE demonstration of compliance must:

- “(a) Identify those design features of the engineered barrier system, and natural features of the geologic setting, that are considered barriers important to waste isolation.
- (b) Describe the capability of barriers, identified as important to waste isolation, to isolate waste, taking into account uncertainties in characterizing and modeling the behavior of the barriers.
- (c) Provide the technical basis for the description of the capability of barriers, identified as important to waste isolation, to isolate waste. The technical basis for each barrier's capability shall be based on and consistent with the technical basis for the performance assessments used to demonstrate compliance with 10 CFR 63.113 (b) and (c).”

Identifying the multiple barriers is a relatively straightforward task, and a brief discussion of the Upper Natural Barrier, EBS, and Lower Natural Barrier, as well as their key features, is given in a subsequent section. With regard to providing the technical bases for barrier capability, that information is extensive and consists of site characterization data and numerous modeling studies that have been conducted over the past two decades. The technical bases for the multiple barriers are documented in numerous Yucca Mountain Project (YMP) reports, many of which are cited in Sections 1 through 6 of this report.

The main purpose of this section is to describe and explain the performance capability of the individual barriers and their barrier features. For this reason, this section is organized into four general parts:

1. Description of the major radionuclides selected to illustrate barriers capability
2. Identification of multiple barriers and their primary barrier features
3. Description of barrier capability in the absence of disruptive events, as defined by a composite modeling case of the Nominal Modeling Case plus Early Failure Modeling Cases
4. Description of barrier capability in the presence of disruptive events, as defined by Seismic GM Modeling Case.

With regard to item 4, the Seismic GM Modeling case was selected to illustrate barrier capability because: (1) frequencies of occurrence of seismic events are much greater than those for other disruptive events and (2) seismic ground motion has been shown to be important to postclosure performance (Section 8.1.1.2).

Selected performance projections of barrier capability will be presented, which highlight each barrier's inherent capability to isolate (i.e., contain and/or confine) the radionuclides that may pose the most risk over the range of plausible future system states. These projections also provide an additional technical basis for the performance demonstration results presented in Sections 8.1 and 8.2. Because of current programmatic and computational constraints, however, the TSPA-LA results for parts (3) and (4) will be documented in a separate addendum to this report.

8.3.1 Radionuclides Selected to Demonstrate Multiple Barrier Capability

One of the basic and important functions of the multiple barriers is to prevent or substantially reduce the rate of radionuclide movement from the repository to the accessible environment. To demonstrate the range of barrier capability but yet limit the number of calculations, a small subset of radionuclides was selected for the barrier performance demonstrations. That subset was selected from two lists: (1) radionuclides identified as dominating the mean annual doses to the RMEI for the 10,000-year and post-10,000-year time periods, and (2) radionuclides identified as dominating the curie inventory for the 10,000-year and post-10,000-year time periods. The first list of important radionuclides was developed directly from the probabilistic projections of mean annual dose (Section 8.1, Figure 8.1-7). The second list was compiled by examining the inventory decay histories, which is discussed herein.

Calculations of inventory decay histories for the major radionuclides are shown on Figure 8.3-1 for the two compliance periods of 10,000 years and post-10,000 years (i.e., after 10,000 years but within 1,000,000 years). The time dependent behavior of individual radionuclides is the result of simple radioactive decay and, in some cases, decay chain in-growth. From the curves on these plots, one can note that in the first 100 years, two fission products, ^{90}Sr (half-life of 28.8 years) and ^{137}Cs (half-life of 30 years), dominate the inventory; thereafter, the actinide radionuclide ^{241}Am (half-life of 432.7 years) dominates to about 1,000 years, ^{240}Pu (half-life of 6,560 years) dominates to about 7,000 years, and then ^{239}Pu (half-life of 2.41×10^4 years) dominates to about 115,000 years. Dominance then shifts to the fission product ^{99}Tc (half-life of 2.13×10^5 years) for the majority of time in the 1,000,000-year time period.

More refined insights to inventory dominance can be gained by examining plots of the fraction of total activity, at time 't', for each radionuclide; two plots are shown on Figure 8.3-2 for 10,000 years and 1,000,000 years. From these plots, it is evident that the dominant radionuclides in the curie inventory, grouped by compliance period, are:

1. For 10,000 years: ^{137}Cs and ^{90}Sr , ^{241}Am , ^{240}Pu and ^{239}Pu
2. For post-10,000 years: ^{239}Pu , ^{99}Tc , ^{237}Np , ^{233}U , and ^{229}Th .

At closure, the three radionuclides, ^{137}Cs , ^{90}Sr , and ^{241}Am , collectively represent about 85 percent of the total curie inventory with ^{137}Cs representing about 46 percent, ^{90}Sr about 29

percent, and ^{241}Am about 10 percent. From Figure 8.3-2(a), it can be clearly seen that ^{241}Am dominates the curie inventory in the interval from 100 to 1,000 years. Dominance shifts to ^{240}Pu and ^{239}Pu from 1,000 years to 10,000 years with these two actinides alone representing more than 90 percent of the inventory at the end of 10,000 years. As shown on Figure 8.3-2 (b), ^{239}Pu dominance peaks at about 50,000 years. From roughly 100,000 years to 1,000,000 years, the inventory dominance shifts to ^{99}Tc (half-life of 2.13×10^5 years) until roughly 850,000 years then transitions to dominance by three actinides: ^{237}Np (half-life of 2.14×10^6 years), ^{233}U (half-life of 1.59×10^5 years), and ^{229}Th (half-life of 7.3×10^3 years); the curves for ^{233}U and ^{229}Th overlay due to secular equilibrium. The latter three actinides are members of a decay chain in the neptunium series (Figure 6.3.7-4). A summary of the decay history of the total curie inventory and the major contributors is presented in Table 8.3-1; the percentages shown in the table were calculated directly from the data files used to create Figure 8.3-1.

Comparing the above radionuclides that dominate the inventory with the radionuclides identified as important to mean annual doses to the RMEI, namely, for the 10,000-year compliance period: ^{99}Tc , ^{14}C , ^{129}I , ^{239}Pu , ^{36}Cl , ^{79}Se , and ^{240}Pu , and for the post-10,000-year compliance period: ^{226}Ra , ^{242}Pu , ^{237}Np , ^{129}I , ^{233}U , ^{135}Cs , ^{230}Th , ^{99}Tc , ^{229}Th , and ^{231}Pa ; it can be noted that the radionuclides that dominate the inventory also appear in the list of radionuclides important to dose, with the exception of the short-lived ^{137}Cs , ^{90}Sr , and ^{241}Am .

Based on the above considerations, the following subset of radionuclides was selected for use in describing barrier capabilities with regards to reducing or substantially reducing the rate of radionuclide movement:

These radionuclides represent a broad range of nuclear properties, geochemical behavior, and transport characteristics including:

- Large initial inventory and short half-life: ^{137}Cs , ^{90}Sr , ^{241}Am , ^{240}Pu
- Highly soluble, non-sorbing, long half-life and major contributor to dose: ^{99}Tc
- Solubility limited, strongly sorbed, long half-life, transported in dissolved and colloidal phases, and important contributor to dose: ^{239}Pu and ^{242}Pu
- Moderately soluble, low sorbing, very long half-life, and transported in dissolved phase: ^{237}Np and ^{234}U
- Low initial inventory, strongly sorbed, and important contributor to colloids or decay chain in-growth: ^{243}Am , ^{230}Th , and ^{226}Ra .

The primary barrier performance metrics that will be used are: seepage flux, DS and WP lifetimes and radionuclide release rates. The release rates are computed at the following barrier interfaces: (1) EBS outer boundary, (2) UZ and SZ interface, and (3) SZ and accessible environment interface.

8.3.2 Identification of Barriers for Yucca Mountain Repository System

As noted earlier, the Yucca Mountain Repository system is comprised of three barriers, namely, the Upper Natural Barrier, the EBS, and the Lower Natural Barrier (Figure 8-1). Collectively, these three barriers function to: (1) prevent or substantially reduce the rate of movement of water or radionuclides from the repository to the accessible environment, or (2) prevent or substantially reduce the release rate of radionuclides from the repository. A brief description of these barriers and their features is given below:

1. Upper Natural Barrier—Barrier features include the topography and surface soils of the mountain, the unsaturated tuff units above the repository, and the rock in which the repository is constructed.
2. EBS—Barrier features include the emplacement drifts, DSs, WPs, waste forms, cladding (associated with CSNF, DSNF, and NSNF), WP pallets, and ballast in the emplacement drift inverts.
3. Lower Natural Barrier—Barrier features include the volcanic rock in the UZ beneath the repository and the volcanic rock and alluvial material in the SZ between the repository and the accessible environment.

It is important to clarify that the Upper Natural Barrier is the portion of the geologic strata that extends from land surface to the bottom of the repository emplacement horizon. The Lower Natural Barrier extends from the base of the repository horizon to the water table and includes the SZ below the water table that extends from the repository footprint to the accessible environment boundary at approximately 18 km.

For the Upper Natural Barrier, the capability of the barrier features is described with respect to how they prevent or substantially reduce the rate and amount of water that may seep into the repository drifts and, ultimately, to the accessible environment. In contrast, the capability of the EBS features is described with respect to how they prevent or substantially reduce the release rate of radionuclides from the WPs. In the case of the Lower Natural Barrier, the capability of the barrier features is described in terms of how they prevent or substantially reduce the rate of movement of radionuclides from the repository to the accessible environment.

8.3.2.1 Upper Natural Barrier

The Upper Natural Barrier serves to prevent or substantially reduce the rate of water percolating downward through the UZ. More specifically, this barrier is important to postclosure performance because percolating water and the ensuing drift seepage are the only means by which radionuclides could be mobilized from the nuclear waste and transported by groundwater to the accessible environment. This barrier is the thick UZ formation of Yucca Mountain, which is composed of surficial soils and the unsaturated tuffs; namely, the Tiva Canyon welded (TCw), Paintbrush nonwelded (PTn), and Topopah Springs welded (TSw) units.

The location and elevation of the repository provide further advantage to the favorable and unique barrier characteristics of the geologic and hydrogeologic setting of the Yucca Mountain site. These favorable barrier characteristics include:

- A semiarid climate with limited precipitation and significant evapotranspiration
- A thickness of rock and soil above the repository of at least 200 m and up to more than 400 m
- Geologic, geochemical, and geomechanical characteristics that are compatible with the design and construction of an effective EBS
- Geomechanical and thermal characteristics that provide a stable facility with adequate capacity for waste disposal.

The semiarid climate in the Yucca Mountain is characterized by mean annual precipitation rates of 100 to 300 mm/yr (BSC 2004 [DIRS 169734], Figure 7-6).

Topography and Surficial Soils—The topography and surficial soils play a significant role as barriers to water because they limit the net infiltration into the UZ. The surficial soils function as a barrier by diverting (as runoff) some of the water that arrives as precipitation and run-on, and by storage in the soil, some of which is returned to the atmosphere as evapotranspiration. Thus, the volume of water that is diverted or evaporated would not percolate into the UZ rock layers as net infiltration. Recent studies of infiltration (SNL 2007 [DIRS 182145]) for the Yucca Mountain site indicate that the surficial soils are a very effective natural barrier to surface infiltration. For example, detailed water balance calculations for present-day climate shows that the mean net infiltration rate is less than 10 percent of the mean precipitation rate (SNL 2007 [DIRS 182145], Table 6.5.7.4-1); these water balance calculations also indicate that the processes of evapotranspiration and runoff reduce the net infiltration rate by as much as 88 percent.

Unsaturated Tuff Units—Water flow in the fractured welded tuffs that host the repository (i.e., TSw hydrogeologic unit) occurs primarily in widely distributed fractures (e.g., fracture density, spacing, and apertures). In contrast, the PTn hydrogeologic unit (located above the TSw hydrogeologic unit) is characterized by matrix-dominated flow. The matrix-dominated flow in the PTn unit attenuates, or dampens, the amplitude of pulses of percolation caused by variable infiltration and lateral heterogeneity.

As part of the UZ Flow Model development over the past decade, the steady-state nature of the flow fields and the damping of transient pulses were evaluated in different studies. The work of Wang and Narasimhan (1985 [DIRS 108835]; 1993 [DIRS 106793], Figure 7.4.7), for example, indicates that effects of infiltration pulses at the surface are damped by the underlying tuff units, especially the PTn. The welded tuff of the repository horizon exhibited only small changes in saturations, pressures, and potentials from steady-state values in response to the transient pulses. Pan et al. (1997 [DIRS 164181]) investigated transient flow behavior for downward water flow through sloping layers in the UZ.

Wu et al. (2002 [DIRS 161058], pp. 35-1 to 35-12) analyzed the capillary barrier capacities in unsaturated units and indicated that, on average, it took several thousand years for water to travel through the PTn. Wu et al. (2000 [DIRS 154918]; 2002 [DIRS 161058]) analyzed the implications of capillary barrier development in subunits of the PTn for lateral diversion of flow in the PTn. Along sloping layers, strong capillary barriers, if formed, will promote lateral diversions. A more recent study, conducted by Zhang et al. (2006 [DIRS 180273]) using three-dimensional and one-dimensional model results, shows that the PTn can attenuate episodic infiltration pulses significantly, most percolating water is damped by the subunits at the top of the PTn, and a small percentage of that water flux is diverted into faults.

Within the repository horizon, ambient unsaturated flow and thermo-hydrologic processes are favorable to the natural barrier function of preventing or substantially reducing the movement of water into emplacement drifts. Thermo-hydrologic (TH) analyses (SNL 2007 [DIRS 181383]) of the postclosure thermal period, for example, show that the decay heat will raise host rock temperatures and create dry-out zones around the emplacement drifts (Figure 6.3.3-3). These dry-out zones would vaporize locally percolating water and induce water (liquid and vapor) flow away from the drift. Current analyses of TH processes indicate that dry-out zones or vaporization barriers may effectively extend upwards of about 11 meters (Section 6.3.2.1) into the host rock and would persist for several centuries (SNL 2007 [DIRS 181383], Table 6.3-50[a]). As the host rock cools to temperature below boiling, the locally percolating water would create a zone of increased saturation at the drift crown and around the drift opening, resulting in the formation of a capillary-barrier (SNL 2007 [DIRS 181244], Section 6.1.4). The extent of the capillary-barrier effect would be limited to a relatively thin layer around the opening but would be sufficiently effective to divert flow around the drift (Figure 6.3.3-4); this diversion of ambient water flow, in turn, would reduce the seepage flux to drift.

8.3.2.2 Engineered Barrier System

The primary features of the EBS include the emplacement drifts, DSs, WPs, cladding, waste forms, WP pallets, and the drift invert. While the WPs remain without a major breach (e.g., an opening or tear/puncture created by general corrosion or seismic loading), none of the waste can be exposed to water, and no release of radionuclides can occur. The WP outer barrier will be made of Alloy 22, a corrosion-resistant nickel alloy. Even if some WPs were to be breached, the intact DSs would prevent seepage from contacting the waste. Similarly, the Titanium Grade 7 DS is sufficiently resistant to both general corrosion and localized corrosion and should not be breached by either of these nominal processes for more than 10,000 years. The barrier capability of the EBS features, however, can be diminished by potential disruptive events and processes, such as seismic ground motion, fault displacement, and igneous intrusion.

The favorable barrier characteristics of the EBS include:

- A stable thermal, mechanical hydrologic, and chemical environment affected principally by the thermal effects of radioactive decay
- Corrosion resistant metals that are designed to perform and function in the thermal, mechanical, hydrologic, and chemical environments expected in the emplacement drifts

- DS, WP, and cladding materials with designs and fabrication methods that reduce the potential effects of SCC, and other physical-chemical degradation processes
- Generally low radionuclide solubility and high sorption characteristics of radionuclides (on corrosion products) delaying or preventing their release even in rare early breach of WPs
- Delayed transport of radionuclides through the EBS due to low rates of water advection through the EBS features, and the slow diffusion of radionuclides through any continuous water film that is expected to be highly tortuous.

Evidence from natural and man-made openings in unsaturated underground environments indicates that such openings effectively limit the movement of water and often create stable conditions in which fragile materials may be preserved for tens of thousands of years. Analogue data from Pena Blanca indicate that underground openings provide a significant reduction in seepage compared to the amount of water infiltration that enters the UZ (SNL 2007 [DIRS 181244], Section 7.3[a] and 8.1[a]). Emplacement drifts provide the thermal, mechanical, hydrologic, and chemical environment in which the rest of the EBS features function. These environments are affected by the heat caused by the decay of radioactive materials in the waste, in particular, in the CSNF waste form. Although these environments are expected to change with time, in the absence of unlikely disruptive events such as low probability seismic or igneous events, the rate of change is very slow.

Drip Shields—The corrosion resistant Titanium Grade 7 DSs are installed over the WPs prior to repository closure. The DSs divert any potential seepage from the Upper Natural Barrier and water condensate (that may form from the water vapor in the air) around the WPs to the emplacement drift invert. In addition, the DS protects the WP from the potential effects of rockfall.

Waste Packages—As long as they remain without a major breach, WPs prevent contact between water and the waste form and prevent the release of radionuclides. Should water contact the WPs, any resultant corrosion of the Alloy 22 WP is expected to proceed slowly under the nominal corrosion degradation conditions and result in initial breaches in the form of hairline cracks that still limit the movement of water that could potentially contact the waste form, and thus, reduce the release rate of radionuclides from the WPs. Note that localized corrosion of WPs is not expected to occur under the repository condition (Section 6.3.5.2.3). The 50-mm thick stainless steel WP inner vessel and the 25-mm thick transport, aging, and disposal (TAD) canister are expected to provide some performance benefit for waste containment, and could potentially contribute to the reduction of the rate of radionuclide release after WPs are breached. The performance benefit of these specific WP components, much like CSNF cladding, are not considered in the TSPA-LA Model.

Cladding—Zircaloy cladding is a highly corrosion-resistant component of the CSNF that prevents or substantially reduces the contact of water with the waste form and reduces the release rate of radionuclides. For the purposes of conservatism, no performance credit is taken in the TSPA-LA Model for CSNF cladding effects on release rate.

Waste Forms—The waste forms that will be disposed of, include spent nuclear fuel (SNF) and HLW glass (vitrified). These waste forms are solid materials that generally degrade slowly in the unsaturated environment of the repository, thus reducing the release rate of radionuclides, although the DSNF is assumed to degrade instantaneously in the TSPA-LA Model.

Pallets—The WP pallet provides stability of the WP in the case of a seismic event and keeps the WP elevated above the floor of the emplacement drift as well. Thus, direct contact between the WP and the invert is not possible until the pallet degrades. This behavior is not considered in the TSPA-LA Model.

Inverts—The emplacement drift invert is composed of two parts: a steel invert structure and ballast (or fill). In the unsaturated repository environment, the granular materials in the invert slow the diffusive movement of radionuclides into the Lower Natural Barrier.

The features described above give the EBS the ability to prevent or significantly reduce the seepage water from contacting the waste forms, thereby substantially reducing the potential radionuclide release and rate of release from the waste into the Lower Natural Barrier.

8.3.2.3 Lower Natural Barrier

The features of the Lower Natural Barrier include the UZ tuff layers immediately below the repository and the SZ that extends downgradient from the repository to the accessible environment. The UZ includes a portion of the Topopah Springs welded (TSw) and Calico Hills nonwelded (CHn), and Crater Flat undifferentiated (CFu) units (Ortiz et al. 1985 [DIRS 101280]). These unsaturated rock layers represent a total vertical transport path of 250 to 380 m across the repository footprint with an average of about 300 m to the water table for present-day climate conditions. Under the wetter climates, however, the travel distance through the UZ may be as much as 120 m less due to a higher water table under the glacial transition climate. The SZ is composed of volcanic units and alluvial sediments (Section 6.3.10) with the flow path extending a distance of about 18 km to the accessible environment boundary. The first 12 to 14 km of the SZ flow path is in fractured volcanic rocks, while the remainder of the flow path is through alluvial sediments. The projected direction of the groundwater transport pathway is southeast from the repository site, transitioning to a southerly direction towards the designated accessible environment boundary in the Amargosa Desert.

The role of the Lower Natural Barrier is to prevent or substantially reduce the rate of movement of radionuclides from the repository to the accessible environment. The Lower Natural Barrier performs this role through the intrinsic site characteristics that are directly reflected by such factors as: (1) slow advective water transport, (2) matrix diffusion and sorption of dissolved phase radionuclides, (3) dispersion/dilution of dissolved and colloidal phase radionuclides, (4) reversible filtration of irreversible colloidal phase radionuclides, and (5) radioactive decay and ingrowth. Section 6.3.9.1 discusses the flow and transport processes involved in determining the capability of the UZ, while the relevant SZ processes are discussed in Section 6.3.10.1.

Some of the fundamental and important performance characteristics of the Yucca Mountain Lower Natural Barrier include:

- Low deep percolation rates through the UZ as a result of hydrologic efficiency of the Upper Natural Barrier (i.e., surficial soils with characteristically high evapotranspiration, matrix flow through the PTn, and lateral flow focusing into faults in the northern part of the repository footprint)
- High porosity vitric rock layers (i.e., CHn in southern half of repository) and low permeability zeolitic rock layers (i.e., CHn and CFu), both types of rock layers that have capability to strongly sorb a variety of radionuclides and delay transport
- Long transport path through the SZ volcanic units with capability to delay radionuclides as dissolved radionuclides diffuse into and out of the pores in the rock, increasing the likelihood of sorption onto mineral surfaces, as well as promoting retardation of irreversible colloids (via reversible filtration)
- Lower advective transport through the higher porosity SZ alluvium, with capability to delay radionuclides via sorption of selected radionuclides on mineral particle surfaces and dilution as a result of longitudinal and transverse dispersion.

The performance characteristics of the UZ units below the repository have been analyzed using detailed three-dimensional flow and transport modeling studies. The rates of radionuclide movement through the UZ portion have been studied using a particle tracking technique which is documented in *Particle Transport Model and Abstraction of Transport Processes* (SNL 2008 [DIRS 184748]). That report presents sensitivity analysis for breakthrough times for various radionuclide species. Analogous studies have been performed for the SZ to develop an understanding radionuclide transport to the accessible environment. Those studies are documented in *Saturated Zone Flow and Transport Model Abstraction* (SNL 2008 [DIRS 183750]).

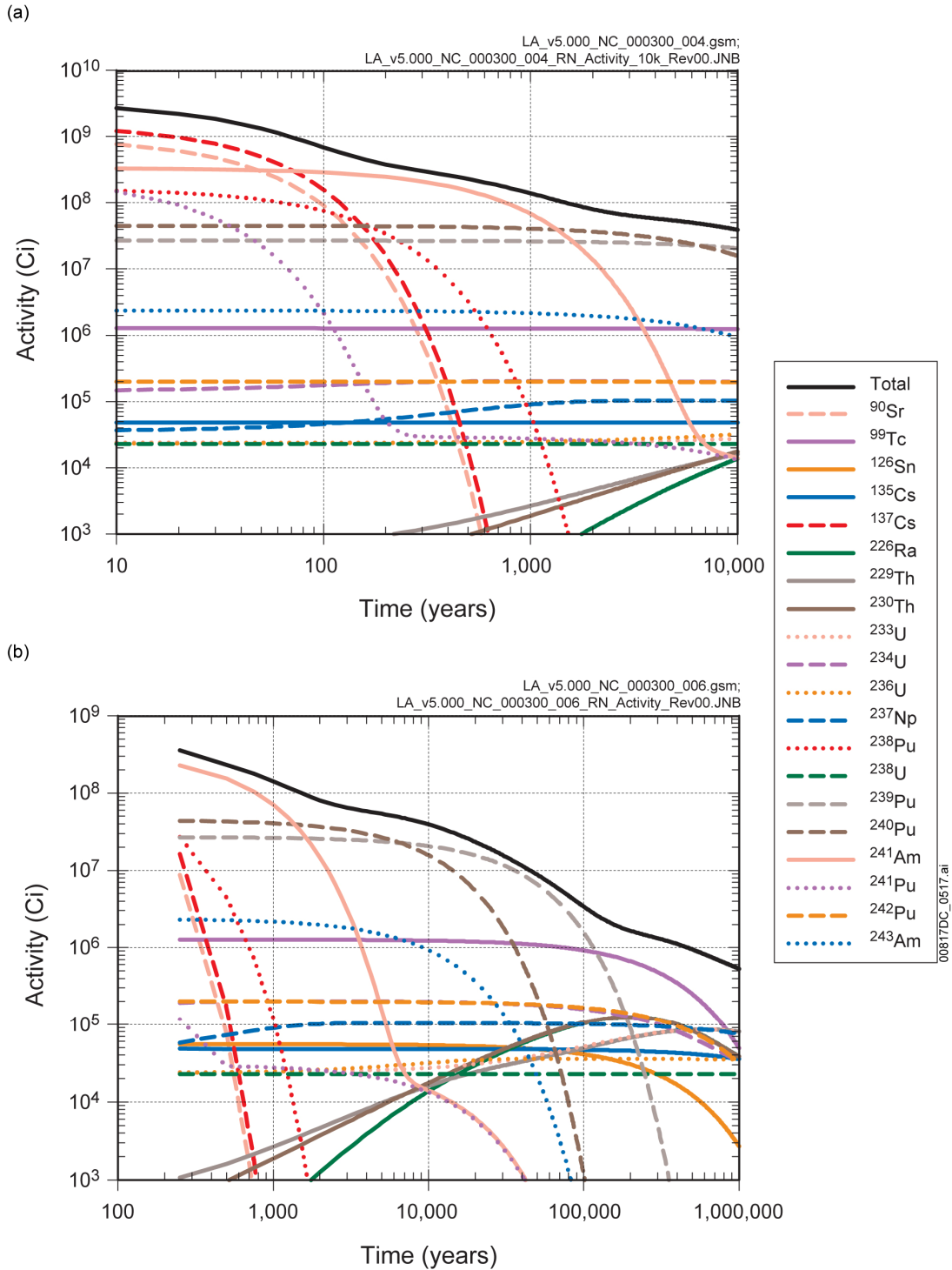
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Table 8.3-1. Decay of Total Curie Inventory as a Function of Time and Dominant Contributors to Total Curie Inventory

Time After Closure (yrs)	Percent of Total Initial Curie Inventory	Major Contributors to Total Inventory at Time after Closure
0	100.00	Cs -137 (46%), Sr-90 (29%), Am-241 (10%)
10	81.2	Cs -137 (45%), Sr-90 (28%), Am241 (12%)
100	20.75	Am-241 (41%), Cs -137 (22%), Sr-90 (13%), Pu-238 (11%)
1,000	4.20	Am-241 (48%), Pu-240 (29%), Pu-239 (19%)
10,000	1.18	Pu-239 (52%), Pu-240 (40%)
100,000	0.10	Pu-239 (46%), Tc-99 (27%)
500,000	0.03	Tc-99 (26%), Th-229 (9%), Th-230 (9%), Ra-226 (9%), U-233 (9%), Np-237 (9%), Pu-242 (8%), U-234 (7%)
1,000,000	0.02	U-233 (15%), Th-229 (15%), Np-237 (14%), Tc-99 (9%), Th-230 (7%), Ra-226 (7%), Cs-135 (7%), U-236 (6%), Pu-242 (6%)

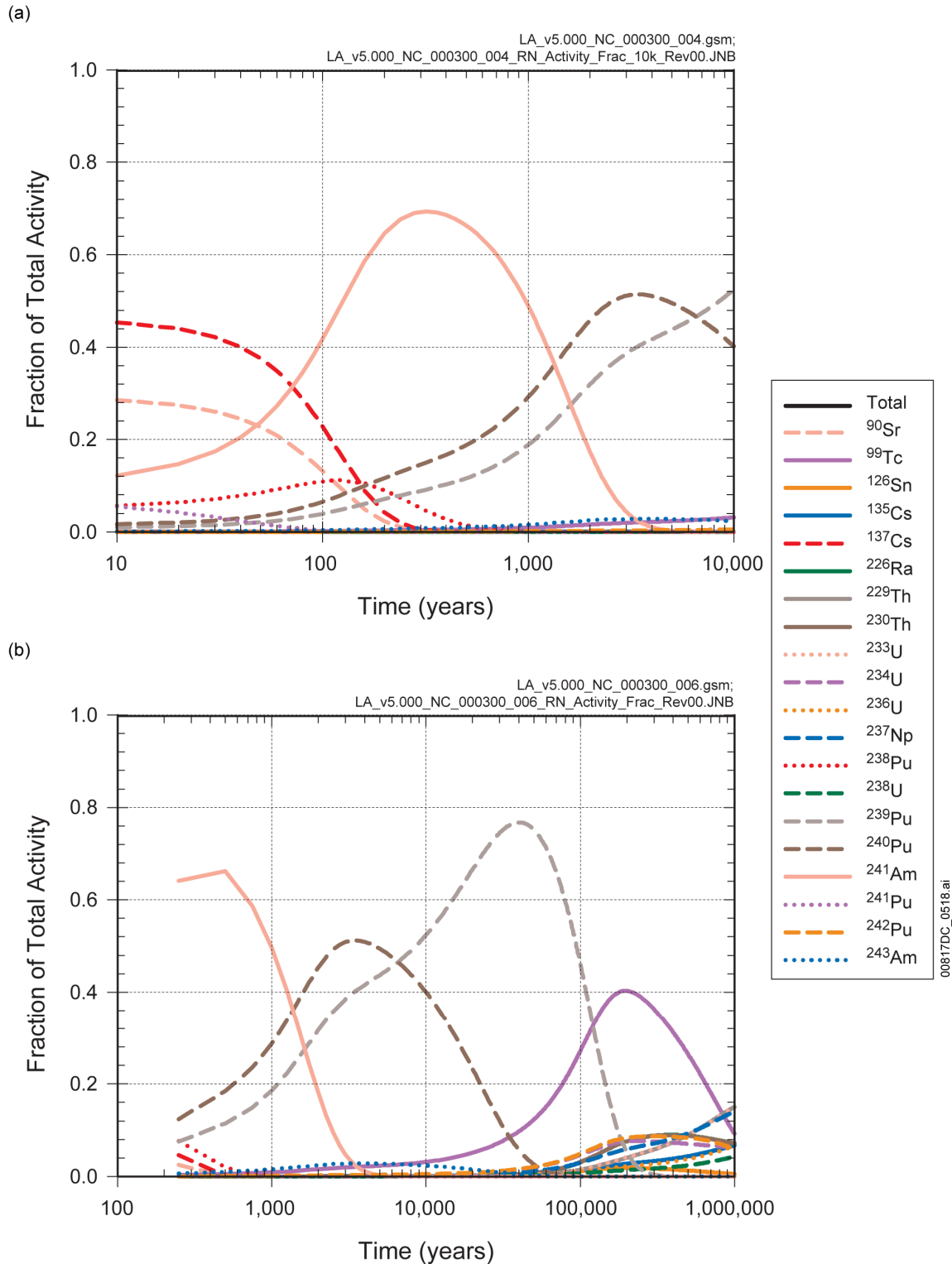
Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

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Source: Output DTN: (a) MO0708TSPAVALI.000 [DIRS 182985]; and (b) MO0709TSPALAMO.000 [DIRS 182981].

Figure 8.3-1. Mean Radionuclide Activities in the Nuclear Waste as a Function of Time for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure



Source: Output DTN: MO0709TSPAREGS.000 [DIRS 182976]

Figure 8.3-2. Mean Radionuclide Contributions to Total Inventory as a Function of Time for (a) 10,000 Years and (b) 1,000,000 Years after Repository Closure

8.4 VALIDITY AND DEFENSIBILITY OF PERFORMANCE DEMONSTRATION

This section outlines the YMP activities that were conducted to ensure that the postclosure performance demonstration for the Yucca Mountain repository would be technically sound and defensible and, therefore, suitable to support the LA. In preparation for the LA, the YMP planned and successfully completed several programmatic activities to ensure that the:

1. TSPA-LA Model and its component models were validated for their intended use, controlled, and documented.
2. Software that implements the TSPA-LA Model was verified and validated through computational testing.
3. Models and input parameters account for the major sources of aleatory and epistemic uncertainties, are supported by appropriate evidence, and are maintained in a controlled database.
4. Performance projections are corroborated, to the extent possible, by other independent means, including auxiliary analyses and comparison with other separate TSPAs.
5. Technical basis for the TSPA-LA methodology was peer reviewed to ensure it is based on well established scientific principles, captures the important phenomena and couplings, and was supported by site characterization data.

The above programmatic activities are presented in Section 7, Volume II, of this report. The description of the TSPA-LA Model is presented in Section 6 of this report; the technical basis for the component models and TSPA input database is documented in several supporting analysis and/or model reports and TSPA Data Input Packages. Moreover, the supporting documentation was prepared in a manner that would ensure the technical basis is auditable and traceable so as to facilitate the licensing review.

Since the issuance of the *Total System Performance Assessment for the Site Recommendation* (CRWMS M&O 2000 [DIRS 143665]), the YMP's TSPA methodology (i.e., models, uncertainty treatment and propagation, code/software, and input parameters) has been significantly improved and refined as a result of:

- Continued scrutiny of the conceptual and mathematical models through internal and external peer reviews
- Further development and testing of TSPA-LA component models and submodels
- Statistical analysis of data and development of improved characterizations of uncertainty.

In addition, plausible and potentially significant model conservatisms have been identified, the rationale for their use explained, and their impact on postclosure performance metrics evaluated.

This section provides a summary of the specific YMP technical and programmatic activities conducted to ensure that the DOE postclosure performance demonstration is suitable to support the LA for the nuclear waste repository at Yucca Mountain.

8.4.1 Validation of TSPA Model and Component Models

At present, the applicable NRC regulation, NRC Proposed Rule 10 CFR Part 63 ([DIRS 178394] and [DIRS 180319]), does not contain requirements for model validation. However, the joint NRC/SKI white paper (Eisenberg et al. 1999 [DIRS 155354], p. 31) notes two essential elements for model validation: (1) procedures for the development of confidence in models, and (2) documentation of results from confidence building activities. These validation elements are incorporated in *Technical Work Plan for: Total System Performance Assessment FY 07-08 Activities* (SNL 2008 [DIRS 184920], Section 2.3.5) and are in accordance with:

- SCI-PRO-002, *Planning for Science Activities*
- SCI-PRO-006, *Models*.

Development and validation of the TSPA-LA Model and its components were planned and controlled in accordance with SCI-PRO-002 and documented in accordance with SCI-PRO-006. It is important to clarify and emphasize that model validation should be viewed as a progressive confidence building process and not an end-state.

The full range of model validation activities conducted for the TSPA-LA is illustrated on Figure 7.1-2. Some of the primary model validation requirements from the two procedures are highlighted here.

During-Development Model Validation—Activities in this first phase of model validation encompassed planning, reviewing, checking, and documenting, which focused on:

- Ensuring the model formulation, assumptions, and simplifications are defensible.
- Ensuring the model theory is consistent with fundamental scientific principles, such as conservation of mass, energy, and momentum.
- Documenting the selection of input parameters and/or input data and explaining how the selection process builds confidence in the model.
- Documenting plans for model calibration activities, and/or initial boundary condition runs, and/or run convergences, and a discussion of how the activity or activities build confidence in the model. Activities also included a discussion of impacts of any run non-convergences.
- Identifying and documenting the potential impacts of uncertainties on model results.

A detailed description of the during-development model validation activities is presented in Section 7.1.2 (Volume II).

Post-Development Model Validation—In this stage of model validation, technical activities involve conducting various computational analyses and evaluations to ensure that the processes modeled are understood and appropriately described. Moreover, the analyses and evaluations demonstrate that the required level of model validation has been fully achieved. The major activities performed for the post-development model validation include:

- Corroborating the abstraction model results to the results of the validated mathematical model or process model from which the abstraction model was derived (Section 7.6)
- Performing auxiliary analyses to corroborate the results with the TSPA-LA Model and/or submodels (Section 7.7)
- Comparing the relevant TSPA-LA component models or submodels with available analogue information (Section 7.8)
- Utilizing internal and external peer reviews to evaluate defensibility of models and identifying potential areas of component model or submodel development improvement (Section 7.9).

More information on the post-development model validation activities is presented in Section 7.1.3.

The TSPA-LA Model and the 11 component models (Section 6) have been validated in accordance with the appropriate procedures. Documentation of the model validation is presented in each of the supporting analysis and/or model reports and includes a description of the validation procedure, validation criteria, and validation results.

8.4.2 Verification and Validation of TSPA Software and Input Data

The TSPA-LA Model and its component models are integrated into the GoldSim software (GoldSim 2007 [DIRS 181903]). The version of the GoldSim software used for the TSPA-LA was verified by the vendor, GoldSim Technology Group, by conducting a broad spectrum of computational tests. The software verification consisted of over 250 tests that covered the program's capabilities, including the user interface, user-defined expressions, internal functions, and distributed processing capabilities. These tests included:

- 93 basic functional tests
- 23 time and Monte Carlo tests
- 130 contaminant transport tests
- 17 reliability module tests.

As part of the verification testing, the software was run through tests that exercise the graphical user interface, internal functions, stochastic processes, contaminant transport code, and result displays for the purpose of demonstrating that the software performed its numerical, logical, and input/output operation correctly. All verification tests were performed in accordance with a verification plan (DOE 2007 [DIRS 181107]).

The YMP staff reran all the verification test cases to provide an independent check; the software testing was performed in accordance with IM-PRO-004, *Qualification of Software*. For added confidence, key functions and capabilities of the simulation software were combined into four TSPA-developed tests, and the results were compared to independently generated results. The independent tests are described in the *Design Document for: GoldSim v9.60* (DOE 2007 [DIRS 181107], Section 7.2.1), and the results are shown in *Software Validation Report for: GoldSim v9.60 on Windows 2000* (DOE 2007 [DIRS 181109], Sections 4.2.34 through 4.2.37). The integrated system software (GoldSim) was qualified in accordance with IM-PRO-004. In addition to verification of the software, testing was performed to verify the computational stability and convergence. This included testing the:

1. Stability of the statistical sampling used in the Monte Carlo simulations (Section 7.3.1)
2. Numerical accuracy of the expected dose calculation (Section 7.3.2)
3. Convergence of the time stepping used in the numerical integration of the model equations (Section 7.3.3)
4. Spatial discretization of the repository system (Section 7.3.4).
5. Stability of transport models using the FEHM particle-tracking methodology (Section 7.3.5).

All of this computational testing was performed for the various modeling cases to ensure that the GoldSim model was ready for use in the performance demonstration for the LA.

The input database for the GoldSim software stores all the TSPA-LA input parameter values and distributions. This input database is used to categorize, store, and retrieve fixed and distributed values of the TSPA-LA Model parameters. The database is programmed with user controls featuring read and write access and audit trails. These controls were designed to ensure the security, integrity, and traceability of the information used in the TSPA-LA Model analyses. A controlled, standalone TSPA Input Database was developed using a commercially available desktop database manager, Microsoft Access 2000. It is a Microsoft Windows based, multi-user relational database solution that allows data entry, viewing, and querying, as well as report preparation. The TSPA Input Database was qualified in accordance with IM-PRO-004.

The parameter values are obtained from controlled sources maintained by the project data and information systems, such as project documents, the Technical Data Management System (TDMS), and the Technical Information Center. The TDMS database is a project-wide database, whereas the TSPA Input Database is used only for the TSPA-LA Model. Each of the parameter sets used in the TSPA Input Database has a data tracking number (DTN) to provide the link to the TDMS database or a reference to the controlled source of the information, such as an analysis and/or model report. The TDMS maintains the qualification status all of its contents. The TSPA Model inputs are controlled and well documented by using the TSPA Input Database. An important configuration management feature of the TSPA Input Database is that it supports the independent verification of every parameter value used in the TSPA-LA Model.

GoldSim software and the TSPA input database have been verified. Additional information on the validation and verification is presented in Section 7.

8.4.3 Uncertainty Characterization Reviews

In preparation for the LA, the Lead Laboratory undertook a substantive effort to systematically review the TSPA uncertainty and variability characterizations for consistency, defensibility, and traceability. A special review team was formed to carefully examine parameter uncertainty and variability representations, correct or modify them as necessary, and ensure that the supporting documentation would facilitate the regulatory review.

The central objective of the uncertainty characterization reviews was to ensure that the treatment of parameter uncertainty and variability was of suitable quality for the LA. To accomplish this objective, the review team focused on:

- Confirming that the stochastic parameter representations appropriately reflect the major sources of uncertainty and/or variability
- Verifying that the probability distributions were supported by appropriate evidence and derived using sound statistical methods and interpretations
- Ensuring that model parameter representations (i.e., probability distributions) were reasonable and defensible, as opposed to depicting extreme variations that could potentially introduce risk dilution (i.e., wider distribution and lower peak mean annual dose).

Uncertainty characterizations were also reviewed with respect to appropriate up-scaling, data modeling, and use of professional judgment in assigning subjective distributions.

A core team of five senior staff members was formed with special expertise in probability and statistics, uncertainty analysis, TSPA modeling, and knowledge of the regulatory guidance regarding consistent treatment of uncertainty and variability. In addition, a small group of SMEs were used to support the core team and to facilitate the reviews of data, parameters, and model abstractions. The review team conducted fifteen formal reviews that were performed to scrutinize the uncertainty characterizations of some forty TSPA input parameters and their associated abstractions.

Because of the relatively large number (~400) of probabilistic parameters used in the TSPA-LA Model, it was necessary to prioritize the parameters for review. The prioritization was developed based on the importance ranking of the scenario modeling cases and parameters within those modeling cases. The ranking of the scenario modeling cases was as follows (from highest to lowest):

1. Seismic Scenario Class, Seismic GM Modeling Case
2. Igneous Scenario Class, Igneous Intrusion Modeling Case
3. Igneous Scenario Class, Volcanic Eruption Modeling Case
4. Early Failure Scenario Class, Waste Package EF Modeling Case

5. Early Failure Scenario Class, Drip Shield EF Modeling Case
6. Seismic Scenario Class, Seismic FD Modeling Case
7. Nominal Scenario Class, Nominal Modeling Case.

A list of 40 plus TSPA-LA parameters (Section 7.4, Tables 7.4-1 through 7.4-3) was compiled. The review focused on parameters whose uncertainty or variability (when propagated through the TSPA-LA Model) would have the greatest influence on the magnitude of the dose results, and the uncertainty in the dose results.

The 15 reviews produced several findings, recommendations, and corrective actions that resulted in improvement in the technical basis and documentation of the parameter uncertainty characterizations. A few of the key improvements to parameter uncertainty characterizations included:

- Unsaturated zone (UZ)—effective soil depth for surface infiltration and weighting factors for UZ flow fields
- Engineered Barrier System (EBS)—general corrosion of the drip shield and waste package internals
- Saturated zone (SZ)—groundwater specific discharge multiplier and flowing interval thickness.

A more detailed discussion of the uncertainty characterization reviews conducted for the TSPA-LA is presented in Section 7.4.

8.4.4 Corroboration of TSPA-LA Model Results

A number of methods (Section 7.0) have been utilized to build confidence in the TSPA-LA Model. One of those approaches is based on corroborating the TSPA-LA performance projections using a simplified and independent computer model (Section 7.7). Performance projections presented in Sections 8.1 and 8.2 were corroborated through: (1) use of a simplified TSPA-LA analysis, and (2) comparison with the Electric Power Research Institute (EPRI) independent assessment of postclosure performance for the proposed repository. The simplified TSPA-LA analysis and the EPRI results have been used to check for consistency of results for the important modeling cases.

The confidence building achieved through these corroborations of the TSPA-LA Model results is emphasized in the following paragraphs.

8.4.4.1 Comparison with Simplified TSPA Analysis Results

A separate and simpler implementation of the performance assessment of postclosure performance (Appendix L) was developed for the specific purpose of corroborating the performance projection results produced with the TSPA-LA Model. That implementation is in a standalone and separate computer code designated as the Simplified TSPA. The Simplified TSPA code uses a Monte Carlo approach to incorporate epistemic and aleatory uncertainties to the probabilistic calculation of mean annual doses. While the formulation of its component

models generally parallels those of the TSPA-LA Model, the Simplified TSPA code employs several simplifications. A detailed enumeration of the differences between the TSPA-LA Model and Simplified TSPA is presented in Section 7.7.2, Table 7.7.2-1. Three important simplifications are: (1) the repository is represented as a single homogeneous region, (2) flow and transport through UZ and SZ are one-dimensional, and (3) approximation for decay chain transport is more conservative.

Thus far, the Simplified TSPA has only been applied to four of the seven TSPA-LA modeling cases (Section 6.1.2) to calculate the mean annual doses to the RMEI. More specifically, the Simplified TSPA has been applied to: Nominal, Waste Package EF, Seismic GM, and Igneous Intrusion. These specific modeling cases were selected because previous TSPAs indicated they would have the greatest influence on the total mean annual doses to the RMEI. This observation is also confirmed in the TSPA-LA results presented in Section 8.1.

Nominal Modeling Case Results—The TSPA-LA Model projections for this case are shown on Figure 8.2-1 and for Simplified TSPA on Figure 7.7.2-5. The TSPA-LA Model projects a dose history where the exposure begins at about 130,000 years and increases, reaching a peak at 1,000,000 years. The Simplified TSPA results show a similar trend, but the WP failures occur later, so exposures are delayed until roughly 300,000 years. TSPA-LA Model results show that a peak mean annual dose for this modeling case is about 0.4 mrem, which compares reasonably well with the Simplified TSPA projection of about 0.2 mrem. A time-slice comparison between the two mean annual dose projections is discussed Section 7.7.2.2 and shown on Figure 7.7.2-6. With regard to radionuclides important to dose, the TSPA-LA Model results show ^{129}I as the main contributor, with small contributions from ^{99}Tc , ^{135}Cs , and ^{242}Pu . The Simplified TSPA results also show the mean annual dose to be dominated by ^{129}I , with lesser contributions from ^{242}Pu . The main difference between the two model results for the Nominal Modeling Case is the timing of WP failures and radionuclide releases. This is explained by the fact that the Simplified TSPA represents the repository as a single block with average properties, and as such does not account for spatial variability.

Waste Package Early Failure Modeling Case Results—For this modeling case, the TSPA-LA Model results are shown on Figure 8.2-5 and the Simplified TSPA results on Figure 7.7.2-2. The TSPA-LA Model results show a relatively flat mean annual dose history over the 1,000,000-year time period, with doses varying from about 10^{-3} to 10^{-1} mrem. The Simplified TSPA shows a similar trend, but the doses vary from about 10^{-2} to 10^{-1} mrem. A time-slice comparison between the two mean annual dose projections is discussed Section 7.7.2.1 and shown on Figure 7.7.2-3. The TSPA-LA Model results indicate that ^{242}Pu , ^{226}Ra (daughter product of ^{230}Th), and ^{237}Np dominate the mean annual dose. In the Simplified TSPA projections, the mean annual dose at 1,000,000 years is dominated by ^{229}Th (daughter product of ^{233}U) and ^{242}Pu and with lesser contributions from ^{237}Np . The differences in radionuclides contributing to the dose are attributed to how the decay chain transport is approximated in the Simplified TSPA code.

Seismic Ground Motion Modeling Case Results—Graphical results for this modeling case are shown on Figure 8.2-11 for the TSPA-LA Model and on Figure 7.7.2-8 for the Simplified TSPA. Both the TSPA-LA Model and Simplified TSPA results show the mean annual doses histories to rise to a peak mean value and then remain relatively flat. However, the TSPA-LA Model shows a peak mean annual dose of about 0.1 mrem, whereas the Simplified TSPA shows a higher value

of about 1 mrem. A time-slice comparison between the two mean annual dose projections is discussed Section 7.7.2.3 and shown on Figure 7.7.2-9. The TSPA-LA Model results show ^{99}Tc and ^{129}I as the main contributors to dose for the majority of the 1,000,000-year time period, whereas the Simplified TSPA dose history is largely dominated by ^{242}Pu , ^{237}Np , and ^{229}Th .

Igneous Intrusion Modeling Case Results—For this modeling case, the TSPA-LA Model results are shown on Figure 8.2-7 and the Simplified TSPA results on Figure 7.7.2-11. The TSPA-LA Model results show a mean annual dose history that rises until reaching a peak of 1 mrem at about 80,000 years and then remaining relatively flat thereafter. The Simplified TSPA results indicate a similar trend but reach a peak mean annual dose of 10 mrem. A time-slice comparison between the two mean annual dose projections is discussed Section 7.7.2.4 and shown on Figure 7.7.2-12. In both the TSPA-LA Model and Simplified TSPA projections, the initial peak dose is attributed to ^{239}Pu . From about 200,000 years to 1,000,000 years, the TSPA-LA Model results show that dose is largely attributed to ^{226}Ra , ^{242}Pu , and ^{237}Np . For this same time interval, the Simplified TSPA results show ^{242}Pu , ^{237}Np , and ^{229}Th as dominating the dose. As noted previously, the daughter products, such as ^{229}Th and ^{226}Ra , are computed differently in the two transport codes, which explain the differences.

Summary of Corroboration—For most of the modeling cases, the performance projections for two separate models compare reasonably well both in terms of the mean annual dose levels and the radionuclides important to the dose. The few differences noted are explainable and attributed to use of different approaches for calculating decay chain transport (i.e., differential transport of precursor and daughter product).

A more detailed discussion of the corroboration of the TSPA-LA Model results using the Simplified TSPA Analysis is presented Section 7.7.2.

8.4.4.2 Comparison with Electric Power Research Institute TSPA Analysis

The EPRI developed its own TSPA Analysis, which is implemented in the Integrated Multiple Assumptions and Release Code (IMARC) (EPRI 2005 [DIRS 178580]). EPRI developed IMARC to provide an independent assessment of key technical and scientific issues associated with the proposed repository at Yucca Mountain. Like the TSPA-LA Model, the EPRI TSPA Analysis uses coupled component models to simulate the response of the repository to changing conditions and disruptive events. With regard to treatment of uncertainty, the EPRI TSPA Analysis accounts for epistemic uncertainties in the model parameters but does not consider aleatory uncertainties associated with occurrence of disruptive events. At present, the EPRI TSPA Analysis only considers four scenarios consisting of nominal, igneous, seismic, and human intrusion. In contrast to the TSPA-LA Model, the EPRI TSPA Analysis for the nominal scenario includes early failure of one WP and one DS. The available EPRI documentation presents a simulation of postclosure performance for the combined nominal and early failure modeling case for a 1,000,000-year time period.

Combined Nominal and Early Failure Modeling Case Results—The TSPA-LA Model projections for this case are shown on Figure 7.7.3-3 and for IMARC on Figure 5-10 of *Program on Technology Innovation: Evaluation of a Spent Fuel Repository at Yucca Mountain, Nevada, 2005 Progress Report* (EPRI 2005 [DIRS 182229]). The mean annual dose history curves for

TSPA-LA Model and IMARC show a similar trend in the dose histories, with a significant increase in dose to the RMEI after about 100,000 years. The TSPA-LA Model projections indicate a peak mean annual dose of 0.4 mrem at 1,000,000 years, whereas the IMARC result shows a peak of about 0.02 mrem. In both the TSPA-LA Model and IMARC model results, the peak mean annual dose is dominated by a single radionuclide, ^{129}I . In addition to peak dose, the other notable difference between the two performance results is the initial breakthrough; the TSPA-LA Model results indicate dose exposure starting at about 400 to 500 years, whereas the IMARC results occur at about 7,000 years.

The main differences in the two independent projections are largely attributed to a distinct technical basis for the EBS and release rate calculation. For example, the EPRI TSPA Analysis only accounts for CSNF waste and considers failure of DS, WP, and cladding, whereas the TSPA-LA Model accounts for CSNF, DSNF, and HLW, but does not take credit for cladding in CSNF WPs. In addition, seepage rates used in the EPRI TSPA Analysis are significantly lower than the corresponding rates used in the TSPA-LA Model, which appear to cause a delay in radionuclide release from the EBS. There are other differences related to the abstraction of features and processes. It is important to keep in mind that the performance projections published by EPRI were intended to status the development of their independent capability and not to present a performance demonstration in accordance with NRC regulation.

Summary of Corroboration—The comparisons of performance projections generally corroborate the TSPA-LA results and therefore provide additional confidence in the validity and defensibility of the performance demonstration. A more detailed discussion of the comparison of the EPRI TSPA Analysis and TSPA-LA Model results is presented in Section 7.7.3

Performance Margin Analysis

A Performance Margin Analysis (PMA) was conducted to quantitatively evaluate the differences in repository performance due to significant explicit and implicit conservatisms embedded in the TSPA-LA Model subcomponents. The conservatisms were evaluated to (1) confirm that they are conservative with respect to the mean annual dose of the TSPA-LA Model; (2) quantify the extent to which they, individually and collectively, overestimate the projected annual dose; and (3) assess that the evaluated conservatisms did not introduce any inappropriate risk dilution in the TSPA-LA results presented in support of the LA. The PMA was conducted by first modifying selected submodels and parameters of the TSPA-LA Model, including the additional submodels and parameters for the PMA and then repeating the sequence of calculation for a select set of modeling cases that were run for the TSPA-LA. PMA was conducted for both 10,000-year and 1,000,000-year time periods and on the same set of modeling cases as the TSPA-LA Model. The details of approach and results of the PMA are presented in Section 7.7.4 with additional supporting material in Appendix C. Summarizing here, the results show that the conservatism evaluated in the PMA are indeed conservative with respect to the total system performance measures (e.g., peak mean annual dose), as the largest doses calculated in the PMA for the 10,000 year and 1,000,000 year are significantly lower than the doses used in compliance demonstration. The largest calculated PMA mean annual doses are lower by over an order of magnitude and a factor of two over the largest mean annual dose relative to the TSPA-LA Model (Section 8.1) for the time periods of 10,000 years and 1,000,000 years, respectively. Further, the PMA analysis demonstrated that the significant conservatisms did not introduce risk dilution in

the TSPA-LA results presented (Section 8.1.1 in Volume III) in support of compliance with the regulatory dose requirement, as demonstrated by the absence of higher peak doses in the PMA results for both the probabilistic projections of the expected annual dose and the comparison of the projected total mean annual dose for the PMA relative to the TSPA-LA. The PMA results also have different significant modeling cases than the TSPA-LA Model due primarily to the items selected for modification in PMA.

8.4.5 Reviews of YMP TSPA Methodology

Independent reviews have played an important role in the development and advancement of the TSPA-LA methodology, as well as in improving its overall conceptual framework. Peer reviews conducted by teams of external experts were commissioned by the YMP in order to ensure that the TSPA methodology would be based on well-established principles, evidence supported, and be technically defensible for the licensing review. Various national and international expert groups have participated in formal reviews of the TSPA methodology and its supporting technical basis. Moreover, various federal oversight and regulatory agencies have regularly reviewed the TSPA methodology. These agencies have included the Nuclear Waste Technical Review Board (NWTRB), NRC's Advisory Committee on Nuclear Waste, and NRC's Division of Nuclear Materials, Safety, and Safeguards staff.

In the subsequent text, three of the comprehensive reviews that were conducted by groups independent of the TSPA-LA Model development are summarized.

8.4.5.1 TSPA-Viability Assessment Peer Review

The first independent and formal evaluation of the TSPA methodology was conducted in support of the DOE Viability Assessment of the Yucca Mountain site (DOE 1998 [DIRS 101779]). The six member peer review group, designated as the TSPA-Viability Assessment (VA) Review Panel, conducted a phased review of the YMP's TSPA methodology over a two-year period. The TSPA-VA peer review was conducted in accordance with the Management and Operating Contractor's QA procedure, QAP-3-3, *Peer Review*, which was in effect at that time of the review.

With regard to the TSPA-VA methodology, the review panel concluded that the "overall performance assessment framework and the approach used to developing the TSPA-VA were sound and followed accepted methods" (Budnitz et al. 1999 [DIRS 102726], p. 2). However, the panel identified deficiencies in some of the component models; these aspects were judged deficient because they, in the view of the panel, lacked an adequate theoretical basis. Some of the review panel's major comments included: (1) TSPA component models and their adequacy to capture relevant phenomena, (2) proper coupling between TSPA component models and submodels, (3) testing and evaluation of the modeled behavior, (4) adequacy and completeness of the supporting database, (5) treatment of model and parameter uncertainties, and (6) potential non-conservative approaches.

Some of the panel's specific recommendations for improving the YMP's TSPA methodology included:

- Make the TSPA submodels more realistic, where supported by data
- Reduce uncertainty (i.e., epistemic) through additional data collection
- Perform additional auxiliary analyses and sensitivity analyses to better understand complex system behavior
- Perform qualitative and quantitative comparisons of the TSPA model components with several man-made and natural analogues
- Conduct follow-on peer reviews of the TSPA methodology.

The YMP implemented the review panel's recommendations for improving the TSPA methodology, as well as those for additional field and laboratory data collection. The review panel's final report is entitled *Peer Review of the Total System Performance Assessment-Viability Assessment Final Report* (Budnitz et al. 1999 [DIRS 102726]). The YMP staff responses to the review panel's comments and recommendations are documented in *Comment Response on the Final Report: Peer Review of the Total System Performance Assessment-Viability Assessment (TSPA-VA)* (CRWMS M&O 1999 [DIRS 153111]).

Additional information on the TSPA-VA review panel comments is presented in Section 7.9.1 of Volume II.

8.4.5.2 Joint Nuclear Energy/International Atomic Energy Agency Peer Review

At the request of DOE, the Organization for Economic Cooperation and Development's (OECD's) Nuclear Energy Agency (NEA) and the International Atomic Energy Agency (IAEA) assembled a team of experts to peer review the TSPA Model for Yucca Mountain. That version of the TSPA Model was to be used for the assessment conducted to support the DOE site recommendation (SR) process. The 10-member international review group, designated as the International Review Team (IRT), reviewed the following aspects of the TSPA-SR model:

- Technical basis for the performance assessment, including identification and justification of the conditions and characteristics modeled at the system level
- Development of the key conceptual models, including the assumptions made with respect to the representations of relevant FEPs
- Adequacy of the treatment of the undisturbed and disturbed system performance
- Adequacy of the methods used, and the cases considered, in sensitivity and uncertainty evaluations
- Overall clarity and completeness of the technical report describing this system-level performance evaluation.

The IRT review was conducted during June 2001 through December 2001. As indicated in their final report, the IRT concluded that “the TSPA-SR methodology is soundly based and has been implemented in a competent manner” (OECD and IAEA 2002 [DIRS 158098], p. 10). In their final report, the IRT made a total of 27 major recommendations for future improvements of the TSPA methodology. Some of the general recommendations for improvements included:

1. A more systematic treatment of uncertainties
2. An investigation of potential risk dilution
3. Improved characterization of the waste forms
4. Use of natural analogues.

The YMP addressed all of the IRT recommendations in developing the TSPA-LA model and in preparing the postclosure demonstration for the LA. The improved systematic treatment of aleatory and epistemic uncertainties is reflected in Section 6.1.2 (Volume I) and Appendix J (Volume III). The investigation of potential risk dilution was largely accomplished through a critical review of the TSPA parameter uncertainty characterizations, which is discussed in Section 8.4.3. Improved characterization of the waste forms is discussed in Section 6.3.7. The use of natural analogs is described in Section 7.8.

A more detailed summary of the IRT review comments and recommendations is presented in Section 7.9.2.

8.4.5.3 Independent Validation Review Team Technical Review

In early preparations for the LA, the YMP commissioned an independent and comprehensive review of the TSPA methodology with the objectives of: (1) examining the defensibility of the TSPA methodology, (2) implementing a model validation strategy, and (3) making a determination of model validity based on a technical review. The critical review of the technical basis was to be based on reviews of the supporting analysis and/or model reports. Implementation of the model validation strategy would follow the applicable process steps outlined in the NRC/SKI white paper on model validation (Eisenberg et al. 1999 [DIRS 155354], pp. 21 to 26). A determination of TSPA Model validity was to be based on conformance with detailed technical criteria. An independent group, designated as the Independent Validation Review Team (IVRT), was formed and chartered to accomplish the review objectives. This team was independent of the TSPA-LA Model development.

The in-depth critical reviews of the TSPA methodology were largely conducted in 2004, while the implementation of a model validation strategy and determination of model validity was performed in 2005. The technical review was conducted in accordance with the then effective QA procedure, LP-SIII.10Q-BSC, *Models*. The IVRT issued their final report in February 2006 (Booth 2006 [DIRS 176638]).

The following is a summary of the IVRT’s major recommendations and findings that lead to further improvements of the TSPA-LA methodology.

Review TSPA Methodology and Technical Basis—The IVRT critical reviews of the TSPA technical basis were largely conducted in parallel with the preparation of the supporting analysis and/or model reports and the TSPA component models. The concurrent review and development

efforts had the advantage that the IVRT comments and recommendations could be addressed and incorporated in the final analysis and/or model reports and in the updated version of the TSPA Model. With regard to the TSPA methodology, the IVRT critique and recommendations lead to significant improvements (Booth 2006 [DIRS 176638]) in the defensibility of:

- Conceptual framework and mathematical formulation of the seismic ground motion and igneous intrusion component models
- Formulation and coupling of in-package chemistry and the radionuclide transport abstraction
- Biosphere dose conversion factors and number of key radionuclides considered in the calculation of organ dose
- Uncertainty characterizations for the igneous and seismic events.

The IVRT thoroughly reviewed all the draft analysis and/or model reports and provided nearly 400 technical comments; addressing those comments improved the defensibility of the technical basis as well as the transparency of analysis and/or model reports. One of the more important benefits of this part of the IVRT review was the careful scrutiny of the plausible conservatisms, identification of potential optimisms, and inconsistencies in draft component models. As a result of these comments, the TSPA staff corrected the inconsistencies and conducted impact analyses to evaluate the significance of the conservatisms and potential optimisms in terms of the relative movement of the mean dose curve.

Model Validation Process and Determination of Model Validity—As directed by the YMP, the IVRT critically reviewed the TSPA methodology against 20 technical criteria grouped into three categories (BSC 2005 [DIRS 173309], pp. 18 to 20), namely: (1) general criteria for the TSPA Model framework, (2) specific criteria for the component models and submodels, and (3) specific criteria for the total system model. A total of five general criteria were specified for the model framework that focused on examining the appropriateness and reasonableness of the overall conceptual model, identification of FEPs and inclusion in the models, consistency of the process models and derived abstractions, and confidence-building analyses. The specific criteria for submodels, which consisted of five criteria, focused on the mathematical basis of the models (e.g., assumptions and technical bases, time step and spatial discretization, and consistency of simulation outputs and model abstraction). The 10 specific criteria for the total system model covered such aspects as statistical sampling of uncertain parameters, initial and boundary conditions, parameter distributions, number of realizations, and verification of the TSPA-LA Model software.

The IVRT findings with respect to the general and specific review criteria are presented in the attachment to the memo containing reference to the TSPA-LA (Booth 2006 [DIRS 176638], Attachment draft Rev 01E, Sections C.5.1 through C.5.3, pp. C-39 through C-116). The IVRT determinations regarding the 20 criteria being met, or not met, were used to make a peer group judgment with respect to conformance with two key model validation goals (BSC 2005 [DIRS 173309], Section 2.10.1), namely:

1. To describe the postclosure performance of the repository system for the nominal, igneous, and seismic scenario classes
2. To produce an estimate of mean dose (and other performance measures, as appropriate) that is consistent with the degree of conservatism representative of the component abstraction models and parameters (and their uncertainty) that are input to the TSPA-LA Model.

The IVRT interpreted the first goal as to “require an approach to TSPA Model development that produced estimates of the performance measures of interest with neither pessimistic nor optimistic bias” (Booth 2006 [DIRS 176638], Attachment draft REV 01E, Section C.5.0.2, p. C-37). The second goal was interpreted by the IVRT to “require an approach to model development that did not underestimate the performance measures because the component [model] abstraction and parameter probability density functions were developed with a conservative bias” (Booth 2006 [DIRS 176638], Attachment draft REV 01E, Section C.5.0.2, p. C-37).

The IVRT finding for these two model validation goals was that they had not been met by the 2005 version of the TSPA Model. The rationale for their finding centered on their views that:

- Uncertainty was not characterized as realistically as available information and data allowed
- Inconsistent treatment of uncertainty among the barrier components
- Model appeared to contain potentially significant optimisms and conservatisms.

For the most part, the IVRT views were attributed to deficiencies in the component models for the DS and WP, which were not sufficiently mature at the time of the review. The optimisms referred to by the IVRT were attributed to the: (1) DS component model not accounting for degradation and failure mechanisms, and (2) DS and WP component models not accounting for early failures. With regard to conservatisms, the IVRT noted that the representation of the natural barriers was underestimated to the extent of being extremely conservative (Booth 2006 [DIRS 176638], Attachment draft REV 01E, Section C.5.1.1.2.1, p. C-42).

The IVRT’s findings on the two model validation goals prompted additional work to improve the uncertainty treatment, remove the excessive model conservatisms and optimisms. Moreover, the IVRT’s finding altered the TSPA compliance strategy from one that relied heavily on conservatism to one that used more realistic models and more rigorous representation of uncertainties.

Additional information on the IVRT review comments and recommendations are presented in Section 7.9.3.

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9.2 CODES, STANDARDS, REGULATIONS, AND PROCEDURES

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